In vitro studies on the tumorigenic potential of halonitromethanes

3

4	Alicia Marsà ¹ , Constanza Cortés ¹ , Elisabet Teixidó ¹ , Alba Hernández ^{1,2} , Ricard
5	Marcos ^{1,2,*}
6	¹ Grup de Mutagènesi, Departament de Genètica i de Microbiologia, Facultat de
7	Biociències, Universitat Autònoma de Barcelona, Cerdanyola del Vallès, Spain.
8	² CIBER Epidemiología y Salud Pública, ISCIII, Spain.

- 9
- 10

11	*Corresponding author at: Grup de Mutagènesi, Departament de Genètica i de
12	Microbiologia, Facultat de Biociències, Universitat Autònoma de Barcelona, Campus de
13	Bellaterra, 08193 Cerdanyola del Vallès (Barcelona), Spain. Phone: +34 93 581 20 52;
14	Fax: +34 93 581 23 87; E-mail address: ricard.marcos@uab.es
15	
16	
17	
18	
19	Running Title: Carcinogenic potential of water disinfection by-products
20	
21	

1

ABSTRACT

23 Epidemiological data indicate that chronic exposure to water disinfection by-products 24 (DBPs) results in an increased risk of cancer in humans. However, the real carcinogenic potential of the compounds within the DBPs is not known. In this study, we assessed the in 25 vitro carcinogenic potential of trichloronitromethane (TCNM) and bromonitromethane 26 (BNM), two unregulated halonitromethanes commonly found in DBP mixtures at 27 28 comparably high concentrations. Human lung BEAS-2B cells were exposed for 8 weeks to environmentally relevant doses of TCNM and BNM. The acquisition of different in vitro 29 cancer-like features was evaluated throughout the exposure. Matrix metalloproteinase 30 (MMP) activities were measured by zymography while colony formation and promotion 31 were assessed by soft-agar assays. Alterations in cellular morphology and proliferation 32 33 were also analysed. The results indicate that long-term exposure to low doses of TCNM and BNM did not cause carcinogenic transformation of BEAS-2B cells as indicated by the 34 absence of morphological changes, no effects on cell growth, no changes in the level of 35 MMP secretion, and no increased anchorage-independent cell growth capacity. 36 Furthermore, TCNM- and BNM-exposed BEAS-2B cells were unable to enhance tumour 37 38 growth directly or by indirect influence on the lung stroma. Our results indicate that the 39 demonstrated carcinogenic effects in human populations exposed to DBP mixtures cannot be attributed to the evaluated halonitromethanes. To our knowledge, this is the first study 40 evaluating TCNM and BNM under a long-term exposure scenario, and using suitable 41 42 hallmarks of the cancer process.

- 43
- 44

Keywords: halonitromethanes; trichloroniromethane; bromonitrometane; cell
 transformation; BEAS-2B cells; cancer biomarkers

47 **1. Introduction**

48 Disinfection by-products (DBPs) are produced by reactions between the organic 49 matter present in raw water and the chemicals used to disinfect it (Ngwenya et al., 2013; Banach et al., 2015). DBPs in drinking water are a complex mixture containing many 50 different chemical groups with a number of compounds having mutagenic and 51 carcinogenic potential, as extensively reviewed (Richardson et al., 2007). Therefore, DBP 52 53 exposure can have important long-term human health implications (Grellier et al., 2015), with bladder cancer as one of the main documented effects (Villanueva et al., 2007; Costet 54 55 et al., 2011; Hrudey et al., 2015).

Due to their potential risk to public health, some DBPs are regulated in many countries, 56 although there are also some chemical species that belong to non-regulated classes which 57 58 could present health risks (Jeong et al, 2015; Kim et al., 2015; Li et al., 2015). Halonitromethanes (HNMs) constitute a highly abundant emerging class of non-regulated 59 DBPs (Weinberg et al., 2002; Krasner et al., 2006; Serrano et al., 2015). These soluble, 60 low molecular weight compounds are produced when chlorine and/or ozone are used for 61 water treatment. They are structurally similar to the halomethanes, but have a nitro-group 62 63 in place of hydrogen bonded to the central carbon atom. The presence of nitrogenous DBPs in drinking water is of great concern due to their higher genotoxicity and cytotoxicity 64 as compared to regulated DBPs (Bond et al. 2011). Trichloronitromethane (TCNM) and 65 bromonitromethane (BNM) are two well-known HNMs. Although TCNM is the most 66 67 common HNM, special attention must be focused on BNM because brominated DBPs are more reactive than their chlorinated forms (Woo et al., 2002; Kim et al., 2015). 68

Genotoxicity assessment is usually accepted as a surrogate biomarker of potential cancer risk. Due to the complexity and ethical issues posed by long-term carcinogenesis studies using mammalian models, *in vitro* genotoxicity assays are used to evaluate potential for carcinogenic risk. This approach has been exploited to determine the 73 genotoxic potential of various DBPs (Richardson et al., 2007; Liviac et al., 2011; Manasfi 74 et al., 2015; Teixidó et al., 2015). Nevertheless, a more direct way to measure the potential carcinogenic risk of individual DBPs remains to be developed. In vitro cell transformation 75 assays have been proposed as suitable alternatives to long-term animal studies to 76 measure carcinogenic effects (Vasseur and Lasne, 2012). In fact, there is increasing 77 78 evidence that the cellular and molecular processes involved in in vitro cell transformation 79 are similar to those that occur during in vivo carcinogenesis (Creton et al. 2012). It is generally accepted that an exposed cell line becomes tumorigenic when a battery of 80 81 different cancer hallmarks are evident (Hanahan and Weinberg, 2011).

It is also interesting to note that the detection of the harmful effects induced by chemical exposures, including those induced by DBPs, are usually measured in unrealistic scenarios of acute treatments at high concentrations. Nevertheless, when the mechanisms of induced carcinogenicity are studied *in vitro*, it is particularly important to design studies involving more relevant types of exposure in terms of human risk; thus, *in vitro* long-term or chronic exposures with low or non-toxic doses would seem to be a good alternative system to assess realistic exposure conditions (Bach et al., 2014).

Using an experimental approach with long-term (8 weeks) exposure and subtoxic concentrations, we evaluated the potential transforming capacity of two HNMs (TCNM and BNM) in a human bronchial epithelial cell line (BEAS-2B), considering inhalation as the main route of exposure to DBPs (Villanueva et al., 2011). Different hallmarks of cell transformation such as morphological cell changes, anchorage-independent cell growth, and secretion of matrix metalloproteinases (MMPs) were analysed.

95

96

97 2. Materials and methods

98 2.1. Cell culture conditions and in vitro DBP exposure.

99 Bromonitromethane (BNM, BrCH₂NO₂; CAS 563-70-2, 90% purity) was purchased from Sigma-Aldrich (St. Louis, MO, USA) and trichloronitromethane (TCNM, CCl₃NO₂; 100 CAS 76-06-2, 97.5% purity) from Riedel-de-Haën (Seelze, Germany). Human bronchial 101 epithelial cells (BEAS-2B), human colorectal carcinoma cell line (HCT116) and human 102 103 lung fibroblast cell line (MRC5) were maintained in DMEM high glucose medium (Life Technologies, NY, USA) supplemented with 10% foetal bovine serum (FBS; PAA®, 104 Pasching, Austria), 1% of non-essential amino acids (NEAA; PAA®) and 2.5 µg/mL 105 plasmocin (InvivoGen, CA, USA) at 37 °C in a humidified 5% CO2 incubator. BEAS-2B 106 cells were long-term exposed to low concentrations of the HNMs (1 and 5 µM BNM; 5 and 107 108 25 µM TCNM) for 8 weeks. The DBP-containing medium was changed every 48 h, and subconfluent cells were passaged weekly. Three separate 75-cm² flasks were maintained 109 for each treatment, and DBP-exposed cells were in all cases compared with unexposed 110 111 passage-matched controls.

112

113 2.2. Cell viability.

BEAS-2B cells were plated in 6-well plates in triplicates at a density of 50,000 cells per well and incubated overnight in complete medium. Cells were then placed in fresh medium with increasing concentrations of 1 to 50 µM of each DBP. After 24 h of treatment, the cells were washed with PBS, and the cell number was assessed after trypsinization by the Beckman counter method with a ZTM Series coulter-counter (Beckman Coulter, CA, USA). Cell toxicity curves were derived from averaging three independent experiments, and the IC₅₀ values were calculated using GraphPad prism version 5.03.

121

122 2.3. Cell proliferation.

BEAS-2B long-term exposed cells and unexposed passage-matched controls were plated in 6-well plates at a density of 50,000 cells per well in regular conditions. At 24 h intervals, the cells were collected by trypsinization and counted by the Beckman counter method. Proliferation was defined as the time necessary for doubling the cell population, and it was calculated according to the equations referred in <u>http://www.doubling-time.com</u> /compute.php

129

130 2.4. Anchorage-independent cell growth.

131 Colony formation in soft-agar was assessed in long-term exposed BEAS-2B cells and passage-matched controls to determine the capacity for anchorage-independent growth 132 133 (Bach et al., 2014). For this study, BEAS-2B cells were collected and individualized using a 30-µm filter. Triplicates of a total of 10,000 cells were suspended in a 1:1:1 mixture 134 containing 1x DMEM (supplemented with 10% FBS, 1% NEAA, 1% L-Glu 200 mM and 2.5 135 136 µg/mL Plasmocin), 2x DMEM (with 2x supplements) and 1.2% Bacto Agar (DIFCO, MD, USA). The mixture was then plated in 6-well plates over a 0.6% base agar (in 137 138 supplemented 2x DMEM). Plates were incubated at 37 °C for 21 days. After this time, cell colonies were stained overnight with 1 mg/mL of (2-p-iodophenyl)-3-3(p-nitrophenyl)-5-139 140 phenyl tetrazolium chloride (Sigma, MO). Plates were scanned and colony counting was 141 performed using OpenCFU, open-source Software. A modified version of the assay was performed using 48-h conditioned medium (CM) from long-term exposed BEAS-2B cells 142 143 and passage-matched controls to assess their capacity to promote the malignant growth of tumour cells. Triplicates of a total of 10,000 individualized HCT116 cells were resuspended 144 145 in CM and mixed in a 1:1:1 ratio with supplemented 2x DMEM and 1.2% bacto agar. The 146 remainder of the protocol was performed as described above. Plates were incubated for 15 days. In normal conditions, transformed cells stimulate the secretion of promoting 147 factors by cells from the stroma. Thus, to further analyse the cancer-like phenotype in 148

long-term exposed BEAS-2B, their 48-h CM was used to grow the lung fibroblast MRC5
cells. Then, the 48-h CM of stimulated MRC5 stroma cells was collected and used for the
soft agar assay with HCT116 cells as indicated above.

152

153 2.5. Secretion of MMP-2 and MMP-9.

154 The activity of secreted matrix metalloproteinases 2 and 9 was examined in long-termexposed BEAS-2B and passage-matched controls. Cells were cultured in basal DMEM 155 medium (without serum or supplements) for 72 h to obtain a CM that was analysed by 156 157 standard zymography, following the manufacturer's instructions (Bio-Rad, Hercules, CA). The area of protease activity was measured densitometrically using the ImageJ analysis 158 program. Since MMPs are mainly secreted by stroma cells, MRC5 fibroblasts were grown 159 in 48-h CM of long-term exposed BEAS-2B and passage-matched controls. Subsequently, 160 the 48-h CM of stimulated MRC5 stroma cells was collected and used for zymography as 161 162 indicated above.

163

164 2.6. Statistical analysis.

One-way analysis of variance followed by Dunnett's multiple comparison test was performed to compare treatments with untreated time-matched controls. In all cases, a two-sided P < 0.05 was considered statistically significant. All statistical analyses were performed using GraphPad Prism 5 version 5.03.

7

170 3. Results

171 3.1. BNM and TCNM are toxic for BEAS-2B cells at low-micromolar levels.

An initial toxicity experiment was carried out using a wide range of compound 172 concentrations to determine appropriate concentrations to study the effects of long-term 173 exposures to non-toxic concentrations of TCNM and BNM. The obtained results showed 174 175 differences in sensitivity to the two compounds (Fig. 1). Thus, BNM was more toxic than 176 TCNM given their IC₅₀ values of 32 \pm 10 μ M and 111 \pm 17 μ M, respectively. Both compounds were clearly cytotoxic at 50 μ M, with a mean viability of 32.4 ± 7.2% and 64.2 177 178 ± 7.4% for BNM and TCNM, respectively, when compared to control cells. Given these results, two different concentrations of each HNM inducing less than a 20% decrease in 179 viability were chosen to carry out the subsequent experiments. The selected doses were 1 180 181 and 5 µM for BNM and 5 and 25 µM for TCNM.

182

3.2. Long-term exposure to HNMs induced no changes in BEAS-2B cellular morphology or proliferation.

Cells undergoing tumoural transformation are known to progressively decrease their 185 doubling times and to start a process in which cells lose their intrinsic morphological 186 characteristics, known as anaplasia. To determine whether the long-term exposure to 187 TCNM or BNM induces tumoural transformation, cellular proliferation and shape were 188 189 monitored as markers of cancer-like phenotypic changes throughout 8 weeks of long-term 190 treatment. No noticeable changes in cell morphology were observed throughout the whole exposure time (Fig. 2A), and results analysed at the end of the exposure period indicated 191 that neither BNM nor TCNM treatment increased the proliferation rates significantly when 192 compared to time-matched controls (Figs. 2B and 2C). 193

194

195 3.3. Long-term exposure to HNMs did not alter BEAS-2B cell capacity to grow in soft agar.

196	One of the main characteristics of cancer cells is their anchorage-independent growth	
197	capacity (Hanahan and Weinberg 2011). To assess this feature in the HNM-long-term	
198	exposed cells, BEAS-2B cells were subjected to a soft-agar colony formation assay.	
199	BEAS-2B cells formed a considerable number of small-sized colonies spontaneously,	
200	with an average size of 143.49 \pm 0.20 $\mu m.$ Nevertheless, important variations were not	
201	observed after the HNM treatments (Fig. 3). Representative pictures of the colony plates	
202	are indicated in Figure 3A where no significant changes are apparent. As shown, control	
203	samples formed a mean of 273 \pm 21 colonies, while cells treated with BNM long-term-	
204	exposed cells displayed 332 \pm 16 and 246 \pm 40 colonies for its lowest and highest	
205	concentrations, respectively. On the other hand, TCNM long-term-exposed cells formed an	
206	average of 288 ± 16 colonies at 5 μM concentration and 157 ± 25 colonies at 25 $\mu M.$ When	
207	colony sizes were taken into account (

Con formato: Fuente: (Predeterminada) Arial, Inglés (Reino Unido)

208

Figure 3Fig. 3C), the mean size was similar in all cases (142.65 \pm 1.51 μ m for controls vs. 145.68 \pm 1.58 and 145.06 \pm 0.77 μ m for BNM 1 and 5 μ M treatments, respectively, and 141.84 \pm 0.74 and 142.24 \pm 1.81 μ m for TCNM 5 and 25 μ M treatments, respectively).

212

213 3.4. Long-term exposure to HNMs did not alter the BEAS-2B cell secretome.

It is known that the interplay between the different cell populations and their environment determine the formation and malignancy of a tumour. Cells undergoing malignant transformation secrete paracrine signals that enhance the tumorigenic potential of nearby cells (del Pozo Martin et al. 2015). Thus, the capability of long-term exposed BEAS-2B cells in promoting the growth of the tumoural cell line HCT116 in soft-agar was evaluated. As observed in Figure 4, no differences in HCT116 colony number or size were observed after the exposure. Cells grown with untreated BEAS-2B CM displayed an average of 1,338 ± 44 colonies. Similar mean values were displayed by HCT116 cells grown in CM of long-term exposed BEAS-2B, showing an average of 1,382 ± 91 and 1,329 ± 62 for the CM of 1 and 5 μ M BNM, and 1,309 ± 68 and 1,309 ± 61 colonies for the CM of 5 and 25 μ M TCNM.

As CM contains cell-secreted factors known to enhance and not initiate tumour 225 226 effects, the comparisons of big colonies or colony sizes are of significance when assessing the promotion of anchorage-independent cell growthThe average size of HCT116 cells 227 exposed to control CM was 216.00 ± 1.17 µm, whereas HCT116 colonies grown in HNM-228 229 exposed BEAS-2B CM showed mean values of 214.70 ± 2.84 and 213.90 ± 4.34 µm for 1 and 5 μ M BNM, and 210 \pm 2.74 and 206.53 \pm 5.70 μ m for 5 and 25 μ M TCNM, 230 231 respectively. There were no significant differences in the percentage of big colonies 232 among the treatments (Fig. 4C).

233

234 3.5. Long-term exposure to HMNs did not influence the lung stroma secretome.

Another method by which transformed cells induce malignant transformation is to 235 236 prompt stroma cells to secrete growth factors that potentiate tumour effects (Cammarota and Laukkanen, 2016). To assess whether BNM or TCNM could enhance tumour growth 237 238 through this indirect mechanism, MRC5 lung stroma fibroblasts grown in HNM long-term-239 exposed BEAS-2B CM were evaluated by the soft agar assay with HCT116 cells (Fig. 5). As shown, HCT116 cells grown in CM of control MRC5 presented an average number of 240 colonies of 1,822 ± 119, non-significantly different than the number of colonies formed by 241 HCT116 cells grown in CM of exposed MRC5 cells, which was $1,848 \pm 127, 2,306 \pm 149$, 242 1,889 ± 178 and 2,171 ± 204 for 1 μ M BNM, 5 μ M BNM, 5 μ M TCNM and 25 μ M TCNM, 243 244 respectively. When mean colony sizes were analysed, the results consistently showed no HNM-associated effect. Thus, HCT116 cells grown in CM of control MRC5 cells showed a 245 246 mean colony size of 198.61 \pm 2.15 μ m, whereas colony sizes of HCT116 cells grown in 247 CM of exposed MRC5 cells were 199.25 ± 1.66 , 201.06 ± 4.74 , 202.19 ± 2.83 and 200.56248 $\pm 2.24 \mu m$ for 1 μM BNM, 5 μM BNM, 5 μM TCNM and 25 μM TCNM, respectively. The 249 CM of MRC5-exposed cells was also unable to induce changes in the percentage of big 250 colonies (Fig. 5C).

251

3.6. Long-term exposure to HNMs induced no changes in the secretion of MMP-2 andMMP-9.

Extracellular proteinases, such as matrix metalloproteinases (MMPs), are known to 254 255 influence cancer cell invasion at a local and distant level (Kessenbrock et al., 2010). For 256 this reason, increased MMP activity levels are generally used as indicators of cell malignant transformation, especially in the case of MMP2 and MMP9. No significant 257 changes were observed when the secreted MMP 2+9 were analysed in HNM long-term-258 exposed BEAS-2B cells compared to control cells (Fig. 6A). As MMPs are mainly secreted 259 260 by stroma cells after stimulation by tumoral cells, we found it interesting and necessary to confirm these negative results by analysing the secretion of MMP 2+9 in MRC5 cells 261 grown in CM of long-term-exposed BEAS-2B cells. Consistent with previous results, no 262 changes in MMP secretion were observed (Fig. 6B). 263

264

265

267 4. Discussion

DBPs constitute a very complex mixture of chemicals. The amount and variety of their chemical components depend on both the disinfection procedure used and the characteristics of the raw water source. So far, more than 600 DBPs have been discovered, with many of them showing different levels of toxicity, genotoxicity and carcinogenic properties (Richardson et al., 2007).

273 Although there is ample epidemiological evidence linking cancer risk to DBP exposure (Hrudey et al., 2015; Villanueva et al., 2007, 2015), none of the individual species tested to 274 275 date in the limited assays employed has shown sufficient carcinogenic potency to account for the cancer risks projected from the epidemiological studies (Bull et al., 2011). This 276 implies that a more systematic approach to determine the carcinogenic potential of DBPs 277 278 is required. In this sense, our study analysing the carcinogenic potential of two HNMs is an attempt to provide new information to fill in this gap. Our data suggests that long-term 279 exposures to low doses of BNM and TCNM are not able to trigger any of the in vitro cancer 280 phenotypic hallmarks analysed in lung epithelial cells. This would support the view that 281 exposure to these particular DBPs does not increase the carcinogenic risk in humans. 282

283 There are very few in vitro studies together with several in vivo studies dealing with the carcinogenic risk of DBPs. Among the in vivo studies, McDorman et al. (2003) 284 evaluated four DBPs by using rats phenotypically prone to the effects of renal carcinogens. 285 286 Potassium bromate. 3-chloro-4-(dichloromethyl)-5-hydroxy-2(5H)-furanone (MX). 287 chloroform, and bromodichloromethane were tested. No significant increases in renal adenomas and carcinomas were detected. Nevertheless, elevated hyperplasia and 288 289 tumours were observed when rats were exposed to the highest dose of chloroform, as well as in the mixture containing the highest doses of the four DBPs tested. In another study 290 from the same group using the same DBP compounds, transitional epithelial hyperplasia 291 292 and cell proliferation in the urinary bladder were observed in rats exposed to the highest dose of MX. Furthermore, treatment with the four individual DBPs and the mixture induced the development of aberrant crypt foci, considered as putative preneoplastic colon cancer lesions (McDorman et al., 2003b). A third study evaluating the effects of dichloroacetic acid (DCA) in early life observed that DCA is carcinogenic in mice, inducing increases in the incidence of hepatocellular tumours comparable to those of life-long exposures (Wood et al., 2015).

Although the in vivo studies are an accurate approximation of the actual tumour-299 inducing scenario, the high costs and complexity of these type of studies make them non-300 301 applicable to the vast number of potential genotoxic/carcinogenic DBP compounds. This is the basis of our proposal to use in vitro cell transformation assays. Indeed, the 302 303 performance of such assays in predicting carcinogenic potential has been established on 304 several hundreds of chemicals. In fact, the Organisation for Economic Co-operation and Development (OECD) has specific guidelines on "Cell transformation assays for the 305 306 detection of chemical carcinogens" (Vasseur and Lasne, 2012), with accumulated evidence that the cellular and molecular processes involved in in vitro cell transformation 307 308 are similar to those occurring during in vivo carcinogenesis (Creton et al., 2012). In spite of the advantages of these in vitro approaches, the limited number of studies carried out with 309 310 DBPs is surprising, mainly taking into account the environmental risks posed by these 311 chemicals. In one of the few available studies, the ability of MX to promote cell transformation was determined in mouse embryonic fibroblasts. MX was used in both the 312 313 initiation and the promotion phase, after a previous exposure to 3-methylcholanthrene (MC). When MX was added during the promotion phase in the MC-initiated cells, it 314 promoted the development of the transformation foci in a dose-dependent manner. 315 316 Exposure to MX as the initiator slightly enhanced the development of foci, suggesting that MX may rather act via promoting tumour development (Laaksonen et al., 2001). In a 317 318 posterior study, different chlorohydroxyfuranones such as MX, MCA [3,4-dichloro-5-

hydroxy-2(5H)-furanone], CMCF [3-chloro-4-(chloromethyl)-5-hydroxy-2(5H)-furanone], 319 and MCF [3-chloro-4-methyl-5-hydroxy-2(5H)-furanone] were tested to demonstrate their 320 ability to promote foci formation in the two-stage cell transformation assay in BALB/c3T3 321 cells. The results indicate that MX acted as a potent inhibitor of gap-junctional intercellular 322 communication. This action was associated with its ability to promote malignant foci 323 324 formation (Hakulinen et al., 2004). In addition, four DBPs namely tribromomethane, bromochloroacetic acid, dibromonitromethane and tribromonitromethane were tested for 325 their potential to transform normal human colonocytes into malignant cells. The results 326 327 indicated that all DBP-exposed colon cells acquired the ability to grow in soft agar to some extent; however, only cells exposed to tribromomethane were able to grow in media 328 329 lacking serum and enriching growth factors (DeAngelo et al. 2007). Finally, the cell-330 transforming potential of the two unregulated DBPs iodoacetic acid (IAA) and iodoform (IF) were tested in mouse NIH3T3 cells. Exposure to IAA increased the frequencies of cells 331 332 with anchorage-independent growth abilities. In addition, IAA-transformed cells were found to form aggressive fibrosarcomas after inoculation into Balb/c nude mice (Wei et al., 2013). 333 334 The above-mentioned studies show the usefulness of cell transforming assays to 335 demonstrate the tumorigenic potential of DBPs. Nevertheless, due to the complexity of the 336 cell transformation process, it is preferable to investigate the acquisition of a wide range of 337 cancer-like phenotypic features to decipher carcinogenic risk potential. This is the basis of the strategy used in our study. Morphological changes in the exposed cells are considered 338 339 to be typical cancer-like phenotypic indicators (Clancy et al., 2012). In addition, secretion of matrix metalloproteinases (mainly MMP2 and MMP9) is also considered an appropriate 340 biomarker of cell transformation since MMPs play crucial roles in tumour invasion, 341 342 morphogenesis, angiogenesis, metastasis, and wound healing by remodelling the extracellular matrix (Oum'hamed et al., 2004). Moreover, assessment of the anchorage-343 independent growth of cells in semi-solid matrices is considered an intrinsic property of 344

transformed cells (Borowicz et al., 2014). The paracrine signals secreted by cells undergoing malignant transformation enhance the tumorigenic potential of nearby cells cells (del Pozo Martin et al., 2015). Therefore, conditioned media obtained from these cell cultures can be used to determine their ability to induce anchorage-independent growth of tumour or stromal cells. In this way, the usefulness of an indirect soft agar approach has recently been demonstrated (Bach et al., 2016).

As a summary, a battery of *in vitro* cancer-like phenotypic hallmarks must be able to clearly predict the carcinogenic potential of a given environmental chemical. On this basis, the lack of observed (i) morphological and proliferation changes, (ii) variations in the secretion of matrix metalloproteinases, and (iii) anchorage-independent cell growth ability, both directly and indirectly, demonstrate that the two selected HNMs should be considered as non-carcinogenic DBPs.

357

358

361 Conflict of interest statement

- 362 The authors report no conflict of interest. The authors alone are responsible for the content
- and writing of the paper.
- 364

365 Acknowledgements

- 366 C.C. and E.T. were supported by postdoctoral fellowships from the Universitat Autonoma
- de Barcelona (UAB). A.M. was supported by UAB PIF fellowships. This work was partially
- supported by the 'Generalitat de Catalunya' (2014SGR-202).
- 369
-
- 370
- 371

372 References

- Bach, J., Sampayo-Reyes, A., Marcos, R., Hernández, A., 2014. *Ogg1* genetic
 background determines the genotoxic potential of environmentally relevant arsenic
 exposures. Arch. Toxicol. 88, 585-596.
- Bach, J., Peremartí, J., Annangi, B., Marcos, R., Hernández, A., 2016. Oxidative DNA
 damage enhances the carcinogenic potential of chronic arsenic exposure. Arch.
 Toxicol. DOI: 10.1007/s00204-015-1605-7 [Epub ahead of print].
- Banach, J.L., Sampers, I., Van Haute, S., van der Fels-Klerx, H.J., 2015. Effect of
 disinfectants on preventing the cross-contamination of pathogens in fresh produce
 washing water. Int. J. Environ. Res. Public Health 12, 8658-8677.
- Bond, T., Huang, J., Templeton, M.R., Graham, N., 2011. Occurrence and control of
 nitrogenous disinfection by-products in drinking water -a review. Water Res. 45, 4341 4354.
- Borowicz, S., Van Scoyk, M., Avasarala, S., Karuppusamy Rathinam, M.K., Tauler, J.,
 Bikkavilli, R.K., Winn, R.A., 2014. The soft agar colony formation assay. J. Vis. Exp.
 92, e51998.
- Bull, R.J., Reckhow, D.A., Li, X., Humpage, A.R., Joll, C., Hrudey, S.E., 2011. Potential
 carcinogenic hazards of non-regulated disinfection by-products: haloquinones, halo cyclopentene and cyclohexene derivatives, N-halamines, halonitriles, and heterocyclic
 amines. Toxicology 286, 1-19.
- Cammarota, F., Laukkanen, M.O., 2016. Mesenchymal stem/stromal cells in stromal
 evolution and cancer progression. Stem Cells Int. 2016, 4824573.
- Clancy, H.A., Sun, H., Passantino, L., Kluz, T., Muñoz, A., Zavadil, J., Costa, M., 2012.
 Gene expression changes in human lung cells exposed to arsenic, chromium, nickel or vanadium indicate the first steps in cancer. Metallomics 4, 784-793.
- Costet, N., Villanueva, C.M., Jaakkola, J.J., Kogevinas, M., Cantor, K.P., King, W.D.,
 Lynch, C.F., Nieuwenhuijsen, M.J., Cordier, S., 2011. Water disinfection by-products
 and bladder cancer: is there a European specificity? A pooled and meta-analysis of
 European case-control studies. Occup. Environ. Med. 68, 379-385.
- 401 Creton, S., Aardema, M.J., Carmichael, P.L., Harvey, J.S., Martin, F.L., Newbold, R.F.,
 402 O'Donovan, M.R., Pant, K., Poth, A., Sakai, A., Sasaki, K., Scott, A.D., Schechtman,
 403 L.M., Shen, R.R., Tanaka, N., Yasaei, H., 2012. Cell transformation assays for
 404 prediction of carcinogenic potential: state of the science and future research needs.
 405 Mutagenesis 27, 93-101.
- DeAngelo, A.B., Jones, C.P., Moyer, M.P., 2007. Development of normal human colon cell
 cultures to identify priority unregulated disinfection by-products with a carcinogenic
 potential. Water Sci. Technol. 56, 51-55.

- 409 Del Pozo Martin, Y., Park, D., Ramachandran, A., Ombrato, L., Calvo, F., Chakravarty, P.,
 410 Spencer-Dene, B., Derzsi, S., Hill, C.S., Sahai, E., Malanchi, I., 2015. Mesenchymal
 411 cancer cell-stroma crosstalk promotes niche activation, epithelial reversion, and
 412 metastatic colonization. Cell Rep. 13, 2456-2469.
- 413 Grellier, J., Rushton, L., Briggs, D.J., Nieuwenhuijsen, M.J., 2015. Assessing the human
 414 health impacts of exposure to disinfection by-products -a critical review of concepts
 415 and methods. Environ. Int. 78, 61-81.
- 416 Hanahan, D., Weinberg, R., 2011. Hallmarks of cancer: the next generation. Cell 144, 646-417 674.
- Hakulinen, P., Mäki-Paakkanen, J., Naarala, J., Kronberg, L., Komulainen, H., 2004.
 Potent inhibition of gap junctional intercellular communication by 3-chloro-4-(dichloromethyl)-5-hydroxy-2(5H)-furanone (MX) in BALB/c 3T3 cells. Toxicol. Lett.
 151, 439-449.
- Hrudey, S.E., Backer, L.C., Humpage, A.R., Krasner, S.W., Michaud, D.S., Moore, L.E.,
 Singer, P.C., Stanford, B.D., 2015. Evaluating evidence for association of human
 bladder cancer with drinking-water chlorination disinfection by-products. J. Toxicol.
 Environ. Health. B Crit. Rev. 18, 213-241.
- Jeong, C.H., Postigo, C., Richardson, S.D., Simmons, J.E., Kimura, S.Y., Mariñas, B.J.,
 Barcelo, D., Liang, P., Wagner, E.D., Plewa, M.J., 2015. Occurrence and comparative
 toxicity of haloacetaldehyde disinfection byproducts in drinking water. Environ. Sci.
 Technol. 49, 13749-13759.
- 430 Kessenbrock, K., Plaks, V., Werb, Z., 2010. Matrix metalloproteinases: regulators of the 431 tumor microenvironment. Cell 141, 52-67.
- Kim, D., Amy, G.L., Karanfil, T., 2015. Disinfection by-product formation during seawater
 desalination: A review. Water Res. 81, 343-345.
- Krasner, S.W., Weinberg, H.S., Richardson, S.D., Pastor, S.J., Chinn, R., Sclimenti, M.J.,
 Onstad, G.D., Thruston, A.D. Jr., 2006. Occurrence of a new generation of disinfection
 byproducts. Environ. Sci. Technol. 40, 7175-7185.
- Laaksonen, M., Mäki-Paakkanen, J., Komulainen, H., 2001. Enhancement of 3methylcholanthrene-induced neoplastic transformation by 3-chloro-4-(dichloromethyl)5-hydroxy-2(5H)-furanone in the two-stage transformation assay in C3H 10T1/2 cells.
 Arch. Toxicol. 75, 613-617.
- Li, J., Wang, W., Moe, B., Wang, H., Li, X.F., 2015. Chemical and toxicological characterization of halobenzoquinones, an emerging class of disinfection byproducts. Chem. Res. Toxicol. 28, 306-318.
- Liviac, D., Creus, A., Marcos, R., 2011. Mutagenic analysis of six disinfection by-products
 in mouse lymphoma cells. J. Hazard. Mat. 190, 1045-1052.

- Manasfi, T., De Méo, M., Coulomb, B., Di Giorgio, C., Boudenne, J.L., 2015. Identification
 of disinfection by-products in freshwater and seawater swimming pools and evaluation
 of genotoxicity. Environ. Int. 88, 94-102.
- McDorman, K.S., Hooth, M.J., Starr, T.B., Wolf, D.C. 2003a. Analysis of preneoplastic and
 neoplastic renal lesions in Tsc2 mutant Long-Evans (Eker) rats following exposure to
 a mixture of drinking water disinfection by-products. Toxicology 187, 1-12.
- McDorman, K.S., Chandra, S., Hooth, M.J., Hester, S.D., Schoonhoven, R., Wolf, D.C.,
 2003b. Induction of transitional cell hyperplasia in the urinary bladder and aberrant
 crypt foci in the colon of rats treated with individual and a mixture of drinking water
 disinfection by-products. Toxicol. Pathol. 31, 235-242.
- 456 Ngwenya, N., Ncube, E.J., Parsons, J., 2013. Recent advances in drinking water
 457 disinfection: successes and challenges. Rev. Environ. Contam. Toxicol., 222, 111 458 170.
- Oum'hamed, Z., Garnotel, R., Josset, Y., Trenteseaux, C., Laurent-Maquin, D., 2004.
 Matrix metalloproteinases MMP-2, -9 and tissue inhibitors TIMP-1, -2 expression and secretion by primary human osteoblast cells in response to titanium, zirconia, and alumina ceramics. J. Biomed. Mater. Res. A., 68, 114-122.
- 463 Richardson, S.D., Plewa, M.J., Wagner, E.D., Schoeny, R., Demarini, D.M., 2007.
 464 Occurrence, genotoxicity, and carcinogenicity of regulated and emerging disinfection
 465 by-products in drinking water: a review and roadmap for research. Mutat. Res., 636,
 466 178-242.
- Serrano, M., Montesinos, I., Cardador, M.J., Silva, M., Gallego, M., 2015. Seasonal
 evaluation of the presence of 46 disinfection by-products throughout a drinking water
 treatment plant. Sci. Total Environ. 517, 246-258.
- Teixidó, E., Piqué, E., Gonzalez-Linares, J., Llobet, J.M., Gómez-Catalán, J., 2015.
 Developmental effects and genotoxicity of 10 water disinfection by-products in zebrafish. J. Water Health, 13, 54-66.
- Vasseur, P., Lasne, C., 2012. OECD Detailed Review Paper (DRP) number 31 on Cell
 transformation assays for detection of chemical carcinogens: main results and
 conclusions. Mutat. Res., 744, 8-11.
- Villanueva, C.M., Cantor, K.P., Grimalt, J.O., Malats, N., Silverman, D., Tardon, A., GarciaClosas, R., Serra, C., Carrato, A., Castaño-Vinyals, G., Marcos, R., Rothman, N.,
 Real, F.X., Dosemeci, M., Kogevinas, M., 2007. Bladder cancer and exposure to
 water disinfection by-products through ingestion, bathing, showering, and swimming in
 pools. Am. J. Epidemiol., 165, 148-156.
- Villanueva, C.M., Gracia-Lavedán, E., Ibarluzea, J., Santa Marina, L., Ballester, F., Llop,
 S., Tardón, A., Fernández, M.F., Freire, C., Goñi, F., Basagaña, X., Kogevinas, M.,
 Grimalt, J.O., Sunyer, J.; INMA (Infancia y Medio Ambiente) Project, 2011. Exposure
 to trihalomethanes through different water uses and birth weight, small for gestational
 age, and preterm delivery in Spain. Environ. Health Perspect., 119, 1824-1830.

- Villanueva, C.M., Cordier, S., Font-Ribera, L., Salas, L.A., Levallois, P., 2015. Overview of
 disinfection by-products and associated health effects. Curr. Environ. Health. Rep., 2,
 107-115.
- Wei, X., Wang, S., Zheng, W., Wang, X., Liu, X., Jiang, S., Pi, J., Zheng, Y., He, G.,
 Qu, W., 2013. Drinking water disinfection byproduct iodoacetic acid induces

491 tumorigenic transformation of NIH3T3 cells. Environ. Sci. Technol., 47, 5913-5920.

Weinberg, H.S., Krasner, S.W., Richardson, S.D., Thurston, A.D. Jr., 2002. The ocurrence
of disinfection by-products (DBPs) on health concern in drinking water: Results of a
Nationwide DBP occurrence study, EPA/600/R02/068, U.S. Environmental Protection
Agency, National Exposure Research Laboratory, Athens, G.A.

- Woo, Y.Y., Lai, D., McLain, J.L., Manibusan, M.K., Dellarco, V., 2002. Use of mechanism based structure-activity relationship analysis in carcinogenic potential ranking for
 dinking water disinfection by-products. Environ. Health Pespect. 110, 75-87.
- 499 Wood, C.E., Hester, S.D., Chorley, B.N., Carswell, G., George, M.H., Ward, W., Vallanat,
- 500 B., Ren, H., Fisher, A., Lake, A.D., Okerberg, C.V., Gaillard, E.T., Moore, T.M., 501 Deangelo, A.B., 2015. Latent carcinogenicity of early-life exposure to dichloroacetic
- 501 Deangelo, A.B., 2015. Latent carcinogenic 502 acid in mice. Carcinogenesis, 36, 782-791.

503

505 Figure legends

506

Figure 1. BEAS-2B cell viability curves after 24 h of exposure to increasing concentrations of BNM and TCNM, ranging 1 to 50 μ M. Data are presented as mean values of independent experiments ± SEM (*n*=3).

510

Figure 2. Long-term exposure to BNM and TCNM does not induce major changes in cancer phenotypic markers as cellular proliferation and cell morphology in BEAS-2B cells.

A) Representative pictures of cell morphology taken 24, 48 and 72 h after seeding. B-C)
 BEAS-2B doubling time mean values ± SEM.

515

Figure 3. Analyzed HNMs do not induce cells anchorage-independent growth after a long term exposure. A) Representative images of BEAS-2B colonies in soft agar. B) Mean

518 number of colonies per well \pm SEM. C) Mean colony size per treatment \pm SEM (*n*=12). 519 ***P*<0.01 compared with time-matched controls.

520

20

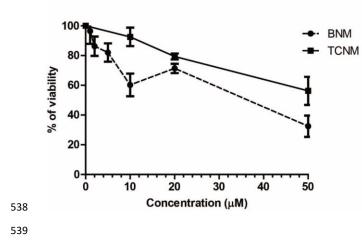
Figure 4. BEAS-2B cells long-term exposed to DBPs do not increase anchorageindependent growth capacity in HCT116 tumor cells. A) Representative images of HCT116 colonies exposed to BEAS-2B conditioned medium in soft agar. B) Mean number of colonies per plate \pm SEM. C) Percentage of colonies bigger than 200 and 300 µm per well formed by HCT116 cells exposed to BEAS-2B CM \pm SEM (*n*=12).

526

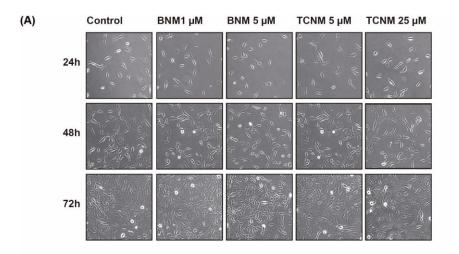
Figure 5. MRC5 fibroblasts grown in exposed BEAS-2B conditioned medium do not stimulate anchorage-independent growth capacity of HCT116 tumor cells. **A)** Representative images of HCT116 cells exposed to MRC5 CM in soft agar. **B)** Mean number of HCT116 colonies per well ± SEM. **C)** Percentage of big HCT116 colonies per well ± SEM (*n*=12).

532

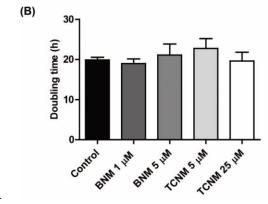
Figure 6. Cells long-term exposed to HNMs do not acquire invasive capacity. **A)** BEAS-2B cells secretion of MMP 2+9 does not increase with long-term BNM nor TCNM exposure when compared to time-matched non-exposed controls. **B)** MRC5 fibroblasts do not show an increase in MMP 2+9 secretion when stimulated with exposed-BEAS-2B CM.







(C)

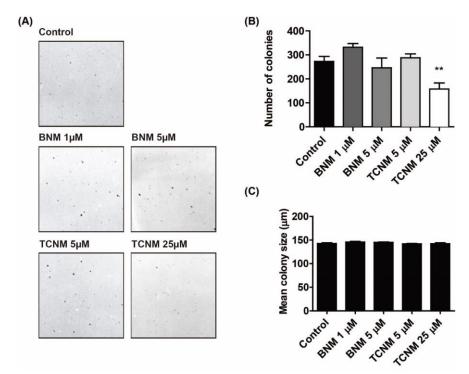


	DT ± SEM (h)
Control	19.88 ± 0.64
BNM 1 µM	18.96 ± 1.17
BNM 5 µM	21.09 ± 1.63
TCNM 5 µM	22.74 ± 1.49
TCNM 25 μM	19.61 ± 1.50

- _

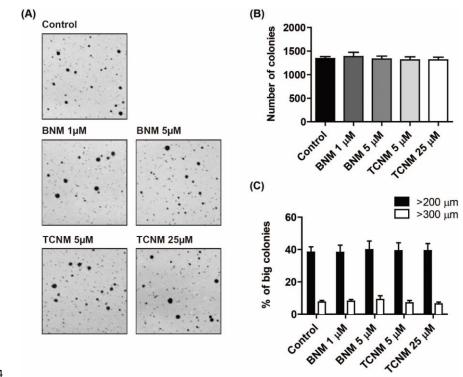
544 Figure 2





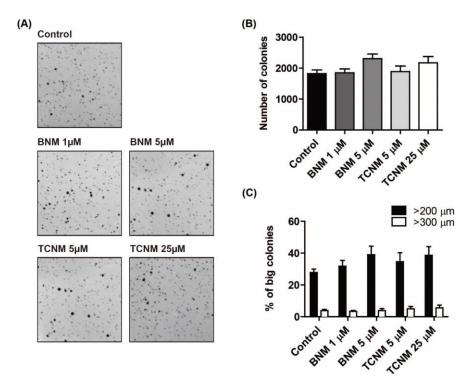
550 Figure 3





556 Figure 4





560 Figure 5







