- 1 Combining Fiber Optic (FO-DTS), Cros Hole ERT and time lapse formation electrical conductivity
- 2 to characterize and monitor a coastal aquifer.
- 3
- 4 Folch, A.^{1,2}, L. del Val^{1,2}, L. Luquot^{3,4,2}, L. Martínez-Pérez^{3,2}, F. Bellmunt⁵, H. Le Lay⁶, V. Rodellas⁷, N.
- 5 Ferrer^{1,2}, A. Palacios^{3,2}, S. Fernández^{3,2}, M. A. Marazuela^{3,2}, M. Diego-Feliu⁷, M. Pool^{3,2}, T.
- 6 Goyetche^{3,2}, J. Ledo⁵, P. Pezard⁴, O. Bour⁶, P. Queralt⁵, A. Marcuello⁵, J. Garcia-Orellana^{7,8}, M.W.
- 7 Saaltink^{1,2}, E. Vazquez-Suñe^{3,2} and J. Carrera^{3,2}
- 8
- ¹ Department of Civil and Environmental Engineering (DECA), Universitat Politécnica de Catalunya,
 Barcelona, Spain.
- 11 ² Associated Unit: Hydrogeology Group (UPC-CSIC).
- 12 ³ Institute of Enviromental Assessment and Water Research, CSIC, Barcelona, Spain
- ⁴ Laboratoire Géosciences Montpellier, UMR 5243, Montpellier, France.
- ⁵ Institut de Recerca Geomodels, Universitat de Barcelona, Spain.
- 15 ⁶ Geosciences Rennes, University Rennes, Rennes, France.
- ⁷ Departament de Física, Universitat Autònoma de Barcelona, Bellaterra, Spain.
- ⁸ Institut de Ciència i Tecnologia Ambiental (ICTA-UAB), Universitat Autònoma de Barcelona,
- 18 Bellaterra, Spain.
- 19
- 20 Keywords: cross hole electrical resistivity tomography, fiber optic distributed temperature
- 21 sensing, alluvial aquifer, sea water intrusion, submarine groundwater discharge, Mediterranean
- 22 sea
- 23
- 24
- 25
- 26
- 27 Preprint submitted to Journal of Hydrology

28 ABSTRACT

29 The characterization of saline water intrusion (SWI) and its hydrodynamics is a key issue to 30 understand submarine groundwater discharge (SGD) and manage groundwater resources in coastal 31 areas. To test and compare different methods of characterization and monitoring, a new 32 experimental site has been constructed in a coastal alluvial aquifer north of Barcelona city 33 (Catalonia, Spain). The site is located between 30 and 90 m from the seashore and comprises 16 34 shallow piezometers organized in nests of three with depths ranging between 15 and 25 m and 4 35 solitary piezometers. The objective of this paper is to combine different recently developed 36 monitoring techniques to evaluate temporal variations in the aquifer hydrodynamics of the site at 37 different spatial scales before and after the dry season of 2015. At the site scale, fibre optic 38 distributed temperature sensing (FODTS), for the first time applied to study SWI, and cross-hole 39 electrical resistivity tomography (CHERT) has been applied. At the meter/borehole scale, electrical 40 conductivity of the formation has been applied not only in a repeated manner ("time lapse"), but 41 also for the first time at relatively high frequency (1 sample every 10 min). CHERT has provided a 42 better characterization of the seawater intrusion than electrical conductivity data obtained from 43 piezometers. The combination of techniques has allowed improving the understanding of the 44 system by: 1) characterizing the extent and shape of SWI; 2) differentiating two different dynamics 45 in the aquifer; and 3) identifying preferential flow paths over different time and spatial intervals. 46 Future challenges and the application of these techniques in other areas are also discussed.

47

48 **1. Introduction**

About 41% of the world population lives in coastal areas (Martínez et al., 2007), where groundwater
is the main source of freshwater. Intensive groundwater exploitation has caused seawater intrusion
(SWI) in the past decades, bore and soil salinization with important losses in agricultural production
globally. Moreover, the impact and magnitude of SWI is expected to be exacerbated by the increase

in the freshwater demand due to the population growth, as well as by climate change and sea-level
rise. Much progress has been made to characterize SWI incoastal aquifers and to understand the
reactions occurring in the reactive mixing zone between terrestrial groundwater and seawater (i.e.
subterranean estuary; Anwar et al., 2014; Moore, 2010, 1999; O'Connor et al., 2015; Santos et al.,
2009; Spiteri et al., 2008). However, the understanding of the dynamics of mixing and dispersion
and its impact on chemical reactions remains a challenge.

59

60 Transport processes in coastal aquifers have recently received increasing attention, particularly in 61 relation to the threat of SWI as well as the complex aquifer-ocean interactions. Density-driven 62 circulation significantly affects dispersion, mixing, and reaction behavior of nutrients and 63 contaminants transported by freshwater, and thus, the supply of these compounds to the coastal 64 sea (Brovelli et al., 2007; Dror et al., 2003). The discharge of groundwater to the sea, commonly 65 referred to as submarine groundwater discharge (SGD), has been recognized as a relevant source of dissolved nutrients (Kim et al., 2011; Rodellas et al., 2015) and metals to the ocean (e.g. (Bone et 66 67 al., 2007; Trezzi et al., 2017; Windom et al., 2006) with important implications for coastal areas. 68 Therefore, guantification of fluxes between coastal aguifers and oceans is critically important, both 69 from a hydrogeological and oceanographic perspective. However, while studying the same process 70 and addressing the same key questions, these two communities have evolved independently.

71

Fluctuations on the fresh-saltwater interface location and width depend on many factors (e.g. recent and past freshwater inputs, permeability of the aquifer, sea-level, and tidal range). The complexity of coastal systems under natural dynamics is aggravated by global change (e.g. sea-level rise, changes in precipitation, increased urbanization in coastal areas, increase on the demand of hydrological resources, etc.) (Michael et al., 2017; Rufí-Salís et al., 2019). The predicted climate change with its associated sea-level rise will modify flow regimes and groundwater discharge 78 conditions in many coastal areas. All these changes and interactions will affect the inland extension 79 and behaviour of SWI as well as the chemical composition and magnitude of SGD. Additionally, 80 despite the transient behaviour of SWI, its current understanding is mainly based on studies that 81 assume steady-state conditions (Werner et al., 2013), and thus implicitly neglect the transient 82 effects and processes that are occurring at different temporal and spatial scales, such as quality 83 patterns (mixing, cation exchange and/or precipitation/dissolution of different minerals). 84 Furthermore, the behaviour of SWI differs between passive and active SWI (Badaruddin et al., 2015; 85 Werner, 2017), and rain events can also have an important influence in the dynamics of these 86 systems at day time-scale (CerdàDomènech et al., 2017; Giambastiani et al., 2017).

87 The dynamics of the SWI can also be affected by other factors offshore, such as beach 88 geomorphology, tidal regime, wave action and the flux of fresh groundwater that discharges 89 offshore (Michael et al., 2005). Moreover, the structure of the aquifer plays an important role in 90 specific contexts, where the presence of low permeable layer as thin as a few decimeters may form 91 a confining layer dividing the system in two aquifers and modifying the dynamics of the system 92 (Pauw et al., 2017). This kind of lithological structure with a low permeable layer changes the 93 behaviour of the SWI and modifies the way in which groundwater discharges into the sea. 94 Furthermore, layered aquifer systems may exhibit enhanced seasonal exchange due to an increase 95 in the length of the fresh-saltwater interface (Michael et al., 2005). Therefore, due to the importance of lithological variability on groundwater flow and salinity patterns, detailed field studies are needed 96 97 to understand these interactions (Pauw et al., 2017).

98 Traditional methods to characterize inland aquifers are also used to depict groundwater systems in 99 coastal areas. Direct information is obtained from piezometers where hydraulic head, 100 physicochemical parameters and isotopes are measured/sampled to study the SWI and coastal 101 groundwater systems (Re et al., 2014 among others). However, as was stated already by Post in 102 2005, in order to successfully apply existing and future models that describe three-dimensional flow, transport and geochemical interaction under variable density conditions, a detailed and accurate
characterization of the subsurface is required. In recent years, there have been many new
techniques focused on this. characterization, which are summarized below.

106 Due to the correlation between water salinity and bulk or formation electrical conductivity, 107 electrical methods such as electromagnetic methods and electrical resistivity tomography (ERT) 108 represent are interesting tools to monitor SWI. Induction logging can also be used to effectively 109 detect and monitor the saltwater wedge at the meter scale and in the near field around boreholes 110 (Denchik et al., 2014; Garing et al., 2013; Pezard et al., 2015). ERT is widely used for hydrological purposes, such as monitoring contaminant plume migration or estimation of aquifer parameters 111 112 (Camporese et al., 2011; Cassiani et al., 2006; Koestel et al., 2008; Müller et al., 2010; Nguyen et al., 2009; Perri et al., 2012; Singha et al., 2015 among others). Some examples of its successful 113 114 application to image the freshwater-saltwater interface in different coastal areas are described by de Franco et al., 2009; Goebel et al., 2017; Morrow et al., 2010; Nguyen et al., 2009; and Zarroca et 115 116 al., 2014. Nevertheless, although widely used, the sensitivity of the ERT measurements depends on 117 the acquisition methodology. One of the most widespread method consists on positioning the 118 electrodes on the surface. The resolution under this configuration decreases as the acquisition 119 depth increase, preventing the quantitative use of the data from a hydrogeological perspective. An 120 alternative method consists of installing the electrodes along the piezometers, which contributes 121 having better resolution acquisitions (Perri et al., 2012). Attempts to link bulk electrical conductivity 122 obtained using surface or surface-toborehole ERT are found in the literature (Beaujean et al., 2017; 123 Huizer et al., 2017; Nguyen et al., 2009). These studies concluded that ERTderived water salinity is 124 usually underestimated. Therefore, attaching electrodes around and along the piezometers, allows 125 to acquire data near the area of interest and cross-hole, limiting the loss of resolution in depth. 126 Temperature can also be used as a tracer of environmental processes in groundwater using the

127 temperature contrast between two endmembers. Fiber optic distributed temperature sensing (FO-

128 DTS) has already proven to be a useful cost-effective tool to perform detailed monitoring of 129 environmental processes (Selker et al., 2006; Tyler et al., 2009). At present, FO-DTS instruments 130 provide temperature resolution of 0.01 °C, a spatial sampling of 0.25 m along the cable and a 131 temporal resolution of fractions of a minute depending on the configuration chosen (Selker et al., 132 2006, Simon et al. 2020). The application of this technology in groundwater environments, has been 133 used in both fractured media and unconsolidated aquifers to determine river-aquifer interaction 134 (Briggs et al. 2016; Rosenberry et al.2016), evaluate groundwater preferential paths, identify 135 fracture connectivity, and approximate aquifer hydraulic and thermal properties (Bakker et al., 136 2015; Bense et al. 2016; Hausner et al., 2016; Klepikova et al., 2014). In coastal aquifers, 137 temperature may be a good indicator for mixing due to the different temperatures of fresh 138 groundwater and seawater. Based on this contrast, Taniguchi (2000) used temperature as a proxy 139 to monitor dynamics in the fresh-saltwater interface at a coastal aquifer in Japan. The same principle 140 was used by Debnath et al., (2015) and Henderson et al., (2008) to monitor interactions between 141 groundwater and sea water using temperature probes and distributed temperature sensing (DTS) 142 respectively. Based on these natural differences in temperature, FO-DTS may be used as a passive 143 sensor to monitor the spatial distribution and temporal fluctuations of the fresh-salt water interface 144 with high-definition. Despite some examples that combine marine ERT and FO-DTS to monitor tidal 145 pumping and SGD (Henderson et al. 2010), FO-DTS has not been applied yet to characterize SWI. 146 Given the importance of fresh groundwater preservation and the complexity of coastal aquifer 147 settings, it is doubtful that a single technique is enough to understand the whole system. For this 148 reason, we have developed a research site in a Mediterranean alluvial aquifer (North of Barcelona, 149 Catalunya, Spain), between 40 and 90 m from the seashore, to compare the performance of 150 different characterization methods and provide new insights to be shared among the SWI and SGD 151 scientific communities. Unlike most studies conducted elsewhere, the selected area presents a

microtidal regime, which allows focusing on other physical driving forces (waves, storms,groundwater table elevation, etc).

154 In this paper we present the preliminary results of the jointly application of different novel 155 monitoring techniques with the objective to evaluate temporal variations of a coastal aquifer at 156 different spatial scales. At the site scale, the selected techniques are: 1) Cross-hole electrical 157 resistivity tomography (CHERT) and 2) FO-DTS, as a first attempt to apply this technique to 158 characterize SWI. At meter/borehole scale, for the first time according to the authors knowledge, 159 we deployed a time lapse induction logging (TILL) with an electromagnetic probe to obtain 160 formation conductivity logs at relatively high frequency (one sample every ten minutes) to study 161 aquifer processes at these time scales. These techniques are compared with traditional measures in 162 piezometers (electrical conductivity and temperature) and applied to obtain two snapshots at 163 different times to characterize the beginning and the end of the dry season (June and September 164 2015).

165

166 **2. Experimental site**

The experimental area is located at the Mediterranean coast close to the mouth of the ephemeral stream Riera d'Argentona, 40 km to the NE of Barcelona (Catalunya, Spain), between the urban areas of Mataró and Vilassar de Mar (Fig. 1). The area is subjected to a Mediterranean semiarid climate with an average annual precipitation of 610 mm (period 2010–2013). The land use is manly divided into agriculture and urban and natural forest (Rufí-Salís et al., 2019). The watercourse is characterized by torrential and non-permanent water flow, which only develops after heavy rain events.

174 The site is located in the lower part of the Argentona stream alluvial aquifer. The geology of the 175 experimental site is dominated by layers of quaternary gravels, sands and clays which are the product of the weathering of the surrounding granitic outcrops (Fig. 2) (MartínezPerez et al., 2018,
Internal communication).

178 In a distance of 30 to 90 m from the seashore, 16 shallow piezometers were installed in a small area 179 of 30 by 20 m, 30 m inland from the seashore. Most piezometers are gathered in nests (N1, N2, N3 180 and N4) of three (15, 20, 25), with depths of around 15, 20 and 25 m (e.g. nest N1 is composed of 181 piezometers N115, N120 an N125). Four standalone piezometers were also installed (PS25, PP15, 182 PP18, and PP20, with depths of 25, 15, 18, and 20 m, respectively). All piezometers were equipped 183 with at least 2 m of blind pipe at the bottom, followed by 2 m of screened tube just above. The only 184 exceptions were piezometers PP15 and PP20, which were screened 11 and 15 m respectively. The 185 study presented here focuses on a cross-section perpendicular to the sea following the piezometers 186 N225, N325, N125, PP15 and PP20 (Fig. 2).

187

188 3. Methodology

189 The methodology described below describes the two stages of the study: (1) initial downhole set-190 up of both FO-DTS and the distribution of electrodes to perform CHERT, and (2) field surveys of FO-

191 DTS, CHERT and time-lapse induction logging.

192

193 *3.1. Installation of annulus fiber optic cable and electrodes during site construction*

194

The fiber optic cable (Brugg Kabel AG, Switzerland) and CHERT electrodes lines were placed during piezometers installation in the annular space between the PVC casing and the formation. The procedure consisted of three steps: (1) electrodes assembly; (2) installing the electrodes along the tubing; and (3) fiber optic cable set-up along the piezometer casing (Fig. 3).

199 The assembly of the electrodes aimed at hindering corrosion, since this is one of the main difficulty

200 with semi-permanent electrodes in a saltwater environment. The electrodes, made of stainless steel

201 meshes, were tested in the laboratory by submerging them (and the cables) in a salty solution (55 202 mS/cm). Then, the electrodes were connected to an electrical cable through which an electrical 203 current was imposed to observe the corrosion. The most sensitive part of the system was the 204 connection between the mesh and the cable. After several tests, we determined that the best 205 strategy to delay corrosion was to fully cover the connection with silicone to minimize the contact 206 with saltwater. The final prototype showed corrosion signs in the laboratory after 20 days of 207 continuous current of 1 A at 3 Hz frequency. The current injected during field experiment is much 208 smaller than 1A and time of injection is only a few mS. This, together with the reducing conditions 209 fonded at the deepest part of the site (Martínez-Perez et al., 2018, Internal communication), should 210 ensure reliable operation for the foreseeable duration of the project. More details in the CHERT configuration can be found in Palacios et al. 2020. 211

212 The installation stage consisted of fixing electrodes and fiber optic lines to the casing by means of 213 nylon flanges, which effectively acted as centralizers to protect the two lines (Fig. 3). Had the wells 214 been deeper, more sturdy protection and centering system would have been needed. The first 215 casing sections were instrumented outside the well and placed vertically inside the auxiliary drilling 216 tubing with the help of the rig crane. The rest of the casing sections were instrumented just after 217 they had been screwed to the casing string already in place. The casing, thus instrumented, was 218 lowered smoothly into the auxiliary tubing for all piezometers. Special attention was paid to prevent 219 dragging casing and instrumentation lines during the extraction of the auxiliary tubing. To this end, 220 we tried to minimize the length over which filter sand in the screened interval, or clay pellets 221 elsewhere, overlapped with auxiliary tubing (Fig. 3). The whole operation requires the pro-active 222 collaboration of drillers, who were carefully trained on what we were trying to do.

The deepest piezometers of each nest, and the stand-alone piezometers (PP15 and PP20) were equipped with 36 electrodes (Fig. 2) to perform both vertical and cross-hole electrical resistivity tomography (CHERT). Distances between electrodes were 40 cm, 50 cm and 68 cm for piezometers with depths of 15 m, 20 m and 25 m, respectively. In the Argentona site, the distance between nests
and pumping wells varies from around 10 m to 26 m while the distance between PP20 and P15, the
shallowest piezometers equipped with CHERT, is of 12.7 m. In the line perpendicular to the coastline,
the CHERT has an aspect ratio (horizontal distance between boreholes divided by depth of
boreholes) between 0.6 and 0.8.

231 In order to perform FO-DTS, fiber optic cables must be installed in all the piezometers of the site. 232 This made it difficult because the cable needed to be cut to extract the auxiliary tubing (Fig. 3). The 233 end of the cable was passed through the tubing to minimize the number of cable segments (or 234 fusions) connections, which cause a loss in signal. In hindsight, this has proven a severe hindrance. 235 The connections between fusions were done with a Prolite-40 Fusion Splicer (PROMAX, Spain) and 236 an EFC-22 fiber optic cutter (Ericson, Sweden). Two continuous lines of fiber optic cable were set up 237 along the site period. Line 1 included wells from nest N1 and N3, and stand-alone wells PP15 and 238 PP20 period. Line 2 included wells PS25, N220, N215, N225. The total length of fiber optic cable 239 installed was of 1,900 m approximately, with 17 connections.

240

241 3.2. Data acquisition

All surveys were performed in 2015. CHERT data were acquired on July 3rd and September 8th in boreholes N125, N225, N325, PP20 and PP15. Temperature with FO-DTS was measured on June 25th – 26th and September 10th, in all piezometers, for durations spanning between one and four hours. Time-lapse induction logging (TLIL) were recorded only in borehole N3-20 on May 11th and 12th. During all field surveys, head levels, groundwater temperature and electrical conductivity were measured in all piezometers including water EC vertical profiles in PP20 using a CTD-Diver Schlumberger.

249

250

3.2.1. Cross hole electrical resistivity tomography (CHERT)

251 CHERT was done in four pairs of piezometers to obtain a cross-section perpendicular to the 252 coastline, from PP20 to N225. The acquisitions were made using 72 electrodes in total. To maximize 253 resolution, the cross-hole configurations used were dipole–dipole, pole-tripole, and Wenner, 254 following Bellmunt et al. (2016) (Fig. 4). We used a ten-channel Syscal Pro resistivity meter and an 255 optimized survey design which allowed to record normal and reciprocal measurements: a total of 256 5842 data points per cross-hole panel in 30 min. The recording of a complete CHERT took 2 the 257 conjugate gradients method. The regularization includes a geostatistical operator that helped 258 removing borehole footprints from the images, caused by the high sensitivity of the method around 259 the electrodes. More detail on the inversion is described by Palacios et al. 2020.

260

261

3.2.2 Time-lapse induction logging (TLIL)

The EM51 downhole induction probe from Geovista[®] has the capacity to perform all measurements 262 263 through PVC tubing in those cases where downhole measurements are complicated by the 264 unconsolidated nature of the sediment. The sonde was deployed the 11th and 12th of May 2015 to 265 measure formation electrical conductivity at meter-scale around piezometer N320. We choose this 266 piezometer instead of N325 because of electrode interference with induction measurements. 267 Logging was performed in a repeated or so-called "time-lapse" manner over short amounts of time, 268 with a period of 10 min between profiles. Compared with the aforementioned techniques, induction 269 logging has shorter acquisition times, allowing to characterize variations in groundwater 270 temperature and/or salinity at short time-scales. Only upward profiles were used in this study 271 despite downward profiles were also recorded.

When comparing field surveys from July and September (CHERT) and induction logging (TLIL), the contribution of the surface conductivity of the grains was suppressed by using the following expression by Waxman and Smits (1968), in which the total formation conductivity is given by the sum of the conductivity of the pore volume and the pore surface:

$$Co = \frac{Cw}{E} + Cs \tag{1}$$

277

276

278 where C_0 is the formation electrical conductivity (mS/m), C_w is the groundwater conductivity, F is 279 the dimensionless formation factor and Cs is the surface conductivity of the pore space at the 280 interface with mineral grains, typically associated to the presence of clay. The electrical formation 281 factor F is a petrophysical parameter that depends on matrix porosity and pore connectivity and 282 describes the efficiency of the fluid-filled pore-space to conduct current. As the Argentona site is 283 close to the sea, groundwater conductivity is expected to be high due to salinity, and the term Cs 284 can be considered negligible. Even if it was not the case, by computing the difference in 285 conductivities between two surveys, and assuming that C₅ is constant because the sediment remains 286 undaunted, the conductivity changes observed in the time-lapse are only related to changes in pore 287 fluid conductivity(C_W).

288

289

3.2.3 Fiber optic distributed temperature sensing (FO-DTS)

290 Temperatures were obtained from both fiber optic lines with an Ultima XT Distributed Temperature 291 Sensor (Silixa, UK). Spatial sampling is set up at 25 cm, with a spatial integration length of 0.5 m. The 292 sampling period was set to approximately 1 min, with an integration time of 10 to 30 s. Data sets 293 collected in June 2015 had an integration time of 10 s. In order to keep consistency between June 294 and September data, the June data were summed up every 30 s. The signal was calibrated using two 295 reference baths of 57 L of water placed in a portable cooler (Ice Cube, Igloo, USA). Temperature was 296 homogenised with aquarium bubblers, and monitored with RBRsolo-T temperature loggers with and 297 accuracy of 0.02 °C (RBR, Canada). One of the baths, common to both lines, was kept at 0 °C thanks 298 to a well-mixed ice and water mixture, whereas a separate bath at ambient temperature was 299 installed in each line.

Four different datasets were calibrated, one for each FO line (line 1 and line 2) and one for each field
 survey (June and September). The calibration approach was adapted to the peculiarities of each
 dataset:

(1) The two datasets bellowing to Line 1 (June and September) were calibrated with the single ended
calibration (Hausner et al., 2011) as no anomalies were found in the fiber connections or calibration
baths. (2) For Line 2 datasets, both June and September acquisitions presented different types of
anomalies, forcing to perform the temperature inversion with different methodologies for each of
them.

308 In the case of the data gathered during June 2015 for line 2, no data from the thermometer 309 monitoring the ambient temperature was available to perform the calibration. Moreover, the 310 differential attenuation between segments of the fiber optic cable connected to form the complete 311 line was different. These two issues prevented the use of a singleended calibration as in the case of 312 line 1 datasets. To solve this, we first had to correct the differential attenuation using another 313 calibration point along the fiber optic line. We choose the screened interval of the 25 m depth 314 piezometer N425. At this depth, the temperature remains constant for the monitoring period 315 (hours), and we could use the temperature data recorded by the pressure-temperature sensor 316 permanently installed at that depth. Secondly, we calculated the inversion parameters using the ice-317 bath and the new calibration point.

In the case of the data collected during the field campaign in September 2015 for line 2, the same problem related to the different differential attenuations between glass segments was solved. Moreover, only part of the reference thermometer data could be used. Thus, the temperature data was extrapolated to those acquisitions times where no temperature from the thermometer was available. This could be done since the monitoring period was small, thus changes in the bath temperature in time were negligible. So ,for this dataset double-ended calibration (van de Giesen et al., 2012) could be applied to account for the change in the differential attenuation. 325

326 4. Results

327

4.1. Cross hole electrical resistivity tomography (CHERT)

328 The bulk conductivity (C0) cross-section perpendicular to the coast obtained by CHERT (Fig. 5) shows 329 a very flat SWI wedge located 2 m below a continuous layer of silt identified at 12 m depth. However, 330 the higher spatial resolution obtained with CHERT between wells PP15 and PP20 appear to indicate 331 two different SWI areas, one in the upper part of the aquifer, close to PP20 and another one in the 332 deeper part of the aquifer. These data do not correlate with the EC measured in the piezometers. 333 Indeed, the EC measurements from fully screened piezometers (PP15 and PP20) show relatively 334 constant values of EC between 4 and 12 m depth (Fig. 6). Below this depth, there is a general 335 continuous increase in EC towards seawater conductivities that does not correlate with CHERT data. 336 Whereas EC patterns observed in PP15 and PP20 are likely to be artificially affected by the use of 337 fully screened piezometers, further research is required to appropriately understand the difference 338 on conductivity patterns derived from EC measurements in the piezometers and CHERT.

339

The higher spatial resolution obtained with CHERT allows differentiating two different mixing zones that could be indicative of two different discharging areas: a shallow one with a recirculation cell perhaps influenced by seawater infiltration from waves/storms, and a deeper one discharging away offshore. However, more data is required to verify these hypotheses, as the conductivity measured in the piezometers (Supplementary material, Table 1), shows higher values (i.e. lower resistivity) in the deepest part of the aquifer.

The ratio between the bulk electrical conductivity from June and September is a good tool to evaluate the seasonal variation in salinity (Fig. 7). Whilst, ratios below 1 (color blue) indicate a decrease of conductivity, ratios above 1 (color red) indicate an increase. The maximum changes occur between 15 and 20 m depth, with a general increase in conductivity mainly in the coarser

350 materials between the piezometers N3 to PP20. This trend is consistent with groundwater level 351 evolution (data not shown) that tend to decrease between 2 and 4 cm between June and 352 September. However, this evolution is not clear in the water electrical conductivity changes 353 measured in the piezometers (Supplementary material, Table 1). There are changes along the 354 vertical electrical conductivity profiles measured with the EC probe, most significant in June, due 355 to the presence of sediments with different grain size that cause different flow distributions along 356 the 2 m screened intervals of most piezometers. Furthermore, compared with piezometers data, 357 CHERT models shows the extension of the area affected by this increase of conductivity in the 358 sands as well as in the weathered granite.

In the upper part of the aquifer there are no significant variations of conductivity as indicated by data obtained in the piezometers PP15 and PP20 (Supplementary material, Table 1), pointing out that a relatively stable SWI between June and September at shallow depths.

362

363 4.2. Time-lapse induction logging (TLIL)

Time-lapse downhole measurements were carried out in piezometer N320 only and at the onset of the dry season (May 2015). Downhole measurements are repeated every 15 min in the same hole, in a timelapse mode, and surface conductivity (C_s) and formation factor (F) were considered constant (Eq. (1)). Consequently, changes in bulk electrical conductivity are attributed to changes in groundwater conductivity due to changes in pore fluid temperature and/or salinity with time.

We considered that changes in formation conductivity (C₀) were directly proportional to changes in water EC, and therefore C₀ measured in May (Fig. 8) correlate with groundwater EC measured a month later in the same piezometer (Supplementary materials, Table 1). The maximum C₀ values were obtained below 14 m depth, in agreement with piezometers data and the June 2015 CHERT acquisition/cross-section. In this way, groundwater at the top of the screened interval for N320 showed values in agreement with the conductivity of 13.3 mS/cm measured in this piezometer in 375 June.

376 When considering C_0 changes over short periods of time (10's of minutes), some depth intervals 377 present significant changes between May 11th and 12th, 2015 (Fig. 8a). Changes can be revealed 378 by calculating pore fluid conductivity ratios (ΔC_0) of each profile taking the May 12th, 2015 profile 379 (IL23) as reference. Close to the surface (from 3 to 9 m depth), were we can find fresher 380 groundwater, all ratios exhibit a decrease in conductivities overnight, with changes up to 30% (Fig. 381 8c, Pore fluid conductivity ratios (ΔC_0)). This decrease can be due to a decrease in groundwater 382 temperature or salinity overnight, or both in the same time period. In the same depth interval, 383 smaller changes of 5 to 10% are noticed between IL3 and IL4 (Fig. 8c, orange and green profiles) 384 that is in less than 15 min. These smaller changes appear in decimeter-thick intervals (near 5.5 385 and 6.5 m depth) and point at small changes in groundwater temperature or salinity in a very short 386 time. These high frequency conductivity changes (grey sections in Fig. 8) are attributed to the 387 presence of preferential flow pathways with relatively high fluid flow at these depths. It can be 388 noticed that the TLIL method identified small zones of preferential flow that cannot be identified 389 with CHERT.

In the brackish to sea-water saturated region below 9 m depth, very little conductivity changes were registered, in occasions less than 0.5%. These horizons correspond to finer grained materials where fluid flow is troublesome. Between 14.5 and 15.5 m depth, noticeable changes up to 5% are measured over a meter-thick interval. These tiny changes underline the precision of TLIL measurements. Below 16 m depth (Fig. 8c), no significant changes in C₀ are observed with TLIL.

395

396 4.3. Fiber optic distributed temperature sensing (FO-DTS)

Rather than presenting the results as temperature depth profiles, we choose to interpolate the data
with a simple linear interpolation perpendicular to the sea to highlight the spatial patterns (Fig.
Unlike the geophysical techniques, which provide a wider distribution of measures in the

400 subsurface, FO-DTS data concentrates on several vertical lines. Therefore, the uncertainty of the 401 interpolation is larger, and any future qualitative analysis would be better based on the un-402 processed temperature data (i.e. the vertical temperature depth profiles). However, interpolated 403 plots allow a good qualitative analysed and comparison between the different techniques 404 performed in this study. The general distribution of temperature follows the same trend as 405 piezometer measurements (Supplementary material, Table 2), with higher temperatures inland 406 and lower closer to the sea (Fig. 9). However, all temperatures measured in the piezometers are 407 higher than those measured with FO-DTS. While the maximum temperature in groundwater 408 according to the FO-DTS is below 19.40 °C, several temperature measurements in piezometers are 409 above this value. This pattern is observed in the wells with a 2 m screened interval as well as those completely screened (PP15 and PP20), where the complete cross-section tends to show 410 411 higher temperature. The higher values measured in the piezometers compared with the 412 distribution observed with the FO-DTS indicates that temperature in piezometers is significantly 413 altered by atmospheric temperature. In contrast, the fiber installed in the annular space of the 414 piezometers, measured temperatures much closer to the expected subsurface temperatures.

415 The highest influence of the atmospheric temperature is above the first 3 m depth, corresponding 416 to the non-saturated zone. In this regard, the annual thermal oscillation extinction point was 417 estimated to be less than 15 m depth applying the solution proposed by Stauffer et al., (2013) 418 with the local atmospheric temperature in the period MaySeptember 2015 (data not shown). 419 Between June and September, this influence can be reduced to 12 m depth. Considering that 420 neither the vertical nor the horizontal distribution of temperature underground is homogenous, 421 different effects can condition groundwater temperature distribution. However, the small influence 422 of atmospheric temperature on groundwater temperature at shallow depths indicate that 423 groundwater flow in this area is relatively important, which is in accordance with the coarser grain 424 of the materials found in the upper part of the aquifer.

425 In June (Fig. 9a), the temperatures are generally lower than in September, showing more spatial 426 changes in the upper part of the aquifer than at the bottom. It is clear that there are two 427 sources of temperature anomalies relating to the coldest and hottest temperatures of the cross-428 section. The coldest anomaly located between N325 and N125 is attributed to the local effect of 429 surface recharge from a discontinuous sewerage discharge channel close to the site, as also 430 observed in temperature data from the N4 nest (data not shown) which is located close to this 431 discharge area. The warmest temperature of the cross-section is observed in the inland part of 432 the experimental site (between N225 and N325) highlighting the thermal effect of groundwater flow recharged inland. The coldest temperatures of the cross-section are measured at the deepest 433 434 part of the aquifer at the bottom of a thick pack of sands below 18 m and close to the weathered 435 granite.

436 Since this is the most affected area by SWI, the coldest temperatures derived from FO-DTS data 437 are related to the intrusion of colder seawater, in contrast with the warmer temperatures 438 observed inland. In this regard, the small areas with the coolest temperatures around at depth 439 of 11 m, close to N125 and PP15, could also be indicating the presence of a shallower saltwater 440 wedge.

441 In September (Fig. 9b), temperature distribution is similar to June, but with slightly higher values. 442 The non-saturated zone, as well as the deepest part of the cross-section, shows higher 443 temperatures. As in June, the warmest temperatures are located at the innermost and shallow 444 part of the aquifer and the coldest temperatures at the deepest part of the aquifer, closer to the 445 sea. However, there is a slight increase of the coldest temperatures in the deeper part of the 446 aquifer, which could be related to the higher temperatures along the profile due to the dry season, 447 an increase of sea water temperature, or both. In the same way, when considering only the shallow 448 part of the aquifer, the coldest temperature corresponds to the part of the profile closest to the 449 sea. However, during summer there is an increase in sea water temperature in the shallow depths that does not seem to affect the temperature distribution in the upper part of the aquifer in thecloser zone to the sea.

452

453 **5. Discussion and integration of techniques**

454 The combination of techniques applied in the Argentona site allows describing the behaviour of the 455 system at the beginning and at the end of the dry season. The shallowest part of the aquifer does 456 not show important salinity changes during the studied period. On the other hand, the deepest 457 part of the aquifer shows an increase in salinity over the season, mainly observed in the bottom part of the sedimentary formation. Despite the basement showing a high electrical conductivity, 458 459 the values measured remain constant during the studied period (i.e. low flow). This lack of dynamism is in agreement with the low transmissivity obtained through short pumping tests 460 461 (Martínez-Perez et al., 2018, Internal communication). At borehole scale, induction logging 462 revealed the presence of preferential flow paths at different depths.

463 Whereas each applied technique provides partial information of the coastal aquifer, the 464 combination of techniques allows obtaining a comprehensive understanding of the characteristics 465 and hydrodynamics of this complex system. CHERT and FO-DTS provide important information 466 at the site scale, whereas TLIL characterizes the system at the meter-scale. With CHERT and FO-DTS 467 data we can differentiate two behaviours in the aquifer. While CHERT identified the main active 468 area of SWI intrusion occurring in the deepest part, FO-DTS does not show important changes 469 between both surveys. However, FO-DTS and TLIL highlighted that the shallowest part of the 470 aquifer is an active system with fresh groundwater flow occurring; a statement that cannot be 471 made by looking only at CHERT.

The active area identified with TLIL in the deepest part of the aquifer at 15 m depth (Fig. 8),
corresponds to the upper part of the active area identified with CHERT (Fig. 7). On the contrary,
no significant changes are observed with TLIL below 16 m unlike what was found by CHERT. These

differences between both methods could be related to the fact that both are deployed at different spatial and temporal scales. In this way, CHERT could point more to seasonal changes, while TLIL may be related to more instantaneous changes. Whilst we cannot evaluate the capacity of CHERT to identify hourly or daily changes with the data collected, we infer that it would be possible to capture such changes with a proper experimental design that allows to

acquire data with a high temporal frequency. More research is needed to understand this issue,
increasing the number of points were C_o is measured and/or by increasing the temporal
frequency of CHERT profiles.

483 Despite the high density of piezometers in the field site, and the screening of the piezometer on 484 only 2 m, these techniques provide higher spatial resolution than direct measurements. Furthermore, they can give more representative information of the SWI extent, especially CHERT. 485 486 In this regard, CHERT data correlates relatively well with water electrical conductivity of most 487 piezometers in the different locations and depths along the study site. However, CHERT provides 488 a 2D representation of the shape and extent of the seawater intrusion and its seasonal variations. 489 This information would be impossible to obtain using only point measurements from piezometers. 490 This is particularly relevant in fully screened piezometers.

491 Temperature measured with FO-DTS in the annular space of the boreholes (i.e., closer to the 492 aquifer matrix) is lower than temperature measured in the piezometers, pointing out the 493 influence of atmospheric/soil conditions on groundwater measurements from piezometers.

The changes in formation electrical conductivity measured with TLIL may indicate preferential flows at a smaller scale than the other techniques, giving information that cannot be obtained and/or approximated with traditional monitoring methods. Only tracer tests in the screened intervals could generate similar data but with lower spatial and temporal resolution, and at higher costs.

499 The study site is located in the Mediterranean basin and therefore subjected to a microtidal regime.

500 In other oceanic coastal areas, the effect of tides is more significant and can influence the dynamics 501 of the system. It is expected that in open sea areas the applied methods could improve the 502 characterization of the system. That is particularly relevant for the FO-DTS, as higher tides increase 503 the dynamism of the SWI, increasing the thermal influence of the sea boundary condition on the 504 aquifer. This assumption could explain why the influence of the sea has been found to be minor in 505 this study, despite various studies indicate that temperature can be used as a useful SWI tracer. In 506 the same way, the application of this technique in those areas with important thermal contrast 507 between sea and groundwater temperature could give better results. Finally, the connection to 508 the sea and the thermal properties of the geological materials could limit the application of this 509 technique.

510

511 **6. Conclusions and future challenges**

512 Different approaches and techniques (direct groundwater measurements from piezometers, 513 CHERT, FO-DTS and TLIL) have been combined for the first time to study a 25-m thick microtidal 514 coastal aquifer during the dry season (before and after summer 2015). CHERT profiles allow a better 515 definition of the shape and distribution of the seawater intrusion, as well as its seasonal changes, 516 than data obtained from point groundwater measurements from piezometers. In this case study, 517 the combination of the different techniques has allowed improving the understanding of the 518 hydrogeological system by: 1) A proper characterization of the extend and shape of the SWI, 2) 519 differentiating two different zones with different dynamics in the deep and upper part of the 520 aquifer and 3) identifying preferential flow paths over different time and spatial intervals. The 521 distribution of the SWI does not follow the typical shape, with main changes between 15 and 18 522 m depth. Despite minor changes in salinity measured in the shallower part of the aquifer, data 523 provided by TLIL and FO-DTS indicate that it is an active system. Although precise characterization 524 of the aquifer was achieved by combining different geophysical techniques, the groundwater

525 discharge process to the coastal sea (i.e. SGD) is still a challenge. Considering the information 526 obtained from the techniques applied, there are two different mixing zones that could be related 527 to two different discharge areas: a shallow recirculation cell closer to the sea, mainly influenced 528 by wave setup and storm effects (in addition to the terrestrial hydraulic gradient), and a deep 529 discharge area, acting at a more seasonal scale, and likely discharging offshore. However, the 530 extension of the discharge of the deep aquifer into the sea is not fully clear. More data are required 531 to fill the blank between the site and the sea, but also inland, to improve the understanding of the 532 system. Yet, a higher temporal and spatial resolution of the already applied techniques will also 533 improve the understanding of the system considering the following:

534

Higher temporal resolution of CHERT would allow understanding how mixing is occurring
and the origin of salinity in the shallow part of the aquifer (convective zones, wave effect,
etc.). At the same time, higher temporal resolution would allow understanding why at the
end of the dry season there is a decrease of salinity simultaneously with an increase of
salinity in the deeper part of the aquifer.

TLIL could be applied in more piezometers at the same time to study the changes in
 conductivity and check whether these changes correspond to preferential flows at
 decimeter scale. In the same way, using this technique with higher frequency could allow
 understanding if these potential preferential flow paths respond to to the heterogeneity of
 the system and/or the recharges/discharges processes occurring at different temporal
 scales (storms events versus seasonal dynamics).

FO-DTS has allowed obtaining more information in areas with no TLIL data and/or where
 the conductivity changes are not significant to obtain representative data with CHERT.
 Nevertheless, only minor differences between both surveys are measured. Therefore, the
 potential use of temperature as a tracer using FO-DTS needs to be evaluated for a longer

period of time as its distribution is significantly affected by the thermal characteristics of the boundary conditions (atmosphere, recharge inland, sea, etc.) and therefore changing along seasons. In this way, it is important to consider that the boundary conditions tend to change in a similar way along seasons but with some lag and different extreme values.

555

556 Although more research is needed, the application of the presented techniques in a well-557 characterized study area such as the Argentona site has allowed describing the effectiveness of FO-558 DTS, CHERT and TLIL to characterize coastal areas dynamics. This information has pointed out the 559 potential of these techniques to be applied in other areas. In this way, the best technique to use 560 when characterizing coastal aquifers dynamics will depend on the temporal and spatial resolution 561 required. The importance of fresh water flow in the system can also indicate which methods 562 should be combined and the amount of data that is required. The structure of the aquifer 563 (unconfined/confined vs multilayer aquifer), and the boundary conditions (recharge patterns, 564 thermal contrast between boundaries, etc.) can also condition the combination of techniques to 565 be used. Lastly, studying zones in small basins as the Mediterranean or on the contrary in open 566 ocean conditions, will also influence the approach to apply due to the different dynamics of costal 567 aquifer in both areas. In all cases, the electrical conductivity and temperature data obtained with 568 the CHERT, FO-DTS and TLIL is expected to be more representative than the same data obtained 569 in the piezometers.

570

571

572 Acknowledgements

573 This work was funded by the projects CGL2013-48869-C2-1-R/2-R and CGL2016-77122-C2-1-R/2-574 R of the Spanish Government. We would like to thank SIMMAR (Serveis Integrals de

575 Manteniment del Maresme) and the Consell Comarcal del Maresme in the construction of the 576 research site. The authors want to thank the support of the Generalitat de Catalunya to MERS (2018 SGR-1588). This work is contributing to the ICTA 'Unit of Excellence' (MinECo, 577 578 MDM20150552). Part of the funding was provided by the French network of hydrogeological 579 observatories H+ (hplus/ore/fr/en) and the ANR project EQUIPEX CRITEX (grant ANR-11-EQPX-580 0011). V Rodellas acknowledges financial support from the Beatriu de Pinós postdoctoral program 581 of the Generalitat de Catalunya (2017-BP-00334). M. Diego-Feliu acknowledges the economic 582 support from the FI-2017 fellowships of the Generalitat de Catalunya autonomous government 583 (2017FI_B_00365). This project also received funding from the European Union's Horizon 584 2020 research and innovation programme under the Marie Sklodowska-Curie Grant Agreement No 722028. 585

586

587 7. References

- Anwar, N., Robinson, C., Barry, D.A., 2014. Influence of tides and waves on the fate of
 nutrients in a nearshore aquifer: Numerical simulations. Adv. Water Resour. 73, 203–
 213. https://doi.org/10.1016/J.ADVWATRES.2014.08.015.
- Badaruddin, S., Werner, A.D., Morgan, L.K., 2015. Water table salinization due to seawater
 intrusion. Water Resour. Res. 51, 8397–8408. https://doi.org/10.1002/
 2015WR017098.
- Bakker, M., Calj, R., Schaars, F., van der Made, K.-J., de Haas, S., 2015. An active heat tracer
 experiment to determine groundwater velocities using fiber optic cables installed
 with direct push equipment. Water Resour. Res. 51, 2760–2772. https://doi.
 org/10.1002/2014WR016632.
- 598 Beaujean, J., Nguyen, F., Kemna, A., Antonsson, A., Engesgaard, P., 2017. Calibration of 599 seawater intrusion models: Inverse parameter estimation using surface electrical

- resistivity tomography and borehole data. Water Resour. Res. 50, 6828–6849.
 https://doi.org/10.1002/2013WR014020.
- Bellmunt, F., Marcuello, A., Ledo, J., Queralt, P., 2016. Capability of cross-hole electrical
 configurations for monitoring rapid plume migration experiments. J. Appl. Geophys.
 124, 73–82. https://doi.org/10.1016/J.JAPPGEO.2015.11.010.
- Bense, V.F., Read, T., Bour, O., Le Borgne, T., Coleman, T., Krause, S., Chalari, A., Mondanos,
 M., Ciocca, F., Selker, J.S., 2016. Distributed Temperature Sensing as a downhole
 tool in hydrogeology. Water Resour. Res. 764 (52), 9259–9273. https://
 doi.org/10.1002/2016WR018869.
- Bone, S.E., Charette, M.A., Lamborg, C.H., Gonneea, M.E., 2007. Has submarine
 groundwater discharge been overlooked as a source of mercury to coastal waters?
 Environ. Sci. Technol. 41, 3090–3095. https://doi.org/10.1021/es0622453.
- Briggs, M.A., Buckley, S.F., Bagtzoglou, A.C., Werkerma, D.D., Lane, J.W., 2016. Actively heated
 high-resolution fiber-optic-distributed-temperature sensing to quantify streambed flow
 dynamics in zones of strong groundwater upwelling. Water Resour. Res. 52, 5179–
 5194. https://doi.org/10.1002/2015WR018219.
- Brovelli, A., Mao, X., Barry, D.A., 2007. Numerical modeling of tidal influence on
 densitydependent contaminant transport. Water Resour. Res. 43, W10426.
 https://doi.org/10.1029/2006WR005173.
- Camporese, M., Cassiani, G., Deiana, R., Salandin, P., 2011. Assessment of local hydraulic
 properties from electrical resistivity tomography monitoring of a three-dimensional
 synthetic tracer test experiment. Water Resour. Res. 47. https://doi.org/10.1029/
 2011WR010528.
- Cassiani, G., Bruno, V., Villa, A., Fusi, N., Binley, A.M., 2006. A saline trace test monitored via
 time-lapse surface electrical resistivity tomography. J. Appl. Geophys. 59, 244–259.

625

https://doi.org/10.1016/J.JAPPGEO.2005.10.007.

- Cerdà-Domènech, M., Rodellas, V., Folch, A., Garcia-Orellana, J., 2017. Constraining the
 temporal variations of Ra isotopes and Rn in the groundwater end-member:
 Implications for derived SGD estimates. Sci. Total Environ. 595, 849–857. https://
 doi.org/10.1016/j.scitotenv.2017.03.005.
- de Franco, R., Biella, G., Tosi, L., Teatini, P., Lozej, A., Chiozzotto, B., Giada, M., Rizzetto, F.,
 Claude, C., Mayer, A., Bassan, V., Gasparetto-Stori, G., 2009. Monitoring the
 saltwater intrusion by time lapse electrical resistivity tomography: The Chioggia test
 site (Venice Lagoon, Italy). J. Appl. Geophys. 69, 117–130. https://doi.org/10.1016/
 J.JAPPGEO.2009.08.004.
- Debnath, P., Mukherjee, A., Singh, H.K., Mondal, S., 2015. Delineating seasonal porewater
 displacement on a tidal flat in the Bay of Bengal by thermal signature: Implications
 for submarine groundwater discharge. J. Hydrol. 529, 1185–1197.
 https://doi.org/10.1016/j.jhydrol.2015.09.029.
- Denchik, N., Pezard, P.A., Neyens, D., Lofi, J., Gal, F., Girard, J.-F., Levannier, A., 2014. Nearsurface CO₂ leak detection monitoring from downhole electrical resistivity at the
 CO₂ Field Laboratory, Svelvik Ridge (Norway). Int. J. Greenh. Gas Control 28, 275–
 282. https://doi.org/10.1016/J.IJGGC.2014.06.033.
- Dror, I., Amitay, T., Yaron, B., Berkowitz, B., 2003. Salt-pump mechanism for contaminant
 intrusion into coastal aquifers. Science 80, 950. https://doi.org/10.1126/
 science.1080075.
- Garing, C., Luquot, L., Pezard, P.A., Gouze, P., 2013. Geochemical investigations of saltwater
 intrusion into the coastal carbonate aquifer of Mallorca. Spain. Appl. Geochem. 39,
 1–10. https://doi.org/10.1016/J.APGEOCHEM.2013.09.011.
- Giambastiani, B.M.S., Colombani, N., Greggio, N., Antonellini, M., Mastrocicco, M., 2017.

- 650 Coastal aquifer response to extreme storm events in Emilia-Romagna Italy. Hydrol.
 651 Process. 31, 1613–1621. https://doi.org/10.1002/hyp.11130.
- Goebel, M., Pidlisecky, A., Knight, R., 2017. Resistivity imaging reveals complex pattern of
 saltwater intrusion along Monterey coast. J. Hydrol. 551, 746–755. https://doi.
 org/10.1016/j.jhydrol.2017.02.037.
- Günther, T., Rücker, C., Spitzer, K., 2006. Three-dimensional modelling and inversion of dc
 resistivity data incorporating topography II Inversion. Geophys. J. Int. 166, 506–517.
 https://doi.org/10.1111/j.1365-246X.2006.03011.x.
- Hausner, M.B., Suárez, F., Glander, K.E., Giesen, N.V.D., Selker, J.S., Tyler, S.W., 2011.
 Calibrating single-ended fiber-optic Raman spectra distributed temperature sensing
 data. Sensors 11, 10859–10879. https://doi.org/10.3390/s111110859.
- Hausner, M.B., Kryder, L., Klenke, J., Reinke, R., Tyler, S.W., 2016. Interpreting Variations in
 Groundwater Flows from Repeated Distributed Thermal Perturbation Tests.
 Groundwater 54, 559–568. https://doi.org/10.1111/gwat.12393.
- Henderson, R.D., Day-Lewis, F.D., Lane, J.W., Harvey, C.F., Lanbo, L., 2008. Characterizing
 Submarine Ground-Water Discharge Using Fiber-Optic Distributed Temperature
 Sensing and Marine Electrical Resistivity. Sageep 1–11. https://doi.org/
 10.4133/1.2963319.
- Henderson, R.D., Day-Lewis, F.D., Abarca, E., Harvey, C.F., Karam, H.N., Liu, L., Lane Jr., J.W.,
 2010. Marine electrical resistivity imaging of submarine groundwater discharge:
 sensitivity analysis and application in Waquoit Bay, Massachusetts, USA. Hydrogeology
 J. 18, 173–185. https://doi.org/10.1007/s10040-009-0498-z.
- Huizer, S., Karaoulis, M.C., Oude Essink, G.H.P., Bierkens, M.F.P., 2017. Monitoring and
 simulation of salinity changes in response to tide and storm surges in a sandy coastal
 aquifer system. Water Resour. Res. 53, 6487–6509. https://doi.org/10.1002/

675 2016WR020339.

- Kim, G., Kim, J.-S., Hwang, D.-W., 2011. Submarine groundwater discharge from oceanic
 islands standing in oligotrophic oceans: Implications for global biological production
 and organic carbon fluxes. Limnol. Oceanogr. 56, 673–682. https://doi.org/10.4319/
 lo.2011.56.2.0673.
- Klepikova, M.V., Le Borgne, T., Bour, O., Gallagher, K., Hochreutener, R., Lavenant, N., 2014.
 Passive temperature tomography experiments to characterize transmissivity and
 connectivity of preferential flow paths in fractured media. J. Hydrol. 512, 549–562.
 https://doi.org/10.1016/j.jhydrol.2014.03.018.
- Koestel, J., Kemna, A., Javaux, M., Binley, A., Vereecken, H., 2008. Quantitative imaging of
 solute transport in an unsaturated and undisturbed soil monolith with 3-D ERT and
 TDR. Water Resour. Res. 44, W12411. https://doi.org/10.1029/2007WR006755.
- Martínez, M.L., Intralawan, A., Vázquez, G., Pérez-Maqueo, O., Sutton, P., Landgrave, R.,
 2007. The coasts of our world: Ecological, economic and social importance. Ecol.
 Econ. 63, 254–272. https://doi.org/10.1016/j.ecolecon.2006.10.022.
- L. Martínez-Perez M.A. Marazuela L. Luquot A. Folch L. del Val T. Goyetche M. DiegoFeliu
 N. Ferrer V. Rodellas F. Bellmunt J. Ledo M. Pool J. Garcia-Orellana P. Pezard M.
 Saaltink E. Vazquez-Suñe J. Carrera Integrated methodology to characterize hydrogeochemical properties in an alluvial coastal aquifer affected by seawater intrusion
 (SWI) and submarine groundwater discharge (SGD). 25th Saltwater Intrusion Meeting
 2018 Gdansk, Poland.
- Michael, H.A., Mulligan, A.E., Harvey, C.F., 2005. Seasonal oscillations in water exchange
 between aquifers and the coastal ocean. Nature 436, 1145–1148. https://doi.org/10.
 1038/nature03935.

699 Michael, H.A., Post, V.E.A., Wilson, A.M., Werner, A.D., 2017. Science, society, and the

coastal groundwater squeeze. Water Resour. Res. 53, 2610–2617. https://doi.org/10.
1002/2017WR020851.

Moore, W.S., 1999. The subterranean estuary: a reaction zone of ground water and sea water. Mar. Chem. 65, 111–125. https://doi.org/10.1016/S0304-4203(99)00014-6. Moore, W.S., 2010. The Effect of Submarine Groundwater Discharge on the Ocean. Ann. Rev. Mar. Sci. 2, 59–88. https://doi.org/10.1146/annurev-marine-120308-

706 081019.

- Morrow, F.J., Ingham, M.R., McConchie, J.A., 2010. Monitoring of tidal influences on the saline
 interface using resistivity traversing and cross-borehole resistivity tomography. J.
 Hydrol. 389, 69–77. https://doi.org/10.1016/J.JHYDROL.2010.05.022.
- Müller, K., Vanderborght, J., Englert, A., Kemna, A., Huisman, J.A., Rings, J., Vereecken, H.,
 2010. Imaging and characterization of solute transport during two tracer tests in a
 shallow aquifer using electrical resistivity tomography and multilevel groundwater
 samplers. Water Resour. Res. 46, W03502. https://doi.org/10.1029/ 2008WR007595.
- Nguyen, F., Kemna, A., Antonsson, A., Engesgaard, P., Kuras, O., Ogilvy, R., Gisbert, J.,
 Jorreto, S., Pulido-Bosch, A., 2009. Characterization of seawater intrusion using 2D
 electrical imaging. Near Surf. Geophys. 7, 377–390. https://doi.org/10.3997/18730604.2009025.
- O'Connor, A.E., Luek, J.L., McIntosh, H., Beck, A.J., 2015. Geochemistry of redox-sensitive
 trace elements in a shallow subterranean estuary. Mar. Chem. 172, 70–81.
 https://doi.org/10.1016/J.MARCHEM.2015.03.001.
- A. Palacios J.J. Ledo N. Linde L. Luquot F. Bellmunt A. Folch A. Marcuello P. Queralt P.A.
 Pezard L. Martínez D. Bosch J. Carrera Time-lapse cross-hole electrical resistivity
 tomography (CHERT) for monitoring seawater intrusion dynamics in a Mediterranean
 aquifer 2020 Earth Syst. Sci. Discuss Hydrol 10.5194/hess-2019-408, accepted.

- Pauw, P.S., Groen, J., Groen, M.M.A., van der Made, K.J., Stuyfzand, P.J., Post, V.E.A., 2017.
 Groundwater salinity patterns along the coast of the Western Netherlands and the
 application of cone penetration tests. J. Hydrol. 551, 756–767. https://doi.org/
 10.1016/j.jhydrol.2017.04.021.
- Perri, M.T., Cassiani, G., Gervasio, I., Deiana, R., Binley, A., 2012. A saline tracer test
 monitored via both surface and cross-borehole electrical resistivity tomography:
 Comparison of time-lapse results. J. Appl. Geophys. 79, 6–16. https://doi.org/10.
 1016/j.jappgeo.2011.12.011.
- Pezard, P.A., Abdoulghafour, H., Denchik, N., Perroud, H., Lofi, J., Brondolo, F., Henry, G.,
 Neyens, D., 2015. On Baseline Determination and Gas Saturation Derivation from
 Downhole Electrical Monitoring of Shallow Biogenic Gas Production. Energy
 Procedia 76, 555–564. https://doi.org/10.1016/j.egypro.2015.07.910.
- Re, V., Sacchi, E., Mas-Pla, J., Menció, A., El Amrani, N., 2014. Identifying the effects of
 human pressure on groundwater quality to support water management strategies
 in coastal regions: A multi-tracer and statistical approach (Bou-Areg region,
 Morocco). Sci. Total Environ. 500–501, 211–223.
 https://doi.org/10.1016/j.scitotenv.2014.08.115.
- 742 Rodellas, V., Garcia-Orellana, J., Masqué, P., Feldman, M., Weinstein, Y., 2015. Submarine 743 groundwater discharge as a major source of nutrients to the Mediterranean Sea. 744 Natl. Acad. U. S. https://doi.org/ Proc. Sci. Α. 112, 3926-3930. 745 10.1073/pnas.1419049112.
- Rosenberry, D.O., Briggs, M.A., Delin, G., Hare, D.K., 2016. Combined use of thermal
 methods and seepage meters to efficiently locate, quantify, and monitor focused
 groundwater discharge to a sand-bed stream. Water Resour. Res. 52, 4486–4503.
 https://doi.org/10.1002/2016WR018808.

- Rücker, C., Günther, T., Wagner, F.M., 2017. pyGIMLi : An open-source library for modelling
 and inversion in geophysics. Comput. Geosci. 109, 106–123. https://doi.
 org/10.1016/j.cageo.2017.07.011.
- Rufí-Salís, M., Garcia-Orellana, J., Cantero, G., Castillo, J., Hierro, A., Rieradevall, J., Bach,
 J., 2019. Influence of land use changes on submarine groundwater discharge.
 Environ. Res. Commun. 1, 031005. https://doi.org/10.1088/2515-7620/ab1695.
- Santos, I.R., Burnett, W.C., Dittmar, T., Suryaputra, I.G.N.A., Chanton, J., 2009. Tidal pumping
 drives nutrient and dissolved organic matter dynamics in a Gulf of Mexico
 subterranean estuary. Geochim. Cosmochim. Acta 73, 1325–1339. https://doi.org/
 10.1016/J.GCA.2008.11.029.
- Selker, J.S., Thévenaz, L., Huwald, H., Mallet, A., Luxemburg, W., Van De Giesen, N., Stejskal,
 M., Zeman, J., Westhoff, M., Parlange, M.B., 2006. Distributed fiber-optic
 temperature sensing for hydrologic systems. Water Resour. Res. 42, W12202.
 https://doi.org/10.1029/2006WR005326.
- Simon, N., Bour, O., Lavenant, N., Porel, G., Nauleau, B., Pouladi, B., Longuevergne, L., 2020.
 A comparison of different methods to estimate the effective spatial resolution of FODTS measurements achieved during sandbox experiments. Sensors 20, 570.
 https://doi.org/10.3390/s20020570).
- Singha, K., Day-Lewis, F.D., Johnson, T., Slater, L.D., 2015. Advances in interpretation of
 subsurface processes with time-lapse electrical imaging. Hydrol. Process. 29, 1549–
 1576. https://doi.org/10.1002/hyp.10280.
- Spiteri, C., Slomp, C.P., Charette, M.A., Tuncay, K., Meile, C., 2008. Flow and nutrient
 dynamics in a subterranean estuary (Waquoit Bay, MA, USA): Field data and reactive
 transport modeling. Geochim. Cosmochim. Acta 72, 3398–3412. https://doi.org/10.
 1016/J.GCA.2008.04.027.

- Stauffer, F., Bayer, P., Blum, P., Giraldo, N.M., Kinzelbach, W., Bayer, P., Blum, P., Giraldo,
 N.M., Kinzelbach, W., 2013. Thermal Use of Shallow Groundwater. CRC Press.
 https://doi.org/10.1201/b16239.
- Taniguchi, M., 2000. Evaluations of the saltwater-groundwater interface from borehole
 temperature in a coastal region. Geophys. Res. Lett. 27, 713–716. https://doi.org/10.
 1029/1999GL002366.
- Trezzi, G., Garcia-Orellana, J., Rodellas, V., Masqué, P., Garcia-Solsona, E., Andersson, P.S.,
 2017. Assessing the role of submarine groundwater discharge as a source of Sr to the
 Mediterranean Sea. Geochim. Cosmochim. Acta 200, 42–54. https://doi.org/10.
 1016/J.GCA.2016.12.005.
- Tyler, S.W., Selker, J.S., Hausner, M.B., Hatch, C.E., Torgersen, T., Thodal, C.E., Schladow,
 S.G., 2009. Environmental temperature sensing using Raman spectra DTS fiber-optic
 methods. Water Resour. Res. 45, W00D23. https://doi.org/10.1029/ 2008WR007052.
 van de Giesen, N., Steele-Dunne, S.C., Jansen, J., Hoes, O., Hausner, M.B., Tyler, S., Selker,
 J., 2012. Double-ended calibration of fiber-optic Raman spectra distributed
 temperature sensing data. Sensors 12, 5471–5485. https://doi.org/10.3390/
 s120505471.
- Waxman, M.H., Smits, L.M., 1968. 1863-A Electrical Conductivities in Oil-Bearing Shaly
 Sands. Soc. Pet. Eng. J. 8, 107–122.
- Werner, A.D., Bakker, M., Post, V.E.A., Vandenbohede, A., Lu, C., Ataie-Ashtiani, B., Simmons,
 C.T., Barry, D.A., 2013. Seawater intrusion processes, investigation and management:
 Recent advances and future challenges. Adv. Water Resour. 51, 3–26.
 https://doi.org/10.1016/J.ADVWATRES.2012.03.004.
- Werner, A. D., 2017.On the classification of seawater intrusion. J. of Hydrol. 551, 619631.
 https://doi.org/10.1016/j.jhydrol.2016.12.012.

800	Windom, H.L.,	Moore, W.S	S., Nienc	heski, L.F.	H., Jahn	ıke, R.A., 20	06. Submarine	groundwater
801	discharg	e: A large,	previou	usly unrec	ognized	source of	dissolved iron	to the south
802	Atlantic	Ocean.	Mar.	Chem.	102,	252–266.	https://doi.o	rg/10.1016/J.
803	MARCHE	M.2006.06	.016.					

- Zarroca, M., Linares, R., Rodellas, V., Garcia-Orellana, J., Roqué, C., Bach, J., Masqué, P.,
- 806 mean of electrical resistivity imaging, 224Ra and 222Rn. Hydrol. Process. 28, 2382

2014. Delineating coastal groundwater discharge processes in a wetland area by

807

805

808 Figure captions

809

Figure 1. Location and pilot site setup in the NW Mediterranean basin. a) Maps depicting the location of the field site with respect to the surrounding water bodies and infrastructures. b) Map showing the distribution of the installed piezometers at the experimental site. The color scale indicates the depth of the screened section of each borehole. The red line depicts the position of the vertical cross-sections shown throughout the article, from A (inland) to B (seawards).)

815

Figure 2. Cross section of the experimental site perpendicular to the seashore (Fig. 1b). The different color lines represent the contact between weathered granite and the alluvial formation (red) and between the anthropic materials and the alluvial sediments (Brown). Black lines represent the correlation between fine materials levels in the different piezometers integrating gamma log data for all boreholes (data not shown). Grey layer indicates a continuous layer of silt that crosses all piezometers.

822

Figure 3. a) Schematic description of piezometer nest monitoring system, including CHERT electrodes (actually, 32 electrodes were installed in each piezometer) and fiber optic cable for DTS. (b) photo of the electrodes and cable installation in the piezometric tube. During extraction of the drilling auxiliary pipe (c), the fiber cable has to be cut, requiring fusion points. The black bar represents the nylon flanges to fix the electrodes and the fiber optic cable on the piezometer.

828

Figure 4. Electrode configurations for the acquisition of CHERT. A and B desig-nate the current electrodes, and M and N the potential electrodes. In the cross-hole dipole–dipole array (CH AB-MN), the current electrodes are in the first borehole while potential electrodes are in the second borehole. In the cross-hole pole-tripole array (CH AMN-B/A-BMN), a current is imposed in the two 833 bore-holes while potential electrodes are in the same borehole.

834

Figure 5. Bulk electrical conductivity model obtained from CHERT data. The anomaly in red, extending throughout the cross-section, 2 m below the continuous layer of silt placed at 12 m depth (in grey), indicates the presence of seawater in the aquifer. The stratigraphic columns of the Argentona site are displayed as a reference for interpretation.

839

Figure 6. EC profiles recorded at the fully screened piezometer PP20 (Fig. 2) in 2015. Fresh and sea
water values are marked with a dashed line in the figure for comparison.

842

Figure 7. Cross-section of the bulk electrical conductivity (CO) ratio between June and September 2015 CHERT surveys. The colour scale varies from a decrease to half of the value of CO compared to June 2015 (blue), to an increase by two in the value of CO compared to June 2015 (red). The main change during summer is a twofold increase in bulk EC observed 2 m below the silt layer at -12 m depth represented in grey. The stratigraphic columns of the Argentona site are displayed as a reference for interpretation..

849

Figure 8. a) Downhole Induction logging (IL) profiles of formation electrical conductivity (C_0) measured in borehole N320 (Fig. 1) on May 11th, 2015 (profiles IL1, IL2 and IL3) between 3:00 and 3:45 PM, and on May 12th, 2015 at 11:30 AM (profile IL23). (b) Stratigraphic column of N325 (Fig. 2) located 1.5 m away from N320 (c) Pore fluid conductivity ratios (ΔC_0), taking the May 12th, 2015 profile IL23 as reference. Grey shadings across the entire figure point at levels where high frequency conductivity changes between profiles recorded on 2015 were detected.

857

858	Figure 9. Thermal profiles of the June and September field surveys resulting from linear
859	interpolation of data in each borehole. Temperatures above 20 °C are all drawn with the same
860	color. The stratigraphic columns of the Argentona site are displayed as a reference for
861	interpretation. Grey layer indicates a continuous layer of silt that crosses all piezometers.

Figures



Figure 1



Figure 2



Figure 3



Figure 4



Figure 5







Figure 7



Figure 8



Figure 9

Supplementary material

Table 1. Groundwater electrical conductivities and temperatures measured in the 2 m screened interval of each piezometer except PP12 and PP15 that are completely screened.

Diamanatan	Electrical Co	nductivity (mS·cm ⁻¹)	Temperature (°C)		
Plezometer	June (16/06/15)	September (10/09/15)	June (16/06/15)	September (10/09/15)	
N115	1.60-2.16	1.56-2.08	19.02-19.70	18.97	
N120	43.0-52.0	42.20-43.22	19.50-18.99	18.86	
N125	39.6-44.4	35.09-35.72	18.91-18.84	18.79	
N215	1.10-1.15	0.82-10.17	19.75-19.61	21.3-21.2	
N220	8.45-14.06	12.30-21.7	19.47-19.37	20.4-20.4	
N225	30.06-32.42	-	19.01-18.98	-	
N315	1.64-1.72	1.45-1.46	19.14-19.19	10.29-11.29	
N320	13.33-19.66	25.83-26.30	19.18-17.92	19.60-19.50	
N325	36.34-41.81	38.99-39.55	19.1-19.03	19.03-19.98	
PP20	8.07-45.33	8.60-39.17	20.11-18.87	23.2-19.6	
PP15	2.24-3.11	2.24-2.78	19.3-19.07	18.8-18.9	