Disparate effects of global-change drivers on mountain conifer forests:

warming-induced growth enhancement in young trees vs. CO2 fertilization

in old trees from wet sites

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Running head: CO<sub>2</sub> and tree growth in subalpine forests

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## Abstract

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2 Theory predicts that the post-industrial rise in the concentration of  $CO_2$  in the atmosphere  $(c_a)$ should enhance tree growth either through a direct fertilization effect or indirectly by 3 4 improving water use efficiency in dry areas. However, this hypothesis has received little support in cold-limited and subalpine forests where positive growth responses to either rising 5 6  $c_{\rm a}$  or warmer temperatures are still under debate. In this study we address this issue by 7 analyzing an extensive dendrochronological network of high-elevation *Pinus uncinata* forests in Spain (28 sites, 544 trees) encompassing the whole biogeographical extent of the species. 8 9 We determine if the basal-area increment (BAI) trends are linked to climate warming and 10 increased  $c_a$  by focusing on region- and age-dependent responses. The largest improvement in BAI over the past six centuries occurred during the last 150 years affecting young trees and 11 being driven by recent warming. Indeed, most studied regions and age classes presented BAI 12 13 patterns mainly controlled by temperature trends, while growing-season precipitation was only relevant in the driest sites. Growth enhancement was linked to rising  $c_a$  in mature (151-14 15 300 years old trees) and old mature trees (301–450 years old trees) from the wettest sites only. This finding implies that any potential fertilization effect of elevated  $c_a$  on forest growth is 16 contingent on tree features that vary with ontogeny and it depends on site conditions (for 17 18 instance water availability). Furthermore, we found widespread growth decline in droughtprone sites probably indicating that the rise in  $c_a$  did not compensate for the reduction in water 19 availability. Thus, warming-triggered drought stress may become a more important direct 20 21 driver of growth than rising  $c_a$  in similar subalpine forests. We argue that broad approaches in 22 biogeographical and temporal terms are required to adequately evaluate any effect of rising  $c_a$ on forest growth. 23 Keywords: basal area increment, climate warming, dendroecology, Generalized Additive 24 Mixed Models, subalpine forests, Pinus uncinata. 25

## Introduction

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27 The post-industrial planet has experience a steep increase in  $c_a$  which is linked to warmer air temperatures (IPCC, 2013). However, current  $c_a$  does not suffice to saturate photosynthesis of 28 C<sub>3</sub> tree species (Norby & Zak, 2011). Therefore, future growth trends of those tree species 29 will depend on how forests respond to a rapidly CO<sub>2</sub>-enriched and warmer world (Bonan, 30 2008). Short-term experiments based on young trees growing under high  $c_a$  levels do not 31 allow a full understanding how mature trees will respond to elevated  $c_a$  in the field where 32 growth is mediated by ontogeny and limited by temperature oscillations and water availability 33 (Körner, 2013). For example, rising  $c_a$  stimulates photosynthesis and decreases stomatal 34 35 conductance through an improved intrinsic water-use efficiency (iWUE), i.e. the carbon gain per unit of water lost (Norby et al., 1999). However, such improved iWUE does not 36 necessarily translate into enhanced growth in the field (Peñuelas et al. 2008, 2011; Linares & 37 38 Camarero, 2012). The questions of whether, where, and when tree growth is influenced by post-39 industrial  $c_a$  still remain open and controversial. Numerous field studies performed in 40 different biome and forest types worldwide have found either positive (e.g., LaMarche et al., 41 1984; Hari & Arovaara, 1988; Graybill & Idso, 1993; Idso & Kimball, 1993; Nicolussi et al., 42 1995; Soulé & Knapp, 2006; Voelker et al., 2006, Wang et al., 2006; Martinez-Vilalta et al., 43 2008; Cole et al., 2010, Koutavas, 2013), or neutral to negative (Kienast & Luxmoore, 1988; 44 Graumlich, 1991; Tognetti et al., 2000, Gedalof & Berg, 2010; Andreu-Hayles et al., 2011; 45 Girardin et al., 2011; Nock et al., 2011; Lévesque et al., 2014) growth responses to higher c<sub>a</sub>. 46 A possible explanation for these contrasting findings can be that the influence of  $c_a$  on tree 47 growth depends either on local site conditions or on tree ontogeny. 48 Local conditions modulate the effect of recent rising temperature on growth since 49 post-industrial temperature trends vary in space (Hättenschwiler et al., 1997; Rolland et al., 50

1998, Hember  $et\ al.$ , 2012). Such spatial variability of long-term climate warming may be a more important driver of tree growth than changing  $c_a$  in cold mountain forests (Salzer  $et\ al.$ , 2009). Direct warming effects on growth could thus confound indirect effects due to rising  $c_a$  per se since growth is more sensitive to temperature than photosynthesis across the expected range of temperatures in montane environments (Körner, 2003). Several authors speculate that such fertilization effect of rising  $c_a$  on growth would happen in dry sites where iWUE should be mainly improved (Huang  $et\ al.$ , 2007), whereas others regard mesic sites with high N deposition rates as being the most sensitive sites to any  $c_a$ -enhanced growth (Knapp  $et\ al.$ , 2001; McMahon  $et\ al.$ , 2010; Silva & Anand, 2013).

The consideration of ontogeny is also critical to detect any fertilization effect on growth since trees are organisms continuously adjusting to changing environmental conditions ( $c_a$  and climate) through aging and changing size (Luyssaert et al., 2008; Phillips et al., 2008; Voelker, 2011). Under experimental conditions, trees subjected to warming allocate more carbon to primary than to secondary growth being taller and skinnier than control trees (Way & Oren, 2010). Such allocation rules can shift through ontogeny because young trees tend to display growth stimulation under controlled CO<sub>2</sub> concentrations while mature trees do not (Körner et al., 2005; Norby et al., 1999, 2010; Battipaglia et al., 2013). Consequently, it is essential to consider the individual responses of trees to changing  $c_a$  across wide climatic and biogeographical gradients and encompassing long temporal frameworks (Johnson & Abrams, 2009; Carrer et al., 2010). Dendrochronology may allow tackling these questions, once its shortcomings are adequately treated. First, the ability to detect potential influences of rising  $c_a$ on tree growth may be limited by the use of detrending techniques that remove long-term growth trends (Briffa & Melvin, 2011; Bowman et al., 2013). Second, a biased sampling of big or fast-growing (small or slow-growing) trees could produce spurious rising (declining) growth rates (Cherubini et al., 1998; Brienen et al., 2012). Therefore, it is important to sample coexisting trees of diverse sizes, growth rates and ages. Finally, rising  $c_a$  and temperatures may also exert nonlinear influences on growth (Lloyd *et al.*, 2013) by stimulating them up to certain site- or age-dependent thresholds (Fig. 1).

Here we evaluate these ideas by quantifying the long-term growth responses of *Pinus uncinata* trees to rising  $c_a$  and air temperature. These trees were sampled in 28 forests subjected to different climatic conditions across the entire distribution area of the species in Spain. Specifically, we test two hypotheses: (i) pre-industrial trees (age>150 years) present lower juvenile growth rates than post-industrial trees (age≤150 years) established later and subjected to higher  $c_a$ , and (ii) tree growth will be enhanced in the driest sites where iWUE is improved and growth is not so constrained by low temperatures. Thus, we expect growth stimulation by rising  $c_a$  being more noticeable either in young trees (cf. Voelker *et al.*, 2006) or in relatively dry and mild sites (cf. Huang *et al.*, 2007).

# Material and methods

90 Study site and species

Pinus uncinata Ram. is a long-lived and shade intolerant conifer occurring in subalpine forests from the Alps, the Vosges, the Pyrenees and the Iberian System. Across the study area in NE Spain, *P. uncinata* is subjected to a wide range of topographical and soil conditions and the southernmost limit of its distribution is located in this area (Ceballos & Ruiz de la Torre, 1979). In this species wood formation mainly occurs from May to October (Camarero *et al.*, 1998). We used data of 544 *P. uncinata* trees with complete topographic and size data (see below) selected among 642 living trees sampled from 1994 to 2010 (see details in Galván *et al.*, 2012, 2014). These trees were sampled in 28 forests located throughout the entire geographic range of the species in northern Spain (Supporting Information, Table S1; for a complete description of the studied regions see Galván *et al.*, 2012, 2014). Most sampled sites

(25 out of 28) were located in the Pyrenees where the species is dominant and forms the alpine treeline ecotone, while 2 sites were sampled in the Iberian System and another one in the Pre-Pyrenees. The climate of the sampled populations is continental but varied from sub-Mediterranean (relatively dry summers) to temperate conditions (wet-cool summer) (Supporting Information, Fig. S1). The macroclimate of the study area is strongly influenced by east—west and north—south gradients with increasing warmer conditions eastwards and southwards, whereas dry and continental conditions prevail in the Central Pyrenees (Galván *et al.*, 2012). In the Central Pyrenees mean annual temperatures range between 3.0 and 3.5 °C, whereas total precipitation vary between 1150 and 1280 mm (Supporting Information, Figure S1). Across the Spanish Pyrenees total precipitation amounts increase westwards (about 1500-2500 mm), whereas the driest sites are located southwards either in the Pre-Pyrenees or in the Iberian System (1100-1200 mm). During the past century, the whole study area has experienced a significant warming trend (on average +0.015 °C yr<sup>-1</sup>; see Galván *et al.*, 2014) but no evident changes in the amount of precipitation have been observed (Supporting Information, Fig. S2).

Tree dataset

We reconstructed growth trends for 544 living *P. uncinata* individuals. In each sampled site 10 to 65 dominant trees representative of the population size and age distributions were sampled. Trees that were at least 10 m apart from each other were randomly selected. Topographic (elevation, slope, and aspect) and biometric (dbh, diameter at breast height measured at 1.3 m, and tree height) variables were recorded for each tree. From each sampled tree, two or three cores were taken at 1.3 m height using increment borers. The diameter, excluding bark, was converted to basal area assuming a circular shape of the stem. Cores were visually cross-dated and measured to a precision of 0.01 mm using a LINTAB measuring

device (Rinntech, Heidelberg, Germany). We evaluated the cross-dating and measurement accuracy with the program COFECHA, which calculates cross correlations between individual series of each core and a master chronology, obtained by averaging all measured series in each site except the one being checked (Holmes, 1983).

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In the cases of cores without pith, the distance to the pith was estimated by fitting a template of concentric circles with known radii to the curve of the innermost rings (Norton et al., 1987). This allowed the estimation of the missing radius length to transform it into the number of missing rings. In those trees in which the central core section could not be estimated because the innermost rings did not curve, we used the dbh of each tree to estimate the tree radius without bark. The conversion of the radius length into missing rings was carried out using a subset of cores with pith (n = 41), and considering the innermost 40 rings, by using a regression calculating the mean number of rings (y) for the estimated distance to the pith (x): y = 0.0109 x ( $R^2 = 0.99$ , P < 0.001). After calculating the missing rings to the pith, we estimate the cambial age (at 1.3 m) adding the number of calculated rings to the total number of measured rings. We estimated tree age at 1.3 m following three criteria depending on the method used and the suitability of each core in assessing the tree age: (a) when the innermost part of the core lacks a tree-ring curvature and hence the pith locator method cannot be used, we estimated age as a function of the dbh; (b) when both methods (dbh-age regression and pith locator) have been applied to different cores of the same tree we considered the age estimated by using the pith locator method; and (c) when the pith locator method was used, we chose the age assessed from the core with a smaller estimated distance to the pith. At the same time, we followed three criteria to establish a scale of uncertainty in age estimation: (i) low uncertainty, when the core includes the pith or when the core has an estimated distance to the theoretical pith smaller than 1.5 cm; (ii) medium uncertainty, when cores with estimated distance to the pith was longer than 1.5 cm, and (iii) high uncertainty,

when the innermost tree rings of the core did not curve and tree age was estimated as a function of dbh. On average, *P. uncinata* trees take 20 years to reach a height of 1.3 m (Camarero, 1999). Therefore we added 20 years to the estimated age at 1.3 m to estimate the age in the base of the trunk.

Climate and atmospheric CO<sub>2</sub> data

We used homogeneous and interpolated monthly climate data corresponding to the CRU TS 3.1 database produced by the Climate Research Unit for the period 1901-2009 at a  $0.5^{\circ}$  spatial resolution (Harris *et al.*, 2014). We selected  $0.5^{\circ}$  grids including the study sites and the monthly climatic variables which most growth variability of *P. uncinata* explain, namely previous mean maximum November temperature (hereafter abbreviated as pTMx11), current mean minimum May temperature (hereafter abbreviated as pTMx11) and June precipitation (hereafter abbreviated as pTMx11) (see Tardif *et al.* 2003; Galván *et al.*, 2014). Annual pTMx110 and pTMx111 were computed from monthly values taken from the Mauna Loa (Hawaii) observatory from 1958 onwards (see http://www.esrl.noaa.gov/gmd/ccgg/trends/).

Growth models

The use of measurements such as basal area increment (BAI) partially avoids the problem of removing long-term growth trends since BAI captures the sigmoid trajectory of tree growth (Biondi & Qaedan, 2008). The dynamics of BAI usually display a young phase of exponential growth and a stable phase when trees age and reach height and BAI plateaus (Fritts, 2001). To model tree growth accurately, individual tree-ring width series were converted into BAI, which allows removing the geometrical constraint of adding a volume of wood to a stem of increasing radius (Biondi & Qaedan, 2008) as:

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$$BAI = \pi (r_t^2 - r_{t-1}^2)$$
 (1)

where  $r_t^2$  and  $r_{t-1}^2$  are the squared stem radial increments corresponding to rings formed in years t and t-1, respectively.

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Since we expect that BAI trends and response to  $c_a$  are age dependent, trees were classified in four groups according to their age: young trees (individuals with age  $\leq 150$ years), mature trees (151–300 years old trees), old mature trees (301–450 years old trees) and old trees (individuals with age  $\geq 451$  years). Age classes were defined to represent different life stages and environmentally contrasting periods (e.g., pre- vs. post-industrial periods) but keeping a robust enough sample size for each class. Given that BAI trajectories vary strongly among individuals and within populations flexible statistical methods can be used to represent growth such as Generalized Additive Mixed Models (GAMM; Wood, 2006). This approach allows describing how BAI varies with the additive increase of cambial age or calendar year. BAI depends on site conditions as well as on the particular growth conditions to which each individual tree is subjected along its ontogeny (e.g., Pokharel & Froese, 2009). In order to avoid underrepresentation of some age- (4 classes) and site (28 levels) replications, sites were grouped into regions with similar growth characteristics based on a previous study by Galván et al. (2012). They considered five regions showing coherent geographical, topographical and growth conditions (similar growth trends and responses to climate) shared by trees inhabiting each region (see Table 1).

The GAMM we fitted was of the form:

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$$BAI_i = s(dbh_i) + s(tree \ age_i) + s(cambial \ age) + Z_iB_i + v_i$$
 (2)

where the BAI of a tree i is modelled as smooth functions (s) of three predictors (dbh, tree age, cambial age of tree-ring formation) and tree identity ( $Z_iB_i$ ) considered a random effect since multiple measurements were performed for each individual tree along its life. Since BAI of year t depends on the previous-year (t-1) BAI we also included in the model an error term ( $v_i$ ) with an AR1 (p=1, q=0) correlation structure. The smooth terms were

represented using thin plane regression splines (Wood, 2006). The degree of smoothness is determined through an iterative fitting process (see Wood, 2003). Cambial age refers here to a variable which represents tree ageing from its establishment ( $age\ 0$ ) to the date in which the annual ring was formed ( $age\ n$ ), whereas tree age is a fixed value representing the age of the tree. Therefore, cambial age varies within each tree along time, whereas tree age is a fixed value which changes between trees. Since BAI has a skewed distribution, we first log (BAI +1) - transformed this variable. To test if BAI depends on the particular time period in which each individual tree was established separated models were performed for the trees grouped in different age classes. In addition, BAI trends were characterized for each study region and age class, and also since the 1950s when  $c_a$  started steeply rising.

To reveal whether recent trends in P. uncinata growth are only due to biological features (e.g., tree ageing) or, alternatively, they are mainly influenced by climate and  $c_a$ , we correlated age- and site-specific growth trends over the period 1958-1994 with selected climatic variables (pTMx11, TMi5, P6) and annual values of  $c_a$ . Since we expect age-dependent growth responses to climate, separate analyses were performed for each age class. In addition, we also considered site-mediated growth responses to  $c_a$ . For each age class, the following model was proposed to study the influence of year-to-year variability in climate variables on BAI after removing mid-term fluctuations not related to climate:

$$BAI_{res} = pTMx11s + TMi5 + P6 + c_a + Z_iB_i + \varepsilon_i$$
(3)

where  $BAI_{res}$  represent BAI residuals after extracting biological trends (i.e. after removing the influence of dbh and tree age), and pTMx11, TMi5 and P6 are climatic variables.

Separate analyses were carried out for each site by following a multi-model inference approach based on information theory (Burnham & Anderson, 2002). This approach calculates the probability that a given model is more appropriate than other competing models

in explaining the response variable. Multi-model inference is recommended for analysing large-scale datasets across broad geographical ranges (Zuur *et al.*, 2009). We ranked all the potential models that could be generated with the different explanatory variables according to the second-order Akaike information criterion (AICc). For each model, we considered its  $\Delta$ AICc (i.e. the difference between AICc of each model and the minimum AICc found for all models). After selecting the set of best candidate models (those having a  $\Delta$ AIC < 2) we selected the coefficients for the model with the lowest  $\Delta$ AIC. We evaluated the fit of the model by graphical examination of the residual and fitted values (Zuur *et al.*, 2009). All statistical analyses were performed using the R statistical software (R Development Core Team, 2014). The *mgcv* package was used to calculate the GAMMs (Wood, 2006) and the *MuMIn* package to perform multi model selection (Barton, 2013).

#### Results

The most abundant age classes were formed by trees with 251 up to 300 years old (Fig. 2). During the first years of the tree life BAI values showed a marked increase with cambial age, being larger for young trees, followed by mature, old mature and old trees (Fig. 3a; see also Supporting Information, Table S2). In general, BAI decreased as trees aged, but the slope of the decrease was lower for the old mature and old trees. For instance, at the age of 40 and 80 years, trees in the youngest age class (< 150 years) display significantly larger values of BAI than trees in the rest of age classes (Fig. 3a).

We observed both accelerating and decelerating BAI trends in the first and second halves of the past century, respectively, which mainly corresponded to improved growth of young trees in response to the warm 1940s and 1950s (Fig. 3b). Such growth enhancement seems to be exceptional when compared with other age classes in the context of the past seven

centuries. A slight growth deceleration was observed in all age classes during the second half of the 20<sup>th</sup> century but it was more noticeable in young trees.

The models proposed to study the long-term BAI dynamics showed different patterns depending on the age class, site and time periods considered (Figs. 4 and 5). For example, BAI of young trees is best represented by an exponential function that reaches an asymptote and decreases at age varying from 80 (e.g., Central Pyrenees) to 60 years (e.g., Iberian System and Pre-Pyrenees). Trees in the remaining three classes also show a constant exponential BAI trend that reaches an asymptote, although in several cases (e.g., Central Western Pyrenees) the trend was represented by a straight line. An important part of the BAI variance was accounted for by tree size (dbh) and age (Supporting Information, Table S2). However, the relevance of tree size on determining growth decreased with tree age.

Significant BAI increases were only observed since 1950 in old-mature and mature trees from the Western and Central-Western Pyrenees, and also in old-mature trees from the Iberian System and Pre-Pyrenees (Fig. 5). Negative or low BAI trends were common among the oldest trees and also in all age classes of the Central and Eastern Pyrenees. We found a positive influence of temperature on BAI in the different analysed age classes, particularly in the Central and Eastern Pyrenees (Table 2). Precipitation also enhanced BAI of young and mature trees from the driest sites in the Iberian System and Pre-Pyrenees region.

Contrastingly, we only found a significant and positive influence of rising  $c_a$  on BAI in old mature trees from the wet Western Pyrenees (Table 2), where the highest BAI enhancement since 1950 was also observed (Fig. 5).

## **Discussion**

We found that tree growth responds more to other factors (ontogeny, temperature, water availability) than to rising  $c_a$ . We detected little evidence of widespread growth enhancement

in subalpine *Pinus uncinata* forests in response to increasing  $c_a$  and a related fertilization effect. The much higher growth rates of post-industrial young trees than pre-industrial old trees for the same age confirm our first hypothesis, namely that pre-industrial trees present lower juvenile growth rates than post-industrial trees subjected to higher  $c_a$ . This must be interpreted with caution since higher post-industrial growth rates could be also the result of age-related selection against fast growing individuals irrespective of  $c_a$  trends (Voelker, 2011). The resurgence in radial growth was observed in old mature trees from the wettest study region (Western Pyrenees) and not in young trees from the driest areas (Iberian System and Pre-Pyrenees), where iWUE should be the most benefitted because of rising  $c_a$ , thus contradicting our second hypothesis. Thus, the rapid growth increase of young or post-industrial trees, mostly established in the 19<sup>th</sup> and 20<sup>th</sup> centuries, seems to be a response to rising air temperatures across the study area. These findings have several implications for studies of forest growth as related to global-change drivers such as climatic warming and increasing  $c_a$ .

Firstly, long-term growth stimulation in response to rising  $c_a$  is unlikely in the field because tree growth is more constrained by low temperatures or by low water availability than by accessible carbon (Körner *et al.*, 2005; Körner, 2013). Along these lines, reviews of free-air  $CO_2$  enrichment experiments also reveal that the initial  $CO_2$  stimulation of tree growth under optimal nutrient conditions disappears after one or two decades suggesting ontogenetic limitations (Kimball *et al.*, 2007; Norby & Zak, 2011). Secondly, trees showing higher iWUE values or increments do not necessarily have to present growth enhancement resulting from rising  $c_a$  as has been evidenced at local (Silva *et al.*, 2010; Nock *et al.*, 2011; Linares & Camarero, 2012; Gómez-Guerrero *et al.*, 2013; Lévesque *et al.*, 2014) and global scales (Peñuelas *et al.*, 2011; Silva & Anand, 2013). In the eastern Pyrenees the iWUE reconstructed from carbon isotopic ratios of *P. uncinata* wood doubled during the 20th century whereas

radial growth declined at a mean rate of 0.002 mm yr<sup>-1</sup> (Andreu-Hayles *et al.*, 2011). These authors interpreted such iWUE rise as a response to significant increases in leaf intercellular  $CO_2$  concentration ( $c_i$ ), but it did not suffice to enhance tree growth. Overall, our results do not support any  $c_a$ -enrichment effect even if we studied trees growing at high elevations where low partial pressures of  $CO_2$  could make it more limiting than at lower elevations (Körner *et al.*, 1991). Our findings suggest long-term anatomical and physiological adaptations for adjusting  $c_i$  to increasing  $c_a$  such as changing stomatal density or photosynthetic rates (Ward *et al.*, 2005). Such adjustments could explain why rising  $c_a$  might modify climate sensitivity of trees, but not overall growth rates (Gerhart *et al.*, 2012).

The interactions between temperature and water availability can determine the tree growth responses to rising  $c_a$  (Norby et al., 2010). Rising  $c_a$  should enhance growth as long as water is not limiting; but we found negative correlations between  $c_a$  and basal area increment among all assessed age classes, particularly in old trees. Our findings agree with observations in boreal black spruce (Picea mariana) forests where old trees also responded negatively to rising  $c_a$  (Girardin *et al.*, 2014). Elevated CO<sub>2</sub> conditions increased the maximal photosynthetic rate and iWUE of *Pinus mugo* at the alpine treeline where soil water availability was high (Streit et al., 2014). Consequently, the amount of carbon available to sinks (growth, storage, root exudation, metabolic respiration) increased under those conditions but again this did not translate into enhanced basal area growth (Handa et al., 2006; Dawes et al., 2011, 2013). In fact warming could also induce drought stress and override any potential effect of rising  $c_a$  and improved iWUE on carbon supply and growth if water supply is not adequate to sustain sink demand (Lévesque et al., 2014). This could explain the growth enhancement observed in the Western Pyrenean P. uncinata forests, where climate conditions are wet. In addition, in that area warming was more pronounced in those months most strongly related to growth (Supporting Information, Fig. S2).

The postulated CO<sub>2</sub> fertilization was observed in old mature trees but this effect could be mediated by other features than tree ontogeny per se such as tree height, leaf or sapwood area. In old trees the efficiency of water transport is maintained because xylem elements grow progressively larger with tree age and only in very tall trees the hydrostatic gradient induces limits on xylem functionality towards the top of the tree (Ryan & Yoder, 1997). Since P. *uncinata* is a shade-intolerant species, tree height and competition for light play minor roles as drivers of growth in these open forests (Supporting Information, Fig. S3). Accordingly, tree size was not a significant factor controlling growth in older age classes confirming that once those old trees reach a senescent phase their BAI rates do not depend on tree size. Regarding ontogenic processes, young or small trees could also preferentially assimilate respired CO<sub>2</sub> when growing in closed canopies altering the isotopic composition of their wood and biasing inferred iWUE trends (Bert et al., 1997), but this effect seems to be unlikely since most of the study sites are open stands. Thus, other size-related constrains related to leaf and sapwood production could explain recent BAI trends (Galván et al., 2012). Nevertheless, future growth and iWUE assessments should consider the whole range of tree sizes and ages (Nehrbass-Ahles et al., 2014).

Regarding the roles played by  $c_a$ , temperature and water availability as drivers of long-term growth trends, a survey of studies reporting positive growth responses to rising  $CO_2$  suggests either a potential fertilization effect in fast-growing hardwood species growing in moist sites such as *Populus* trees (Cole *et al.*, 2010; Battipaglia *et al.*, 2013) or improved growth due to improved iWUE, particularly in dry sites (Soulé & Knapp, 2006). In the first case the presumed stimulation of growth has been poorly discussed and could represent initial stimulatory (ontogenetic) effects of growth in rapidly-growing sprouting *Populus* species. In the second case, recent results dismiss the generality of this pattern since declining growth trends have been described in the same conifer species coexisting either in xeric or in mesic

sites (Lévesque  $et\ al.$ , 2014). Despite a reduction in stomatal conductance in response to higer  $c_a$  has been observed in broadleaf species drought-induced stomatal closure diminishes growth in the long term (Lévesque  $et\ al.$ , 2014). Still, in these discussed cases the effects of low temperatures on growth are minor compared with those described in this and other cold-limited forests (Salzer  $et\ al.$ , 2009; Silva & Anand, 2013). Our findings also imply considering warming-related drought stress as a major driver of growth in old trees from mountain forests (Bellassen  $et\ al.$ , 2011). In the Pyrenees a warming trend has characterized the past 700 years (Büngten  $et\ al.$ , 2008). Due to the rise in air temperature, young trees experience longer seasons nowadays than older trees did when they were in their juvenile phase, resulting in large growth rates nowadays (Way & Oren, 2010). Finally, we only considered one tree species so future studies should investigate several coexisting tree species. Those species could grow differently in response to rising  $c_a$  as a function of different photosynthetic and phenological adjustments affecting carbon uptake and wood production.

To conclude, we show that forest growth responses to rising  $c_{\rm a}$  and climate warming depend on site conditions and tree ontogeny and are nonlinear. Our results illustrate how growth of post-industrial P. uncinata trees has improved during the past 150 years. Such growth improvement was linked to climate warming, while a positive response of tree growth to rising  $c_{\rm a}$  was only detected in relatively old trees from wet sites. We argue that growth, stemwood production and productivity will increase as  $c_{\rm a}$  rises in those wet sites, whereas climate warming can directly enhance growth in other sites or even lead to growth decline if rising temperatures amplify water shortage in the most xeric sites. Our findings emphasize the need of using broader perspectives in temporal (ontogeny), biogeographic (site conditions) and also functional (e.g., comparing conifers and broadleaf species) terms to comprehensively evaluate long-term  $c_{\rm a}$  effects on forest growth.

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# Table 1. Number of trees per age class studied in each region and site. See additional site

information in Supporting Information, Table S1.

Region (code)	Site		No. trees					
		Total	Age classes (age)					
			Young trees (≤ 150 years)	Mature trees (151–300	Old mature trees (301– 450 years)	Old trees (≥ 451 years)		
				years)				
Western Pyrenees	Contienda-Larra	20	0	5	13	2		
(WP)	Atxerito	12	0	5	2	5		
Central-Western	Respomuso	14	1	8	4	1		
Pyrenees (CWP)	Pic d'Arnousse	6	1	3	2	0		
Central Pyrenees (CP)	Mirador del Rey	17	1	16	0	0		
	Las Cutas	10	10	0	0	0		
	Bielsa	10	1	7	1	1		
	Senda de Cazadores	32	2	14	9	7		
	Sobrestivo	38	1	9	19	9		
Eastern Pyrenees (EP)	Airoto	14	0	9	3	2		
	Amitges	24	0	7	13	4		
	Conangles	20	1	7	10	2		
	Corticelles	23	0	2	7	14		
	Estanys de la Pera	19	0	10	5	4		
	Estany Gerber	28	1	6	11	10		
	Barranc de Llacs	31	1	12	6	12		
	Lladres	7	0	2	3	2		
	Mata de València	8	0	6	2	0		
	Mirador	31	1	7	14	9		
	Monestero	25	0	9	12	4		
	Estany Negre	40	0	14	10	16		
	Ratera	2	1	1	0	0		
	Sant Maurici	20	0	20	0	0		
	Tessó del Son	9	0	5	1	3		
	Mulleres	10	0	3	2	5		
Iberian System and	Vinuesa	22	1	7	7	7		
Pre-Pyrenees (IS-PP)	Teruel	30	14	8	7	1		
<u>-</u>	Guara	13	4	9	0	0		

**Table 2.** Influence of climatic variables and atmospheric CO<sub>2</sub> concentrations on basal area increment from different regions and considering five age classes (period 1958-1994). For each region, the influence of the annual increment in CO<sub>2</sub> is shown by the t statistic and its associated probability (P). Coefficients were only considered for regions with more than 5 trees per age class. Abbreviations of climatic variables: pTMx11, previous mean maximum November temperature; TMi5, current mean minimum May temperature; P6, current June precipitation. Significant levels:  $P \le 0.05$ ; \*\*  $P \le 0.01$ . Grey cells show significant positive effects of CO<sub>2</sub> on basal area increment.

Age class (age)	Region	pTMx11	TMi5	P6	$CO_2$	No. models	Relative weight
Young trees	Western Pyrenees	_	_	_	_	_	_
$(\leq 150 \text{ years})$	Central-Western Pyrenees	_	_	_	_	_	_
	Central Pyrenees	0.056**	0.019**	-0.018**	_	1	1
	Eastern Pyrenees	_	_	_	_	_	_
	Iberian System and Pre-Pyrenees	0.057**	_	0.037**	-0.006*	1	1
Mature trees	Western Pyrenees	0.036**	-0.006	0.013*	0.006	6	0.98
(151-300	Central-Western Pyrenees	0.031**	0.042**	-0.014	0.003	3	0.99
years)	Central Pyrenees	0.058**	0.015*	0.012*	-0.008**	1	1
	Eastern Pyrenees	0.056**	0.044**	0.018*	-0.005**	1	1
	Iberian System and Pre-Pyrenees	0.044**	0.022*	0.020**	-0.007**	1	1
Old mature	Western Pyrenees	0.009	0.018*	0.009	0.013**	6	0.99
trees (301-	Central-Western Pyrenees	0.021*	0.027*	-0.005	_	6	0.99
450 years)	Central Pyrenees	0.049**	0.029**	-0.007	-0.008**	2	0.99
	Eastern Pyrenees	0.044**	0.043**	0.012*	-0.003**	1	1
	Iberian System and Pre-Pyrenees	0.030**	0.036**	0.003	0.003	3	0.99
Old trees (≥	Western Pyrenees	0.011	0.018	0.011*	_	5	1
451 years)	Central-Western Pyrenees	_	_	_	_	_	_
- ,	Central Pyrenees	0.041**	0.007	-0.018*	-0.012*	2	0.99
	Eastern Pyrenees	0.034**	0.042**	-0.002	-0.002*	2	0.99
	Iberian System and Pre-Pyrenees	0.037**	0.034**	-0.007	_	2	0.99

# Figure legends

Figure 1. Expected responses of photosynthetic rates (a) and growth or basal area increment (b, c, d) to rising atmospheric  $CO_2$  concentrations ( $c_a$ ) and temperatures as a function of either tree ontogeny (cambial age) or time. (c) Postulated ontogenetic trajectories of growth as a function of temperature for three levels of tree performance (low –clear grey–, mid –dark grey– and high performance –black line). Symbols indicate the trajectories of warming-induced growth increase for different performance levels. (d) We expect a departure of the observed basal area increment (BAI) (red continuous lines, circles) as compared with ontogenetically expected BAI (dashed blue line, triangles) in response to rising  $c_a$  (grey area). Panels a, b and c are adapted and modified from Sage & Kubien (2007), Phillips *et al.* (2008), and Way & Oren (2010).

**Figure 2.** Age structure of *Pinus uncinata* trees sampled in the Iberian Peninsula. The age classification was based on 25-year age classes considering three levels of uncertainty in age estimation (see *Material and Methods*).

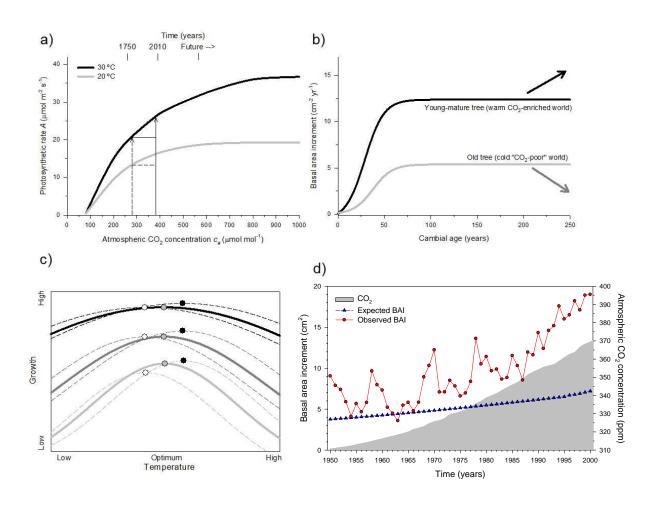
**Figure 3.** Basal area increment (mean  $\pm$  SE) as a function of cambial age (a) and calendar year (b) for the trees in the four age classes considered. The upper and lower insets show basal area increment trends from 20 up to 100 years and for the  $20^{th}$  century, respectively. Different colours of lines indicate different age classes: red lines, young trees (individuals with age  $\leq$  150 years); green lines, mature trees (151–300 years old trees); blue lines, old mature trees (301–450 years old trees); and black lines, old trees (individuals with age  $\geq$  451 years).

**Figure 4.** Long-term trends in basal area increment (BAI, *y* axes) displayed as a function of cambial age (*x* axes) for trees grouped in four age classes (rows) and five study regions (columns; see regions' abbreviations in Table 1). The lines show the fitted GAMMs used to model age-dependent trends in BAI while dots show raw values (the darker the point fills the larger the tree diameter).

**Figure 5.** Recent trends in basal area increment (BAI, *y* axis) of *Pinus uncinata* as a function of year grouped in four age classes (rows) and considering the five study regions (columns; see Table 1). The continuous and dotted lines indicate the observed and fitted BAI trends, while the vertical line shows the first year with data of atmospheric CO<sub>2</sub> concentration used in BAI models. In each graph the correlations show the associations between mean BAI values (black lines) and year (significant trends are written in bold characters). Dots show raw BAI values (the darker the point fills the larger the tree diameter). Note that the fact that the observed BAI exceeds that modelled by GAMMs is an artefact since GAMMs were fitted to a longer period than the most recent 50 years presented in the figure.

# **Figures**

Figure 1



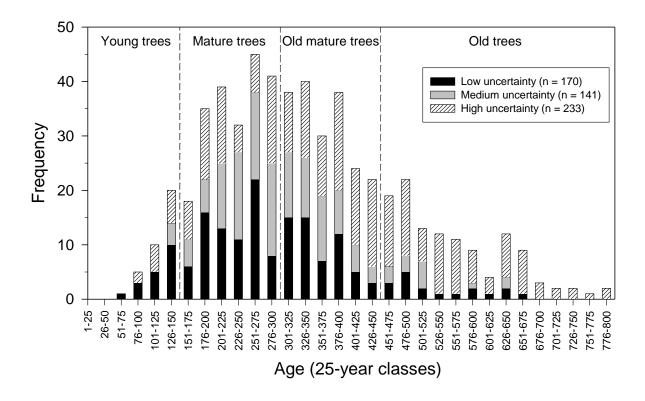


Figure 2

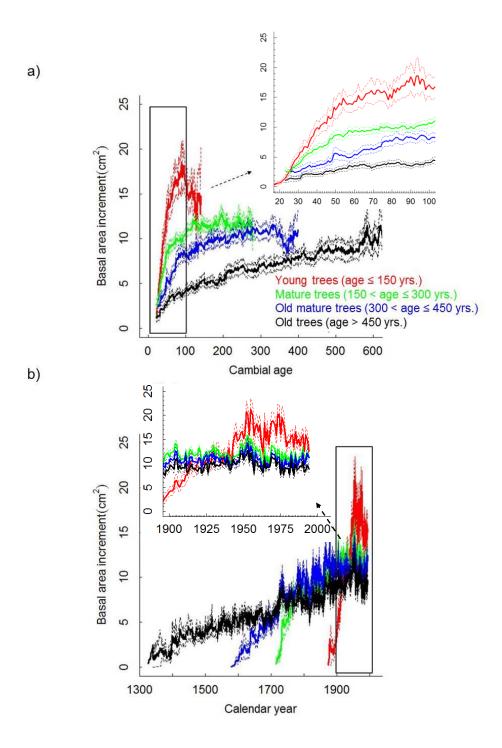


Figure 3

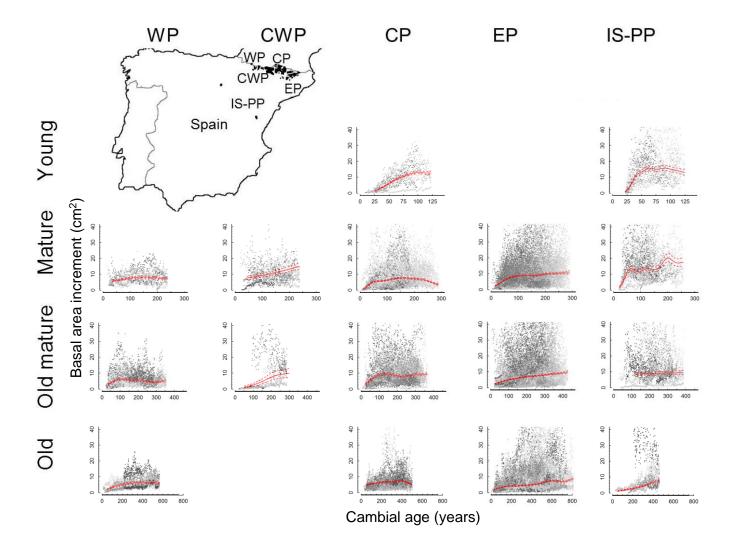


Figure 4

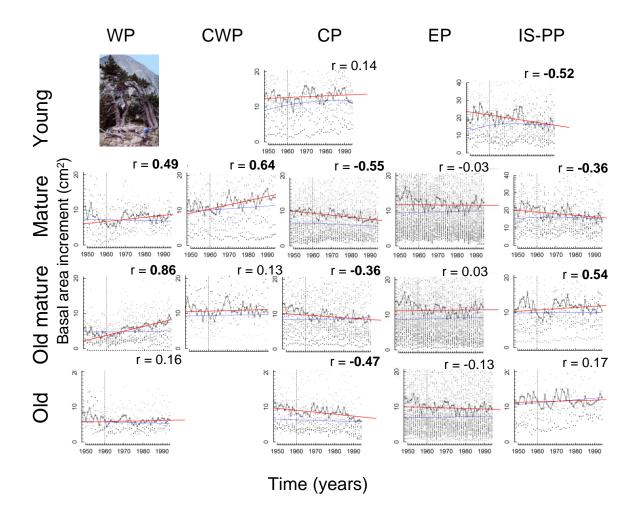


Figure 5