

# Vertical patterns of metabolism in three contrasting stratified lakes

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## ABSTRACT

Recent advances in open water measurements suggest significant temporal and spatial variability of gross primary production (GPP), net ecosystem production (NEP), and respiration (R) with implications for understanding carbon cycling in lakes. This study applied high frequency depth profiles in three stratified lakes of different trophic status to investigate 1) the importance of vertical variations in metabolic rates, 2) the effects of changes in the depth of the mixed layer ( $Z_{mix}$ ) and the photic zone ( $Z_{eu}$ ), and 3) the photoacclimative responses of the aquatic autotrophs to changes in these conditions. Taking account of vertical differences in metabolism improved the reliability of whole-areal NEP estimates during stratification. Whereas the hypolimnion was always heterotrophic, and the epilimnion was mostly autotrophic, the metalimnion had  $NEP > 0$  when  $Z_{eu} > Z_{mix}$ . Although most of GPP and R occurred in the epilimnion, between 0-20% of GPP and 4-37% of R took place in the metalimnion. Areal metabolic estimates based on surface measurements deviated up to 60% for GPP and 80% for R when  $Z_{eu} > Z_{mix}$ . The vertical variability in metabolism was driven by available light in both the epi- and metalimnion. Coupling between GPP and R was low in all layers and indicated increasing background R with depth. Light utilization efficiency was significantly higher under low light conditions, indicating photophysiological acclimation of phytoplankton to decreasing light in the metalimnion.

## 1 **Introduction**

2       Measurements of the production and consumption of organic material has been a focus of aquatic science  
3 for more than 80 years during which an important aim has been to quantify the magnitude and understand the  
4 drivers of variability of metabolic rates and carbon processing. Estimating ecosystem primary production and  
5 respiration from measurements of diel, open water changes in dissolved oxygen (DO) concentrations has  
6 become a widely accepted method in aquatic science and has been particularly popular in lakes (Gelda and  
7 Effler 2002; Staehr et al. 2012a). The recent popularity of the diel open water technique among researchers and  
8 managers relates to awareness that it is important to capture the variable temporal dynamics in metabolic rates  
9 which can be obtained from sensors moored in lakes.

10       As the use of open-water techniques has expanded, new insights and new questions have emerged on the  
11 controls of the temporal and spatial variability of ecosystem gross primary production, (GPP), ecosystem  
12 respiration (R), and net ecosystem production (NEP = GPP – R) (Staehr and Sand-Jensen 2007; Van de Bogert  
13 et al. 2007; Solomon et al. 2013). Here, we investigate vertical patterns of lake metabolism in lakes deep  
14 enough to experience vertical temperature stratification. While this is not a new topic (Bella 1970; Hornberger  
15 and Kelly 1974; Melack 1982), technological advances in providing high frequency DO profiles, and analytical  
16 improvements have made it possible to determine depth specific production and respiration in detail (Staehr et  
17 al. 2012b) and evaluate if vertical day-to-day patterns in metabolism compare with those predicted from models  
18 of light and nutrient dependency.

19       In clear water lakes, the photic zone may extend below the thermocline resulting in elevated primary  
20 production and respiration below the upper mixed layer (Stefan et al. 1995). Such conditions make the  
21 metalimnion a zone of primary production when sufficient light is available and nutrient levels are depleted in  
22 the upper mixed layer (Staehr et al. 2012b). This agrees well with earlier findings that vertical patterns in water  
23 column productivity are controlled by variations in mixed layer depth ( $Z_{mix}$ ) and water clarity through  
24 regulation of light and nutrient availability (Fee 1976; Grobbelaar 1985; Stefan et al. 1995). Under turbid

25 conditions where  $Z_{\text{mix}}$  exceeds the depth of the photic zone ( $Z_{\text{eu}}$ ) phytoplankton cells may spend considerable  
26 time under non-productive dark conditions (Grobbelaar 1990). Under clear water conditions where  $Z_{\text{eu}} > Z_{\text{mix}}$   
27 deeper peaks in phytoplankton biomass and primary production can occur in the metalimnion, as sufficient light  
28 and nutrients are available. Experimental work (Grobbelaar et al. 1995; Mouget et al. 1999; Mellard et al. 2012)  
29 have furthermore shown that phytoplankton light utilization efficiency may increase under conditions of low  
30 light and high nutrient availability.

31 This study evaluated vertical patterns in lake metabolism and in photophysiological acclimation of  
32 phytoplankton by careful analysis of a dataset of high frequency oxygen, temperature and light profiles in three  
33 temperate lakes of different trophic status (from mesotrophic to hypereutrophic). Recent developments in the  
34 use of inverse modeling techniques to determine metabolic rates from diel changes in DO (Hanson et al. 2008;  
35 Batt and Carpenter 2012) allowed us to evaluate the variability in parameters describing the photophysiological  
36 state of the phytoplankton community, including the light utilization efficiency.

37 Our first objective was to quantify the importance of GPP, R, and NEP with depth and investigate  
38 relationships between R and GPP. From previous studies (Coloso et al. 2008; Sadro et al. 2011a; Staehr et al.  
39 2012b) we hypothesize that i) metabolic rates will be higher in the epilimnion compared to the meta- and  
40 hypolimnion due to more available light and strong coupling between GPP and R, ii) GPP will decrease more  
41 than R with depth, and iii) the balance between GPP and R (~NEP) will be related to the prevailing light and  
42 mixing conditions causing the metalimnion to vary between net autotrophic and heterotrophic conditions on a  
43 daily basis. Furthermore we hypothesize that the relative importance of epi-, meta-, and hypolimnetic zones will  
44 reflect their spatial extent and volume specific rates, such that most of the areal production will occur in the  
45 epilimnion, especially in the eutrophic lake where light will be limited to the surface waters, while clear water  
46 lakes are hypothesized to have significant contributions to GPP, NEP, and R from the metalimnetic zone. With  
47 regard to the relationship between R and GPP, ecosystem respiration depends on both autochthonous (~GPP)  
48 and allochthonous (terrestrially derived) organic matter inputs (del Giorgio and Williams 2005). Recent studies

49 (Solomon et al. 2013) show a tighter relationship between R and GPP in oligotrophic than in eutrophic lakes,  
50 where substantial primary production escapes immediate respiration and becomes buried or exported (Caraco  
51 and Cole 2004). Within lakes, we therefore expect a stronger relationship between R and GPP in the epilimnetic  
52 zone (as compared to the meta- and hypolimnion zones) where high light availability stimulates autochthonous  
53 production. Furthermore, we hypothesize that background respiration (i.e., the rate of respiration at  $GPP = 0$ ),  
54 will increase under conditions where the concentration of non-photosynthetic organic material is high (del  
55 Giorgio and Williams 2005). As depth increases and light availability decreases, we expect the relative  
56 contribution of background respiration to total respiration to increase.

57 Our second objective was to investigate the extent to which vertical patterns in water column primary  
58 production are controlled by variations in mixed layer depth ( $Z_{mix}$ ) and the depth of the photic zone ( $Z_{eu}$ )  
59 together determining the ambient light availability. Assuming R to be relatively constant with depth and GPP to  
60 be strongly linked to light, the net ecosystem production (NEP) is hypothesized to show a strong light  
61 dependency. The light dependency is expected to be more pronounced in the metalimnion than the epilimnion,  
62 with significantly higher NEP in the metalimnion during periods where minimum light requirements are met  
63 (Stefan et al. 1995; Domingues et al. 2011). Assuming light to be the dominant driver of vertical changes in  
64 primary production, we finally want to investigate vertical changes in phytoplankton light acclimation. Field  
65 (Grobbelaar 1985; Kishino et al. 1986; MacIntyre et al. 2002) and laboratory studies (Grobbelaar 1995; Mouget  
66 et al. 1999) have previously shown a photoacclimative response to decreasing light by increasing the light  
67 utilization efficiency ( $\alpha$ ; slope of the initial part of the photosynthesis vs. light curve). Experimental studies  
68 (Staeher and Sand-Jensen 2006) have shown that  $\alpha$  is higher under nutrient replete conditions. Therefore, within  
69 the photic zone ( $> 1\%$  surface light) we hypothesize  $\alpha$  to be higher in meta- than epilimnion due to less light  
70 and more nutrients available.

## 71 **Methods**

### 72 *Study sites*

73 This study was conducted in three Danish lakes which stratify most of the summer and differ in nutrient  
74 and dissolved organic carbon (DOC) loads and therefore in water clarity. The lakes in this study also differ in  
75 morphometry, residence time and mixed layer depth (Table 1). Lake Hampen (9.4°E, 56.1°N) is a mesotrophic  
76 polymictic lake, and was sampled from April to December 2007. Lake Vedsted (9.4°E, 55.2°N) is a small,  
77 eutrophic and dimictic lake, and was sampled from January to December 2008. Lake Castle (12.3°E, 55.9°N) is  
78 a hypereutrophic and polymictic lake, and was sampled from September to November 2006 (Fig. 1).

### 79 *Monitoring stations*

80 Continuous monitoring of oxygen concentration, electrical conductivity and temperature at different  
81 depths was performed through an automatic profiling mooring station equipped with a multiparametric sonde  
82 (Yellow Springs Instruments 6600), and placed at the deepest area in each lake. The DO sensor has a resolution  
83 of 0.01 mg L<sup>-1</sup> with an accuracy of ± 0.1 mg L<sup>-1</sup> and the temperature sensor has a resolution of 0.01°C with an  
84 accuracy of ± 0.15°C. Dissolved oxygen (DO) was measured with a self-cleaned optical sensor that was  
85 calibrated every 3 weeks during each deployment. No drifts in the sensor were observed between calibrations.

86 The sonde performed automatic profiles every 30 min, measuring data at specified depths from the  
87 surface to the bottom of the water column (Table 1). We allowed 3 min of sensor stabilisation prior to each  
88 measurement. We had a DO and temperature measure for each depth every 30 minutes during the deployment  
89 period. Wind speed (Onset<sup>®</sup> anemometer) and photosynthetic active radiation (2π Onset<sup>®</sup> light logger) were  
90 collected at 5 second intervals and recorded as 10 minute averages at 1.3 m above the water surface at the  
91 mooring location.

92 Underwater light conditions were determined every 10 minutes using a series of four underwater Onset<sup>®</sup>  
93 light sensors placed at 20, 40, 80, and 120 cm depth. The daily vertical light attenuation coefficient ( $K_D$ , 400-  
94 700 nm) was determined as the slope of a linear regression model of irradiance ( $E_z$ ) vs. depth ( $z$ ):  $\ln(E_z) = b +$   
95  $K_D z$ . Continuous irradiance recordings integrated over 30 minute intervals, allowed 48 light profiles each day.  
96 We then calculated a daily mean of vertical light attenuation (during light hours) from  $K_D$  values estimated

97 from regressions with  $r^2 > 0.8$  according to Staehr et al. (2012b). Daily  $K_D$  values determined from HOBO  
98 sensors were finally corrected by linear regression analysis against biweekly  $K_D$  estimates based on underwater  
99 light profiles using a  $4\pi$  LiCor photosynthetic available radiation (PAR) sensor. The depth of the photic zone  
100 (1% of surface light) was calculated for each day as  $4.6 / K_D$ .

### 101 *Vertical stratification*

102 We divided the water column into three layers epilimnion, metalimnion and hypolimnion according to the  
103 procedure in Staehr et al. (2012b). We determined the depth of epilimnion as the shallowest depth with a water  
104 density gradient equal to or above a suitable threshold (around  $0.07 \text{ kg m}^{-3} \text{ m}^{-1}$ ) for each time step according to  
105 Read et al. (2011) and will refer to that depth as  $Z_{\text{mix}}$ . Water density was calculated from temperature (in °C)  
106 assuming negligible effects of solutes as in Read et al. (2011). The thermocline depth was defined as the depth  
107 with the maximum temperature gradient and the lower limit of the metalimnion was calculated as thermocline  
108 depth plus the distance between  $Z_{\text{mix}}$  and the thermocline. Here we assume a symmetric temperature gradient  
109 around the thermocline. The depths of each layer were calculated for every 30 min, from modeled temperature  
110 profiles according to the procedures described in Staehr et al. (2012b). Measured profiles of temperature were  
111 fitted to a continuous curve model (Rimmer et al. 2006), which we only used for determination of  $Z_{\text{mix}}$ ,  
112 thermocline and metalimnion depths. The performance of the curve-fitting model was evaluated by comparing  
113 temperature profiles in another Danish lake measured at 0.25 m and 1 m depth intervals. Only small and non-  
114 significant (Student's  $t$ -test,  $t = 1.82$ ;  $df = 9$ ;  $p = 0.08$ ) differences were observed in the calculated vertical  
115 temperature profiles at 0.1 m depth resolution. We recognize that our smoothed profiles do not represent the  
116 actual fine-resolution variations in temperature or density that would be obtained from microstructure profiles  
117 (MacIntyre 1993, Imberger 1985a).

118 We restricted our analysis to the periods where temperature stratification prevailed (Table 1). Further  
119 details on the automatic profiling system and on the methods to estimate vertical stratification of the water  
120 column can be found in Staehr et al. (2012b).

121 *Metabolic calculations*

122 Data treatment prior to metabolic calculations included simple linear interpolation with depth to obtain  
 123 DO and temperature values for every 30 minutes and every meter interval, followed by a temporal smoothing of  
 124 the time series at each depth through a running average of four hours. The same procedures were applied to  
 125 wind and light data.

126 Metabolic rates for each depth layer were calculated using a methodology that includes biological fluxes  
 127 (metabolism), air-water gas exchange and DO exchange between depth layers driven by mixed-layer deepening  
 128 and eddy diffusivity (Staeher et al. 2012b). The basic model assumes that the DO change between two  
 129 consecutive time steps in a given depth layer  $i$  ( $\frac{\Delta O_{2(i)}}{\Delta t}$ , in  $\text{mmol m}^{-3} \text{h}^{-1}$ ) is described by:

$$130 \quad \frac{\Delta O_{2(i)}}{\Delta t} = \text{NEP}_i + D_{z(i)} - D_{v(i)} - D_{s(i)} \quad (1)$$

131 where  $\text{NEP}_i$  is net ecosystem production,  $D_{z(i)}$  is the flux between layers driven by mixed-layer  
 132 deepening,  $D_{v(i)}$  is the flux between layers driven by eddy diffusivity, and  $D_{s(i)}$  is the air-water gas exchange, all  
 133 expressed in  $\text{mmol m}^{-3} \text{h}^{-1}$ .

134 Above  $Z_{\text{mix}}$   $D_s$  was calculated as  $D_{s(i)} = K_s(O_{2(i)} - O_{2\text{sat}(i)})/Z_{\text{mix}}$ , where  $K_s$  is the gas transfer velocity  
 135 at the in situ temperature ( $\text{m h}^{-1}$ ),  $O_{2(i)}$  is the measured DO concentration and  $O_{2\text{sat}(i)}$  is the DO concentration in  
 136 atmospheric equilibrium. Depth layers below  $Z_{\text{mix}}$  were considered to be isolated from atmosphere.  $K_s$  was  
 137 calculated from the gas transfer velocity at a Schmidt number of 600, which was in turn calculated for the  
 138 cooling and heating periods from wind speed data following the equations of MacIntyre et al. (2010).  $D_{z(i)}$  was  
 139 calculated using changes in  $Z_{\text{mix}}$  (Staeher et al. 2012b), and  $D_{v(i)}$  was calculated from the vertical turbulent  
 140 diffusivity coefficient ( $K_v$ ) following Hondzo and Stefan (1993). This parameterization of eddy diffusivity has  
 141 significant limitations, especially if applied to diel variations, and has not been validated in comparison to other  
 142 approaches based on theoretical and mechanistic understanding of the mixing processes (Imberger 1985b;  
 143 MacIntyre et al. 1999; Yeates and Imberger 2004). Our measurements did not permit application of these

144 approaches. To evaluate the potential effect of uncertainties in  $K_v$  we performed a sensitivity analysis based on  
 145 allowing  $K_v$  to vary up to 3 orders of magnitude (from  $10^{-4}$  to  $10^{-6}$   $\text{m}^2 \text{s}^{-1}$ , which is a large but reasonable range  
 146 for our lakes) and evaluating the variability in the derived R and GPP rates in epi-, meta-, and hypolimnion  
 147 during a 30 day period in Hampen Lake. We observed minor changes in the daily metabolic rates to changes in  
 148  $K_v$ , thus supporting the validity of our results. Differences in R were always below  $2 \text{ mmol O}_2 \text{ m}^{-3} \text{ d}^{-1}$  (mean  
 149 coefficient of variation (CV) was below 8% in all layers). Differences in GPP were always below  $1 \text{ mmol O}_2 \text{ m}^{-3} \text{ d}^{-1}$  (mean CV 2% in epi- and metalimnion, and 13% in hypolimnion).

151 Metabolic rates were calculated using an inverse modeling procedure (Hanson et al. 2008) which  
 152 calculates  $\text{NEP}_i$  from photosynthetically active radiation ( $\text{PAR}_i$ ,  $\mu\text{mol m}^{-2} \text{s}^{-1}$ ) and temperature ( $T_i$ , °C) at each  
 153 specific depth  $i$ . The model includes a saturating response of photosynthesis to PAR (Jassby and Platt 1976) and  
 154 a thermodependent respiration, and is described as:

$$155 \quad \text{NEP}_i = P_{\max} \tanh\left(\frac{\alpha \text{PAR}_i}{P_{\max}}\right) - R_{20} \theta^{(T_i - 20)} \quad (2)$$

156 where  $P_{\max}$  is the maximum photosynthetic rate at saturating light ( $\text{mmol m}^{-3} \text{ h}^{-1}$ ),  $\alpha$  is the photosynthetic  
 157 efficiency ( $\text{mmol O}_2 \text{ m}^{-3} \text{ h}^{-1} (\mu\text{mol photons m}^{-2} \text{ s}^{-1})^{-1}$ ),  $R_{20}$  is the respiration rate at  $20^\circ\text{C}$  ( $\text{mmol m}^{-3} \text{ h}^{-1}$ ) and  $\theta$  is  
 158 a coefficient which stands for the thermal dependence of respiration (set to 1.07; Jorgensen and Bendoricchio  
 159 2001). The first term in Eq. 2 corresponds to gross primary production ( $\text{GPP}_i$ ,  $\text{mmol m}^{-3} \text{ h}^{-1}$ ), and the second to  
 160 ecosystem respiration ( $R_i$ ,  $\text{mmol m}^{-3} \text{ h}^{-1}$ ) for each depth layer.  $\text{PAR}_i$  was obtained from continuous surface  
 161 PAR measurements and from the light attenuation coefficient in the water column ( $K_D$ ,  $\text{m}^{-1}$ ).

162 A model combining Eqs. 1 and 2 was fitted to the DO data for 24 hour periods using a numerical  
 163 minimization algorithm in the non-linear function in the Statistical Analysis System (SAS) software. Thus, by  
 164 fitting the model to the observed DO time series at each depth, we obtained, for every 24 h and every depth  
 165 layer, estimates of the parameters  $P_{\max}$ ,  $\alpha$ , and  $R_{20}$ . We assessed model performance (i.e., how well the model  
 166 fitted the observed DO data) through the coefficient of determination ( $r^2$ ).

167 By applying the parameters ( $P_{\max}$ ,  $\alpha$ , and  $R_{20}$ ) derived in Eq. 2 we calculated hourly metabolic rates for  
168 each depth ( $i$ ) ( $NEP_i$ ,  $GPP_i$ , and  $R_i$ , in  $\text{mmol m}^{-3} \text{h}^{-1}$ ). The daily metabolic rates (in  $\text{mmol m}^{-3} \text{d}^{-1}$ ) were  
169 calculated as the average hourly rates multiplied by 24 h. Whole-lake areal GPP, R, and NEP ( $\text{mmol m}^{-2} \text{d}^{-1}$ )  
170 were computed by multiplying the volumetric daily rates of each depth layer ( $\text{mmol m}^{-3} \text{d}^{-1}$ ) by the water  
171 volume within each layer ( $\text{m}^3$ ), and by summing these quantities and dividing by the surface lake area ( $\text{m}^2$ ).

## 172 *Water analysis*

173 Biweekly water samples were collected from the epilimnion and measurements of chlorophyll *a* (Chl *a*),  
174 total phosphorous (TP), and colored dissolved organic matter (CDOM). Chl *a* samples were filtered through  
175 Advantech  $0.7 \mu\text{m}$  filters, extracted in 96% ethanol for 24 h, and measured according to methods in Jespersen  
176 and Christoffersen (1987) using a Shimadzu ultraviolet (UV)-160AH spectrophotometer. Total phosphorus  
177 (TP) was determined by persulfate digestion according to Eaton et al. (1995) and further assayed according to  
178 Kragh and Sondergaard (2004). Absorbance of CDOM was measured in a GF/F filtrate ( $0.7 \mu\text{m}$ ) at 360 nm  
179 through a 5 cm cuvette. In comprehensive measurements from Danish lakes and streams, DOC concentrations  
180 ( $\text{mg C L}^{-1}$ ) increased linearly with CDOM absorbance at 360 nm ( $\text{m}^{-1}$ ):  $\text{DOC} = 0.454 \text{CDOM}_{360} + 1.9$  ( $r^2 = 0.80$ ,  
181  $n = 399$ , C. Stedmon unpubl.) permitting estimation of DOC from CDOM with an uncertainty of 20%.

## 182 **Results**

### 183 *Vertical patterns in metabolism*

184 As typically observed for Danish lakes (Stæhr et al. 2010b), stratification commenced in all lakes when  
185 surface temperatures passed approximately  $10^\circ\text{C}$  in May (not shown). The duration and stability of the  
186 stratification however varied among the studied lakes. Mixing depth and extent of the metalimnetic zone was  
187 smaller and much more variable in the polymictic Hampen Lake compared to the dimictic Vedsted Lake (Fig.  
188 1; Table 1). Vedsted Lake is the smallest lake and it is sheltered between hills and surrounding trees, causing  
189 the thermocline to be more stable compared to Hampen and Castle lakes which are situated in open and more  
190 wind exposed landscapes. In Hampen Lake the photic zone penetrated far into the metalimnion whereas the

191 metalimnion received much less light in Vedsted Lake and in Castle Lake the metalimnion did not have light  
192 levels above 1% (Fig. 1; Table 1).

193 All lakes had net autotrophy ( $NEP > 0$ ) in the photic zone and heterotrophy in the layers below the photic  
194 depth (Figs. 2,3). Accordingly the autotrophic zone extended deeper in Hampen than in Vedsted Lake and was  
195 limited to the upper meter in Castle Lake (Figs. 2,3). Hampen had larger variability in photic depth (5 to 9 m)  
196 than the two other lakes thus resulting in variability in GPP and NEP within the metalimnion (Fig. 3) which was  
197 higher than the variability in Vedsted and Castle Lake. The metalimnion was heterotrophic in Castle (NEP  
198 range from  $-93.6$  to  $-0.7$   $\text{mmol m}^{-3} \text{d}^{-1}$ ; mean  $-23.1$ ; Fig. 3), mostly heterotrophic in Vedsted (NEP range from -  
199  $84.1$  to  $63.7$   $\text{mmol m}^{-3} \text{d}^{-1}$ ; mean  $-25.7$ ), and varied from heterotrophic to slightly autotrophic in Hampen Lake  
200 (NEP range from  $-96.5$  to  $81.3$   $\text{mmol m}^{-3} \text{d}^{-1}$ ; mean  $-7.8$ ). Whereas heterotrophy progressively increased with  
201 depth in Hampen, the most negative NEP in Castle and Vedsted occurred in the metalimnion, not in the  
202 hypolimnion.

203 GPP decreased with depth in all three lakes, and the highest GPP was observed in the epilimnion except  
204 in Hampen, where GPP rates in the metalimnion occasionally exceeded GPP in epilimnion (Fig. 2 and 3). All  
205 lakes had highest R rates in the upper epilimnion and the metalimnion. Differences in the physical structure and  
206 light penetration between the three lakes were reflected in the contribution of the three depth strata to areal  
207 metabolism (Fig. 4; Table 2). Despite the epilimnion being the main contributor to GPP and R, a considerable  
208 contribution of non-epilimnetic waters was observed, especially in Hampen and Vedsted Lake (Table 2). In  
209 these lakes, epilimnetic respiration accounted for 50-70% of total R. Most of the remaining R (about 30%)  
210 occurred in the metalimnion (Table 2). The metalimnion accounted for up to 18% of the whole-column GPP on  
211 average. In agreement with expectations no primary production occurred in the meta- and hypolimnetic waters  
212 of Castle Lake, due to the shallow photic depth. Only 4% of areal ecosystem respiration occurred below the  
213 upper mixed layer in Castle Lake.

214 Calculated areal rates of NEP based on depth specific estimates over the whole-column were lower than  
215 NEP calculated from a single sonde in the upper mixed layer (Fig. 4). Whereas the single-sonde approach  
216 provided positive values of areal NEP, the whole-column estimates were significantly lower than zero for both  
217 Vedsted ( $t = -5.6$ ;  $df = 62$ ;  $p < 0.01$ ) and Castle ( $t = -1.5$ ;  $df = 31$ ;  $p < 0.01$ ) lakes. The discrepancy was most  
218 evident in Vedsted Lake where the metalimnetic zone was generally net heterotrophic. Thus single-sonde  
219 estimates taken at 1m depth overestimated areal GPP, R, and NEP rates (Fig. 4). In Vedsted and Castle lakes  
220 the heterotrophic character of the water column was not observed when using metabolic estimates derived from  
221 single-sonde measurements. Despite that most of the metabolism occurred in the epilimnion, rates of GPP and  
222 R were significantly ( $p < 0.05$ ) overestimated when calculated from 1 meter depth.

### 223 *Coupling between photosynthesis and respiration*

224 Relationships between ecosystem respiration and gross primary production were generally weak,  
225 especially so for the hypereutrophic Castle Lake which had no significant relationships ( $p > 0.05$ ) based on a  
226 whole water column or strata specific comparison (Table 3). Using the coefficient of determination ( $r^2$ ) of the  
227 relationship as a proxy of the direct coupling between R and GPP, suggested a stronger coupling in epilimnetic  
228 waters than in the deeper strata as expected. Background respiration rates were defined as the respiration rates  
229 in absence of primary production (i.e. respiration based on the organic pool of allochthonous organic material).  
230 These were determined as the y intercepts of the R vs. GPP relationships and were significantly lower for  
231 Hampen Lake than the two other lakes when comparing the whole water column ( $p < 0.01$ ), and the epilimnetic  
232 zone ( $p < 0.05$ ). This result agreed with the lower DOC concentrations in Hampen Lake (Kruskall-Wallis  
233  $p < 0.01$ ) compared to the other lakes (Table 1). Within lakes there was a tendency for higher background R in  
234 deeper waters (meta- and hypolimnion) compared to the upper mixed layer in Hampen and Vedsted lakes, but  
235 for Castle Lake the tendency was opposite (Table 3).

### 236 *Community response to light availability*

237 We investigated if variations in light availability caused by changes in mixing depth and depth of the  
238 photic zone, determines the rate of primary production in the different zones of the water column. To test this  
239 we used GPP estimates as an indicator of primary production and divided data for each lake into measurements  
240 performed in epi-, meta- and hypolimnetic layers and whether they were in the photic or aphotic zone. Effects  
241 of depth and light zone were seen in Hampen Lake (Fig. 5) which had the largest variations in  $Z_{\text{mix}}$  and  $Z_{\text{eu}}$   
242 (Table 1). A two way ANOVA showed that the aphotic layers in Hampen Lake had significantly lower primary  
243 production, thus confirming the decrease in GPP with depth seen in Fig. 3. Only epilimnion and metalimnion  
244 were analyzed as hypolimnion data had significantly different variances ( $F_{2,62} = 19, p < 0.01$ ) between the lakes.  
245 Pairwise comparisons with Tukey post hoc test showed that light conditions (photic or aphotic) had a  
246 significant effect on GPP in both epilimnion ( $p < 0.01$ ) and metalimnion ( $p < 0.001$ ). It also showed that for  
247 Hampen Lake, GPP was not significantly different between the photic zones of the epilimnion and metalimnion  
248 and likewise with aphotic zones ( $p > 0.05$ ). In Vedsted Lake pairwise comparisons in epi- and metalimnion  
249 showed that the epilimnion had higher GPP than the metalimnion, where there was no significant difference in  
250 GPP of the photic and aphotic zone ( $p > 0.05$ ). For Castle Lake only the epilimnion was considered since the  
251 photic zone did not penetrate below the mixed layer.

252 Given our observations that R varies less with depth than GPP which is strongly coupled to light, the net  
253 ecosystem production (NEP) is likely to have a strong light dependency. In agreement with these expectations,  
254 relationships between daily volumetric rates of NEP with depth and available light were stronger in the light  
255 exposed epi- and metalimnion than in the hypolimnion, where only Hampen Lake had a few days with light  
256 above 1% surface irradiance (Fig. 6; Table 4). Pooling data for all three lakes, we found a light dependency,  
257 well described in the epi- and metalimnetic layers by the commonly applied light vs. photosynthesis model of  
258 Jassby and Platt (1976). Overall the epilimnion was mostly (55%) net autotrophic, with a minimum daily light  
259 requirement of  $37 \mu\text{mol photons m}^{-2} \text{s}^{-1}$ . In comparison, the minimum light requirement was lower in the  
260 metalimnion ( $15 \mu\text{mol photons m}^{-2} \text{s}^{-1}$ ) but these conditions were rarely met, and the metalimnion was therefore  
261 dominated (86%) by net heterotrophic conditions (Fig. 6).

262           Given the observed dependence of volumetric rates of GPP and NEP on combinations of  $Z_{\text{mix}}$  and  $Z_{\text{eu}}$   
263 (Figs. 2,3,5), we further analysed our data to determine if there was a common threshold in the ratio between  
264  $Z_{\text{mix}}$  and  $Z_{\text{eu}}$  at which areal estimates of metabolism would differ significantly between water column integrated  
265 and those obtained from epilimnetic measurements (~single sonde) only. Combining data for all three lakes, we  
266 found an asymptotic relationship between the ratio of  $Z_{\text{mix}}$  to  $Z_{\text{eu}}$  and the relative contribution of the epilimnetic  
267 zone to whole water column GPP (Fig. 7A). Accordingly, epilimnetic measurements provide representative  
268 estimates of the entire water column when the mixing depth is deeper than the photic zone ( $Z_{\text{mix}} > Z_{\text{eu}}$ ).  
269 However, for most of our measurements the photic zone extended beyond the mixing depth, causing areal GPP  
270 to be underestimated by up to 60% under conditions of high water clarity and a shallow mixed layer (Fig. 7A).  
271 A similar analysis for areal rates of respiration, indicated a hyperbolic relationship to  $Z_{\text{mix}} : Z_{\text{eu}}$  with deviations  
272 up to 80% between R derived from epilimnetic and depth integrated measurements (Fig. 7B). The hyperbolic  
273 relationship was not as strong as for GPP, supporting previous findings that R is not strongly coupled to GPP  
274 via light, but is dependent on the allochthonous inputs which should not vary with depth.

275           For metabolic rates determined within the photic zone (> 1% surface light) we expected light utilization  
276 efficiency ( $\alpha$ ) to increase with depth due to acclimation to less light and higher levels of nutrients. To test this  
277 we compared mean values of  $\alpha$  in the epi- and metalimnion, assuming no vertical gradients within each of these  
278 layers due to internal mixing. The selection criteria of > 1% surface light was met in the epi- and metalimnion  
279 in Hampen and Vedsted lakes, and only the epilimnion of Castle Lake. In both Hampen and Vedsted lakes  $\alpha$   
280 was significantly higher in the metalimnion ( $\alpha = 0.30 \pm 0.18$  and  $0.27 \pm 0.14$  in Hampen and Vedsted,  
281 respectively, mean  $\pm$  standard deviation (SD)) compared to the epilimnion ( $\alpha = 0.15 \pm 0.11$  and  $0.14 \pm 0.11$ ,  
282 respectively) ( $p < 0.01$ ; Wilcoxon paired test).

## 283 **Discussion**

### 284 *Vertical patterns in metabolism*

285           The studied lakes became increasingly heterotrophic with depth during periods of thermal stratification.  
286           This is because GPP is strongly light and therefore depth dependent while respiration is almost independent of  
287           light (apart from photorespiration, and elevated respiration after sunset, following changes in the DOC pool,  
288           Sadro et al. 2011a) and more constant over depth. These patterns are as expected and similar to other studies  
289           (Stefan et al. 1995; Coloso et al. 2008; Staehr et al. 2012b). Respiration peaked around the thermocline as  
290           described in other studies (Salonen et al. 1983; Stefan et al. 1995; Sadro et al. 2011a) and suggested to result  
291           from accumulation of settling particles around the thermocline (Stefan et al. 1995; Staehr et al. 2012b).  
292           However, from this study, it is evident that elevated net heterotrophy occurs in depths which are located both in  
293           the metalimnic zone and below the photic zone (Fig. 2).

294           Variations in light availability and mixing depth had, as expected (Fee 1976; Abbott et al. 1984;  
295           Grobbelaar 1985), a strong effect on depth specific rates of GPP in all three lakes (Figs. 3,5). Within both epi-  
296           and metalimnion in Hampen Lake, GPP was higher on days when light availability was higher than 1% surface  
297           irradiance. As respiration was not related to light conditions (Fig. 3), elevated GPP on days when strata were  
298           within the photic zone, resulted in net autotrophy ( $NEP > 0$ ), even in the metalimnion. Interestingly, GPP was at  
299           comparable levels in the photic zone of epi- and metalimnion in Hampen Lake (Fig. 5), despite that the  
300           epilimnion received three times more light than the photic zone in the metalimnion on average. This suggests  
301           significantly higher light utilization efficiency ( $\alpha$ ) in the metalimnetic zone, as will be discussed later. In  
302           comparison, significantly lower GPP was observed in the photic zone of the metalimnion than in the epilimnion  
303           in Vedsted Lake (not shown). This is due to the definition of the photic zone as 1% of surface irradiance. Thus  
304           if we compare the actual irradiance in the metalimnion, the maximum irradiance in Vedsted was  $34 \mu\text{mol}$   
305           photons  $\text{m}^{-2} \text{s}^{-1}$  but up to  $354 \mu\text{mol photons m}^{-2} \text{s}^{-1}$  in Hampen Lake. Therefore GPP can reach a higher level in  
306           metalimnion in Hampen than in Vedsted with rates comparable to epilimnion. In Castle Lake GPP in  
307           metalimnion is absent due to lack of light.

308           Since GPP were calculated with an inverse model using hourly light values as a driving variable, there  
309 could be a potential bias in our interpretation of light dependency of daily rates. Therefore we compared our  
310 light dependent model with both the standard book-keeping approach (BKA, Hanson et al. 2008; Staehr et al.  
311 2010a) and an inverse modeling approach (IMA) using surface irradiance and not available irradiance as the  
312 forcing function at all depths. These three approaches confirmed the light dependency model. The IMA using  
313 surface irradiance as forcing function at all depths showed no differences in epilimnetic GPP in all three lakes  
314 (median difference < 0.1%), and the descending trend with depth in GPP was maintained. In the metalimnion  
315 the median difference between rates obtained using IMA and BKA was low in Hampen (median 0-4 mmol O<sub>2</sub>  
316 m<sup>-3</sup> d<sup>-1</sup>), higher in Vedsted (median 20-35 mmol O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup>) and relatively large in Castle lake (median 25-70  
317 mmol O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup>), but production rates (up to 1000 mmol O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup>) in Hampen and Castle Lake following  
318 periods of mixing convinced us that these estimates were affected by physical fluxes and did not reflect a  
319 biological signal, as discussed below. The difference in hypolimnetic GPP rates was also relatively low in all  
320 the lakes (median 1-20 mmol O<sub>2</sub> m<sup>-3</sup> d<sup>-1</sup>). Comparing metabolic rates from the traditional BKA with those from  
321 our model (IMA using depth specific irradiance), we found almost identical daily NEP (epilimnetic data:  $r^2 =$   
322 0.9,  $p < 0.01$ ) but more variable GPP and R for the BKA, especially with depth. As a result, light dependencies  
323 of GPP based on the BKA were weaker than the corresponding analysis on GPP derived through IMA. Thus,  
324 while IMA-based rates of GPP confirms the conceptual model of combined effects of light availability and  
325 mixing depth (Fee 1976; Abbott et al. 1984; Grobbelaar 1985), we cannot exclude a bias from the inherent light  
326 dependency of the applied model though the overall pattern is maintained using model approaches independent  
327 of light extinction. Future evaluations of the regulating role of light in stratified lakes, would therefore benefit  
328 from in situ bottle incubations, and should include truly oligotrophic lakes, where it is expected that  
329 metalimnetic GPP should be even higher than in the epilimnion (Sadro et al. 2011a).

330           For metabolic rates calculated below the epilimnion physical variations resulting from internal waves and  
331 seiches may be problematic for dissolved oxygen measurements and compromise the metabolic estimates in the  
332 metalimnion, where these signal oscillations are often greatest (Monismith 1986). We examined the effects of

333 this issue on our estimates of metabolism through a wavelet analysis (Coloso et al. 2008), which identified  
334 significant diel oxygen signals around the metalimnion in both Hampen and Vedsted lakes. We did not register  
335 any significant DO signals at other periods than the diel and therefore we concluded that most of the variation  
336 in DO was due to metabolic processes and not internal water movements. Also, the fit of our model (Eq. 2) to  
337 observed data showed an increasing model error with depth, but with on average 30-50% variability in diel DO  
338 explained in the metalimnion. Because the IMA is less sensitive to the physical variability than the BKA (Batt  
339 and Carpenter 2012) we feel confident that the observed patterns of metabolism are valid. We actually found  
340 that with depth the BKA provided variable rates which poorly correlated with the IMA-based rates, whereas a  
341 good correlation between the two methods was observed only in the surface waters.

#### 342 *Contributions of depth layers to total metabolism*

343 Epilimnetic measurements provide representative estimates of the entire water column when the  
344 epilimnion is well mixed and the mixing depth is deeper than the photic zone ( $Z_{\text{mix}} > Z_{\text{eu}}$ ). When  $Z_{\text{eu}} > Z_{\text{mix}}$  the  
345 deviation of areal metabolic estimates based on epilimnetic measurements only can be as high as 60% for GPP  
346 and 80% for R. Hence, we recommend estimating metabolism from vertically distributed DO measurements in  
347 lakes where the meta- and hypolimnion includes a substantial part of the total lake volume and where the photic  
348 depth is higher than or equal to the mixing depth.

349 Despite the relatively low volume of meta- and hypolimnion in the three lakes and the less favorable light  
350 conditions, the layers under epilimnion contribute significantly to the overall areal metabolism (Table 2). Only  
351 in Castle Lake where meta- and hypolimnion makes up less than 9% of the total volume and are in dark at all  
352 times, the contribution is insignificant. GPP therefore decreases strongly with depth in Castle Lake, and primary  
353 production is absent under the mixed layer due to lack of light (Fig. 3). Even the epilimnion seems to be  
354 separated into a more productive and a less productive zone. This can be related to the existence of secondary  
355 thermoclines within the mixed layer, or at least a mixing process which is slower than the production of oxygen  
356 in epilimnion and gives rise to deviation in single sonde estimates and depth integrated metabolism within  
357 epilimnion, as seen in other studies (Eckert et al. 2002; Coloso et al. 2008).

358 While all of the metabolic activity in Castle Lake occurs in the upper mixed layer, (Table 2) the strong  
359 vertical variability within the epilimnion would not be adequately measured by a single sonde. In comparison,  
360 Hampen and Vedsted had about 80% of areal GPP and 50% to 70% of respiration captured in the epilimnion  
361 (Fig. 4 and Table 2). Respiration in metalimnion represents 26% and 37% of the whole column areal respiration  
362 in Hampen and Vedsted lakes, respectively. This is more than the volume represented by this layer on average  
363 (11% and 18%). Importance of metabolism under the mixed layer has previously been investigated in an  
364 oligotrophic system (Sadro et al. 2011a) and a mesotrophic lake (Staeher et al. 2012b). Here we extend this  
365 analysis to eutrophic lakes and show how metabolism varies with depth in lakes differing in water clarity and  
366 trophic status. We generally confirm previous findings from relative clear water lakes that GPP decrease with  
367 depth while R is less depth dependent (Coloso et al. 2008; Staeher et al. 2012b). However periods with  
368 significant primary production and respiration below the thermocline were also observed (Figs. 2,3).

369 The small contribution of respiration in meta- and hypolimnion of Castle Lake can be explained by a  
370 combination of lake morphometry and anoxic conditions. Castle Lake is a shallow lake with a relatively small  
371 part of the volume in meta- and hypolimnion compared to the other lakes (Table 1). In addition there is anoxic  
372 condition in hypolimnion during the study period (maximum 4 % DO saturation and mean 0.3%) and therefore  
373 aerobic respiration is low or zero and respiration can only be carried out with other electron acceptors which we  
374 did not measure. Due to this we underestimate total respiration in hypolimnion in Castle Lake and cannot  
375 compare this with Hampen Lake where DO was present and Vedsted Lake where oxygen was usually present  
376 (only anoxic at 9 meters) and hence respiration will give rise to oxygen consumption.

### 377 *Coupling between R and GPP*

378 Coupling between GPP and R was weak in our lakes (maximum correlation was 0.49; Table 3) compared  
379 to previous studies in lakes (Sand-Jensen and Staeher 2009; Coloso et al. 2011; Laas et al. 2012) where  
380 correlations were generally higher than 0.9. Whereas our metabolic rates were obtained from DO diel data  
381 using IMA, previous studies have applied the traditional book-keeping approach, where daily GPP is

382 determined as daytime NEP plus total daily R (Hanson et al. 2003). Consequently, the GPP to R coupling in  
383 previous studies, are likely biased because GPP was not determined independently from calculated respiration  
384 values. To further evaluate this we compared GPP and R using the BKA with GPP and R based on IMA. We  
385 found a stronger correlation ( $r^2$  between 0.19 and 0.96,  $p < 0.05$ ) between GPP and R based on the BKA  
386 compared to the rates found through IMA. Therefore it appears that previous studies using the BKA may have  
387 overestimated the dependency of R on GPP. This is further supported by a recent study by Solomon et al.  
388 (2013) who applied IMA and also found a weak coupling between R and GPP.

389 Consistent with expectations of substrate-limited bacterial R in oligotrophic lakes (Sadro et al. 2011b) we  
390 found the strongest coupling between respiration and primary production in the nutrient poorest mesotrophic  
391 Hampen Lake. This agrees with Solomon et al. (2013) who recently showed that the R-GPP coupling is higher  
392 in oligotrophic than eutrophic lakes. Interestingly, the correlation between GPP and R tended to decrease with  
393 depth in all three lakes while background respiration increased. This further supports the interpretation that lake  
394 respiration becomes less dependent on the activity of primary producers as autochthonous carbon is less  
395 available to heterotrophs. Background respiration for the whole water column was higher in Vedsted and Castle  
396 lakes compared to Hampen Lake (Table 3), which was consistent with the higher DOC level in these two lakes  
397 (Table 1). In addition we measured metabolism in a period following a large cyanobacterial bloom in Castle  
398 Lake. Subsequent degradation of this biomass would be interpreted as elevated background respiration  
399 throughout the water column. The lower rates of background respiration determined for the hypolimnetic waters  
400 of Castle Lake are therefore likely a result of absence of changes in oxygen concentrations caused by the anoxic  
401 conditions occurring near the bottom in this eutrophic lake as previously mentioned.

#### 402 *Acclimation to light*

403 Further evidence for the importance of light was provided by the saturating light dependency of NEP for  
404 epi- and metalimnion for all lakes and hypolimnion for Hampen Lake (Fig. 6 and Table 4). Lack of light  
405 dependence in the hypolimnion of Vedsted and Castle Lake occurred because minimum light requirements ( $I_c$ )

406 to sustain positive net photosynthesis were not achieved. Saturating light dependency of NEP confirms the  
407 findings in Staehr et al. (2012b) where NEP is shown to approach balanced ( $NEP = 0$ ) or slightly autotrophic  
408 ( $NEP > 0$ ) conditions in metalimnion when more than 1% of surface light is available, and is heterotrophic  
409 ( $NEP < 0$ ) below this threshold. Interestingly,  $I_c$  values were lower in the metalimnetic zone suggesting  
410 acclimation to lower light conditions through higher light utilization efficiency ( $\alpha$ ). This indication was  
411 supported from the model derived estimates of  $\alpha$  which were significantly higher in the metalimnion than in the  
412 epilimnion in both Hampen and Vedsted lakes. Bottle incubations in Vedsted resulting in higher metalimnetic  $\alpha$   
413 values further supported these results (data not shown). There are two explanations for an increase in  $\alpha$  in  
414 metalimnion. Either it can be due to physiological acclimation in the phytoplankton (Grobbelaar et al. 1995;  
415 MacIntyre et al. 2002; Deblois et al. 2013) or it can be due to an increase in biomass of phytoplankton in this  
416 zone (Amand and Carpenter 1993; Hamilton et al. 2010). To explore if the higher metalimnetic  $\alpha$  values were  
417 dictated by higher levels of phytoplankton biomass with depth, we examined daily profiles of Chl *a*  
418 fluorescence measured in Vedsted Lake. During the stratified period we found no clear pattern with depth and  
419 no significant differences between the epi- and metalimnion ( $p > 0.05$ ; Wilcoxon paired test). Furthermore, bi-  
420 weekly measurements of Chl *a* above and below  $Z_{mix}$  in Hampen and Castle lakes (not shown), showed that for  
421 both lakes, Chl *a* below  $Z_{mix}$ , was less than 50% of epilimnetic values during summer stratification. Thus higher  
422 light utilization efficiency in the low light conditions of the metalimnion appears to be independent of biomass  
423 accumulation. From our study it therefore appears that elevated light utilization efficiency by phytoplankton is a  
424 common phenomenon occurring under low light conditions in the metalimnic zone, where nutrients are  
425 generally more replete (Fee 1976; Abbott et al. 1984). This is similar to experimental conditions where overall  
426 food chain efficiency of aquatic ecosystems has been found to be highest under low light conditions (Dickman  
427 et al. 2008).

428 Our study underlines the importance of measuring depth-specific metabolism in stratified lakes to  
429 reduce a bias towards water column net autotrophy when only surface processes are considered. In comparing  
430 lakes of different trophic status, we found that although most of GPP and R occurred in the epilimnion, up to

431 20% and 37% of GPP and R took place in the metalimnion, having  $NEP > 0$  when the depth of the photic zone  
432 exceeded the depth of the upper mixed layer. Areal metabolic estimates based on epilimnetic measurements  
433 only may accordingly deviate significantly from depth integrated estimates when  $Z_{eu} > Z_{mix}$ . While the observed  
434 importance of light and mixing regime for vertical patterns in primary production and respiration is convincing  
435 and consistent with experimentally based models, further research is needed to support the suggested increase  
436 in background respiration and *in situ* light utilization efficiency with depth.

437

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449

450 **References**

- 451 Abbott, M. R., K. Denman, T. M. Powell, P. Richerson, R. Richards, and C. Goldman. 1984. Mixing and the  
452 dynamics of the deep chlorophyll maximum in Lake Tahoe. *Limnol. Oceanogr.* **29**: 862-878.
- 453 Amand, A. S., and S. R. Carpenter. 1993. Metalimnetic phytoplankton dynamics, p. 210-224. *In* S. R. Carpenter  
454 and J. F. Kitchell [eds.], *The trophic cascade in lakes*. Cambridge Univ. Press.
- 455 Batt, R. D., and S. R. Carpenter. 2012. Free-water lake metabolism: Addressing noisy time series with a Kalman  
456 filter. *Limnol. Oceanogr.: Methods* **10**: 20-30, doi: 10.4319/lom.2012.10.20
- 457 Bella, D. A. 1970. Dissolved oxygen variations in stratified lakes. *Journal Water Pollution Control Federation*  
458 **96**: 1129-1146.
- 459 Caraco, N. F., and J. J. Cole. 2004. When terrestrial organic matter is sent down the river: Importance of  
460 allochthonous C inputs to the metabolism in lakes and rivers, p. 301-316. *In* G. A. Polis, M. E. Power and  
461 G. Huxel [eds.], *Food webs at the landscape level*. University of Chicago Press.
- 462 Coloso, J. J., J. J. Cole, P. C. Hanson, and M. L. Pace. 2008. Depth-integrated, continuous estimates of metabolism  
463 in a clear-water lake. *Can. J. Fish. Aquat. Sci.* **65**: 712-722.
- 464 Coloso J. J., J. J. Cole, and M. L. Pace. 2011. Short-term variation in thermal stratification complicates estimation  
465 of lake metabolism. *Aquat. Sci.* **73**: 305–315.
- 466 Deblois, C. P., A. Marchand, and P. Juneau 2013. Comparison of photoacclimation in twelve freshwater  
467 photoautotrophs (Chlorophyte, Bacillariophyte, Cryptophyte and Cyanophyte) isolated from a natural  
468 community. *Plos One* **8**: 1-14.
- 469 del Giorgio, P. A., and P. J. B. Williams 2005. *Respiration in aquatic ecosystems*. Oxford University Press.
- 470 Dickman, E. M., J. M. Newell, M. J. González, and M. J. Vanni. 2008. Light, nutrients, and food-chain length  
471 constrain planktonic energy transfer efficiency across multiple trophic levels. *Proc. Natl. Acad. Sci.* **105**:  
472 18408–18412.
- 473 Domingues, R. B., A. B. Barbosa, U. Sommer, and H. M. Galvao 2011. Environmental drivers of phytoplankton  
474 in a turbid estuary: nutrient vs. light limitation. *European Journal of Phycology* **46**: 165-166.
- 475 Eaton, A. D., L. S. Clesceri, A. E. Greenberg, and M. A. H. Franson. [eds.]. 1995. *Standard methods for  
476 examination of water and waste water (19<sup>th</sup> edition)*. American Public Health Association.
- 477 Eckert, W., J. Imberger, and A. Saggio. 2002. Biogeochemical response to physical forcing in the water column  
478 of a warm monomictic lake. *Biogeochemistry* **61**: 291-307.
- 479 Fee, E. 1976. The vertical and seasonal distribution of chlorophyll in lakes of the Experimental Lakes Area,  
480 northwestern Ontario: Implications for primary production estimates. *Limnol Oceanogr* **21**: 767-783.
- 481 Gelda, R. K., and S. W. Effler 2002. Metabolic rate estimates for a eutrophic lake from diel dissolved oxygen  
482 signals. *Hydrobiologia* **485**: 51-66.
- 483 Grobbelaar, J. U. 1985. Phytoplankton productivity in turbid waters. *J. Plankton Res.* **7**: 653-663.
- 484 Grobbelaar, J. U. 1990. Modelling phytoplankton productivity in turbid waters with small euphotic to mixing  
485 depth ratios. *J. Plankton Res.* **12**: 923-931.
- 486 Grobbelaar, J. U., L. Nedbal, L. Tichy, and L. Setlik. 1995. Variation in some photosynthetic characteristics of  
487 microalgae cultured in outdoor thin-layered sloping reactors. *J. Appl. Phycol.* **7**: 175-184.
- 488 Hanson, P. C., D. L. Bade, and S. R. Carpenter. 2003. Lake metabolism: Relationships with dissolved organic  
489 carbon and phosphorus. *Limnol. Oceanogr.* **48**: 1112-1119.

- 490 Hanson, P. C., S. R. Carpenter, N. Kimura, C. Wu, S. P. Cornelius, and T. K. Kratz 2008. Evaluation of  
491 metabolism models for free-water dissolved oxygen methods in lakes. *Limnol. Oceanogr.: Methods* **6**:  
492 454-465.
- 493 Hamilton, D. P., K. R. O'Brien, M. A. Burford, J. D. Brookes, and C. G. McBride. 2010. Vertical distributions of  
494 chlorophyll in deep, warm monomictic lakes. *Aquatic Sciences* **72**: 295-307.
- 495 Hondzo, M., and H. G. Stefan. 1993. Lake water temperature simulation model. *J. Hydraul. Eng. ASCE* **119**:  
496 1251-1273.
- 497 Hornberger, G. M., and M. G. Kelly. 1974. A new method for estimating productivity in standing waters using  
498 free oxygen measurements. *Water Resour. Bull.* **10**: 265-271.
- 499 Imberger, J. 1985a. The diurnal mixed layer. *Limnol. Oceanogr.* **30**: 737- 770.
- 500 Imberger, J. 1985b. Thermal characteristics of standing waters: an illustration of dynamic processes  
501 *Hydrobiologia* **125**: 7- 29.
- 502 Jassby, A., and T. Platt. 1976. Mathematical formulation of the relationship between photosynthesis and light for  
503 phytoplankton. *Limnol. Oceanogr.* **21**: 540-547.
- 504 Jespersen, A. M., and K. Christoffersen. 1987. Measurements of chlorophyll *a* from phytoplankton using ethanol  
505 as extraction solvent. *Arch. Hydrobiol.* **109**: 445-454.
- 506 Jorgensen, S. E., and G. Bendoricchio [eds.]. 2001. Fundamentals of ecological modelling, 3<sup>rd</sup> ed. Developments  
507 in environmental modelling, volume 21. Elsevier.
- 508 Kishino, M., N. Okami, M. Takahashi, and S. Ichimura 1986. Light utilization efficiency and quantum yield of  
509 phytoplankton in a thermally stratified sea. *Limnol. Oceanogr.* **31**: 557-566.
- 510 Kragh, T., and M. Sondergaard. 2004. Production and bioavailability of autochthonous dissolved organic carbon:  
511 Effects of mesozooplankton. *Aquat. Microb. Ecol.* **36**: 61-72.
- 512 Laas, A., P. Nõges, T. Kõiv, and T. Nõges. 2012. High-frequency metabolism study in a large and shallow  
513 temperate lake reveals seasonal switching between net autotrophy and net heterotrophy. *Hydrobiologia*  
514 **694**: 57-74.
- 515 MacIntyre, S. 1993. Vertical mixing in a shallow, eutrophic lake: Possible consequences for the light climate of  
516 phytoplankton. *Limnol Oceanogr.* **38**: 798-817
- 517 MacIntyre, S., K. M. Flynn, R. Jellison, and J. R. Romero 1999. Boundary mixing and nutrient fluxes in Mono  
518 Lake, California. *Limnol. Oceanogr.* **44**: 512-529.
- 519 MacIntyre, H. L., T. M. Kana, T. Anning, and R. J. Geider. 2002. Photoacclimation of photosynthesis irradiance  
520 response curves and photosynthetic pigments in microalgae and cyanobacteria. *J. Phycol.* **38**: 17-38.
- 521 MacIntyre, S., A. Jonsson, M. Jansson, J. Aberg, D. E. Turney, and S. D. Miller. 2010. Buoyancy flux, turbulence,  
522 and the gas transfer coefficient in a stratified lake. *Geophys. Res. Lett.* **37**: L24604,  
523 doi:10.1029/2010gl044164
- 524 Melack, J. M. 1982. Photosynthetic activity and respiration in an equatorial African soda lake. *Freshwater Biol.*  
525 **12**: 381-399.
- 526 Mellard, J., K. Yoshiyama, C. A. Klausmeier, and E. Litchman. 2012. Experimental test of phytoplankton  
527 competition for nutrients and light in poorly mixed water columns. *Ecol. Monogr.* **82**: 239-256.
- 528 Monismith, S. 1986. An experimental study of the upwelling response of stratified reservoirs to surface shear-  
529 stresses. *J. Fluid Mech.* **171**: 407-439.
- 530 Mouget, J. L., G. Tremblin, A. Morant-Manceau, M. Morançais, and J. M. Robert. 1999. Long-term  
531 photoacclimation of *Haslea ostrearia* (Bacillariophyta): Effect of irradiance on growth rates, pigment  
532 content and photosynthesis. *Eur. J. Phycol.* **34**: 109-115.

- 533 Read, J. S., D. P. Hamilton, I. D. Jones, K. Muraoka, L. A. Winslow, R. Kroiss, C. H. Wu, and E. Gaiser 2011.  
534 Derivation of lake mixing and stratification indices from high-resolution lake buoy data. *Environmental*  
535 *Modelling & Software* **26**: 1325-1336.
- 536 Rimmer, A., M. Boger, Y. Aota, and M. Kumagai. 2006. A lake as a natural integrator of linear processes:  
537 Application to Lake Kinneret (Israel) and Lake Biwa (Japan). *J. Hydrol.* **319**: 163-175.
- 538 Sadro, S., J. M. Melack, and S. MacIntyre 2011*a*. Depth-integrated estimates of ecosystem metabolism in a high-  
539 elevation lake (Emerald Lake, Sierra Nevada, California). *Limnol. Oceanogr.* **56**: 1764-1780.
- 540 Sadro, S., C. E. Nelson, and J. M. Melack 2011*b*. Linking diel patterns in community respiration to  
541 bacterioplankton in an oligotrophic high-elevation lake. *Limnol. Oceanogr.* **56**: 540-550.
- 542 Salonen, K., K. Kononen, and L. Arvola. 1983. Respiration of plankton in two small, polyhumic lakes.  
543 *Hydrobiologia* **101**: 65-70.
- 544 Sand-Jensen K., and P. A. Staehr. 2009. Net heterotrophy in small Danish lakes: A widespread feature over  
545 gradients in trophic status and land cover. *Ecosystems* **12**: 336-34.
- 546 Solomon, C. T., D. A. Brusewitz, D. C. Richardson, K.C. Rose, M. C. Van de Bogert, P. C. Hanson, T. K. Kratz,  
547 B. Largert, R. Adrian, B. L. Babin, C. -Y. Chiu, D. Hamilton, E. E. Gaiser, S. Hendricks, V. Istvánovics,  
548 D. M. O'Donnell, M. L. Pace, E. Ryder, P. A. Staehr, T. Torgersen, M. J. Vanni, K. C. Weathers, and G.  
549 Zhu. 2013. Ecosystem respiration: Drivers of daily variability and background respiration in lakes around  
550 the globe. *Limnol. Oceanogr.* **58**: 849-866.
- 551 Staehr, P. A., D. Bade, M. C. Van de Bogert, G. R. Koch, C. E. Williamson, P. C. Hanson, J. J. Cole, and T.  
552 Kratz. 2010*a*. Lake metabolism and the diel oxygen technique: State of the science. *Limnol. Oceanogr.:*  
553 *Methods* **1**: 1-2.
- 554 Staehr, P. A., K. Sand-Jensen, A. L. Raun, B. Nielsson, and J. Kidmose. 2010*b*. Drivers of metabolism and net  
555 heterotrophy in contrasting lakes. *Limnol. Oceanogr.* **55**: 817-830.
- 556 Staehr, P. A., and K. Sand-Jensen. 2006. Seasonal changes in temperature and nutrient control of photosynthesis,  
557 respiration and growth of natural phytoplankton communities. *Freshwater Biol.* **51**: 249-262.
- 558 Staehr, P. A., and K. Sand-Jensen. 2007. Temporal dynamics and regulation of lake metabolism. *Limnol.*  
559 *Oceanogr.* **52**: 108-120.
- 560 Staehr P. A., J. Testa, M. Kemp, J. J. Cole, K. Sand-Jensen, and S. V. Smith. 2012*a*. The metabolism of aquatic  
561 ecosystems: History, applications, and future challenges. *Aquat. Sci.* **74**: 15-29.
- 562 Staehr, P. A., J. P. A. Christensen, R. D. Batt, and J. S. Read. 2012*b*. Ecosystem metabolism in a stratified lake.  
563 *Limnol. Oceanogr.* **57**: 1317-1330.
- 564 Stefan, H. G., X. Fang, D. Wright, J. G. Eaton, and J. H. McCormick. 1995. Simulation of dissolved-oxygen  
565 profiles in a transparent, dimictic lake. *Limnol. Oceanogr.* **40**: 105-118.
- 566 Van de Bogert, M. C., S. R. Carpenter, J. J. Cole, and M. L. Pace. 2007. Assessing pelagic and benthic metabolism  
567 using free water measurements. *Limnol. Oceanogr.:* *Methods* **5**: 145-155.
- 568 Yeates, P. S., and J. Imberger. 2004. Pseudo two-dimensional simulations of internal and boundary fluxes in  
569 stratified lakes and reservoirs. *Internat. J. River Basin Management.* **1**: 1-23.  
570

## TABLES

**Table 1.** Morphometrical and limnological descriptors of the three studied lakes (mean  $\pm$  SD during the studied period).

	Hampen	Vedsted	Castle
Mean depth (m)	4.2	5.5	3.7
Maximum depth (m)	13	11	9
Volume ( $10^6$ m <sup>3</sup> )	3.688	0.460	0.827
Area (km <sup>2</sup> )	0.76	0.08	0.22
Water residence time (y)	1.4	4.6	0.5
Chl ( $\mu\text{g L}^{-1}$ )	5.3 $\pm$ 4.3	41 $\pm$ 21	65 $\pm$ 67
TP ( $\mu\text{g L}^{-1}$ )	22.7 $\pm$ 5.8	19.5 $\pm$ 25.6	102 $\pm$ 34
K <sub>d</sub> (m <sup>-1</sup> )	0.7 $\pm$ 0.1	0.8 $\pm$ 0.2	1.7 $\pm$ 0.7
DOC (mg L <sup>-1</sup> )	2.97 $\pm$ 0.08	4.79 $\pm$ 0.56	3.46 $\pm$ 0.29
Studied period	30 May 2007 to 25 Aug 2007	01 Jun 2008 to 03 Aug 2008	08 Sep 2006 to 13 Oct 2006
Number of days	46 days	63 days	33 days
Sampled depths	1, 3, 5, 7, 9	0.5, 1, 2, 3, 4, 5, 6, 7, 8, 9	0.5, 2, 4, 6, 8
Z <sub>mix</sub> (m)	4.7 $\pm$ 1.7	3.8 $\pm$ 0.4	5.1 $\pm$ 1.0
Z <sub>eu</sub> (m)	6.3 $\pm$ 0.9	4.0 $\pm$ 0.7	2.2 $\pm$ 0.6
Metalimnetic width (m)	1.5 $\pm$ 0.7	1.8 $\pm$ 0.2	1.2 $\pm$ 0.4
Fraction of water volume in the epilimnion (%)	73 $\pm$ 13	55 $\pm$ 5	91 $\pm$ 5

**Table 2.** Relative contribution (%) of epilimnion (epi), metalimnion (meta), and hypolimnion (hypo) to total areal metabolism of the water column (mean  $\pm$  SD for GPP and R).

		Hampen	Vedsted	Castle
GPP	epi	79 $\pm$ 17	83 $\pm$ 15	100 $\pm$ 0
	meta	20 $\pm$ 16	17 $\pm$ 15	0 $\pm$ 0
	hypo	1.1 $\pm$ 1.6	0.2 $\pm$ 0.6	0 $\pm$ 0
R	epi	69 $\pm$ 19	54 $\pm$ 17	96 $\pm$ 4
	meta	26 $\pm$ 18	37 $\pm$ 15	4 $\pm$ 4
	hypo	5 $\pm$ 4	9 $\pm$ 7	0.1 $\pm$ 0.1

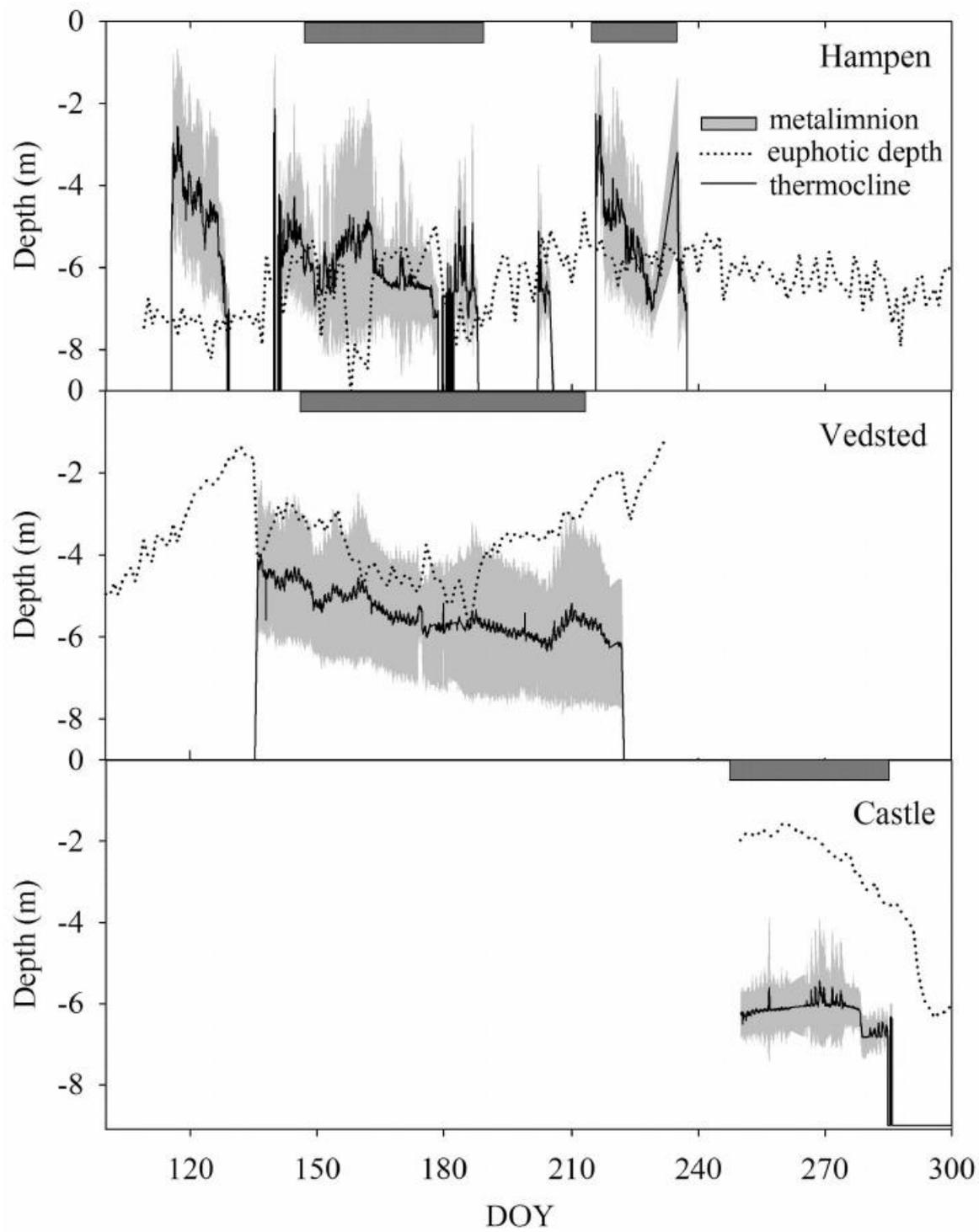
**Table 3.** Summary statistics (slope, intercept, and  $r^2$  of the equation  $R = \beta + aGPP$ ) of the GPP vs. R relationships in the volumetric rates of epi-, meta-, and hypolimnion and in the total areal rates, for the three studied lakes. Note that  $\beta$  is the background respiration, expressed in  $\text{mmol m}^{-3} \text{d}^{-1}$ . Only the significant parameters of a Type II regression are shown ( $p < 0.05$ ). ns: not significant. Data are means  $\pm$  standard error (SE)

	$r^2$	slope	$B$
<b>Hampen</b>			
epi	0.39	0.57 $\pm$ 0.10	7.9 $\pm$ 4.61
meta	-	ns	38.7 $\pm$ 5.63
hypo	-	ns	30.3 $\pm$ 3.47
whole column	0.49	0.68 $\pm$ 0.10	6.1 $\pm$ 3.9
<b>Vedsted</b>			
epi	0.19	0.39 $\pm$ 0.11	17.4 $\pm$ 3.7
meta	0.09	0.52 $\pm$ 0.19	31.3 $\pm$ 2.6
hypo	0.06	-3.66 $\pm$ 1.7	23.6 $\pm$ 1.9
whole column	0.21	0.45 $\pm$ 0.11	21.5 $\pm$ 2.7
<b>Castle</b>			
epi	-	ns	43.8 $\pm$ 5.6
meta	-	ns	23.3 $\pm$ 3.1
hypo	-	ns	6.8 $\pm$ 2.1
whole column	-	ns	26.6 $\pm$ 3.5

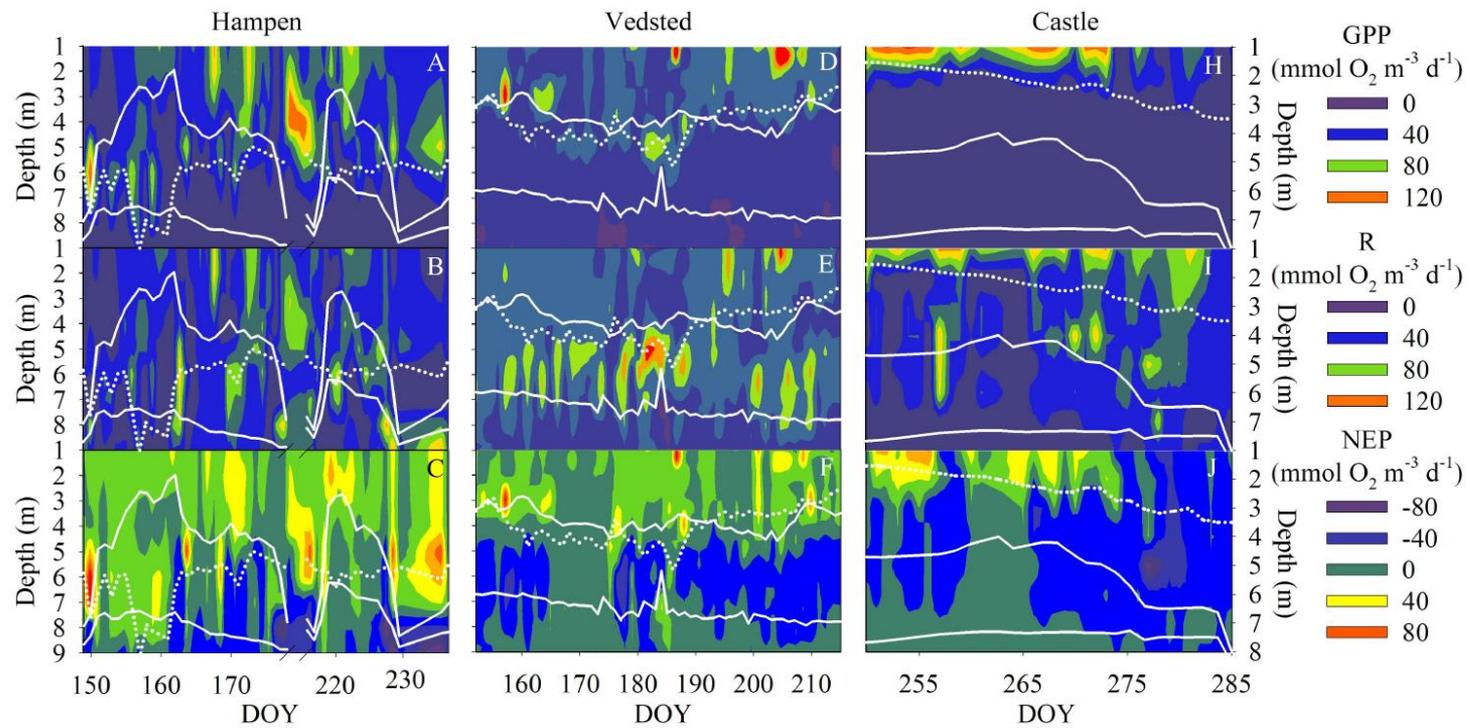
**Table 4.** Correlation (Pearson's  $r$ ) of the Jassby and Platt model relating mean daily volumetric NEP rates in epi-, meta-, and hypolimnion and the corresponding mean daily PAR in each layer. ns: not significant ( $p < 0.05$ )

	Hampen	Vedsted	Castle	all lakes pooled
epi	0.55	0.28	0.44	0.59
meta	0.52	0.30	ns	0.43
hypo	0.53	ns	ns	Ns

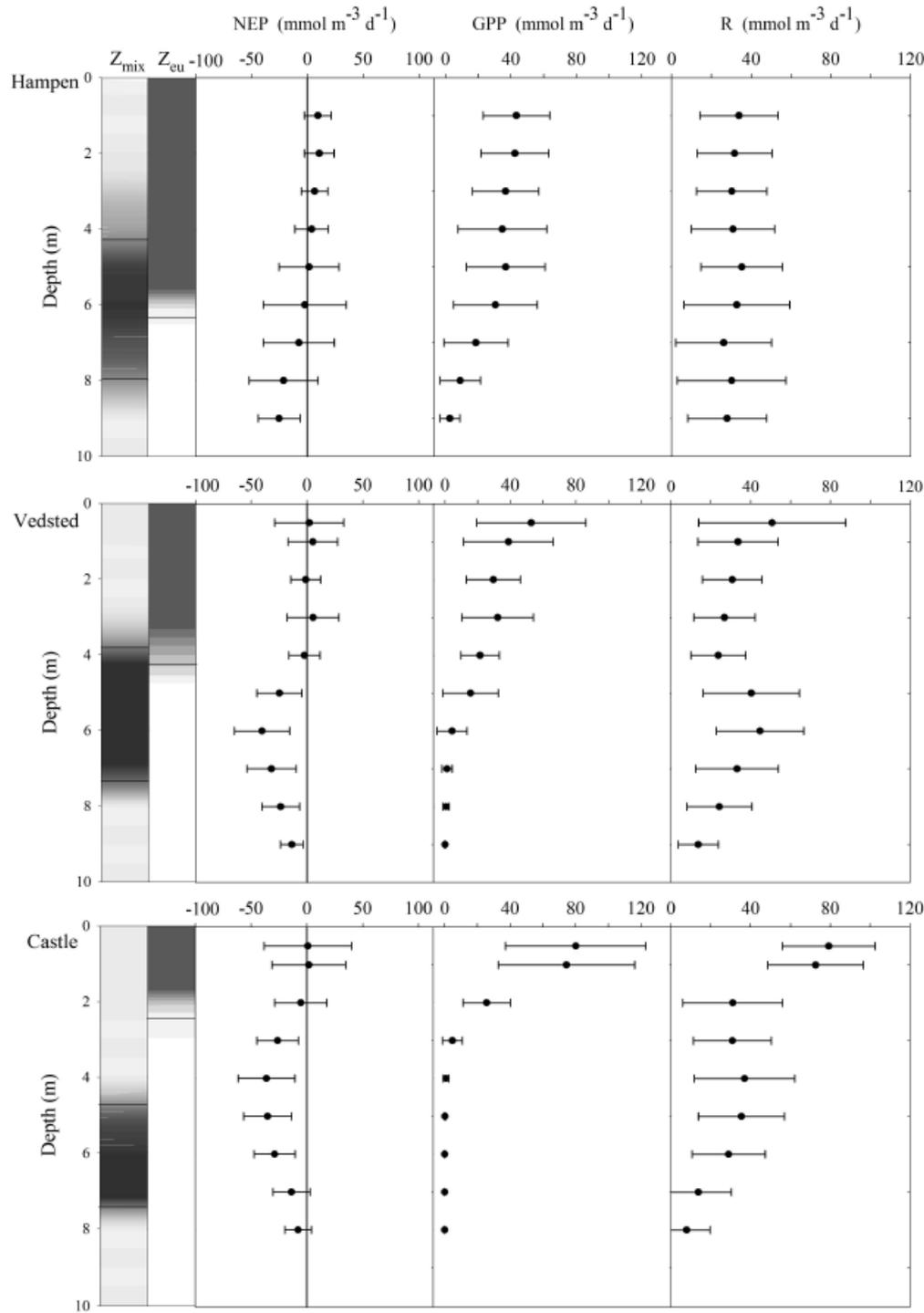
## FIGURES



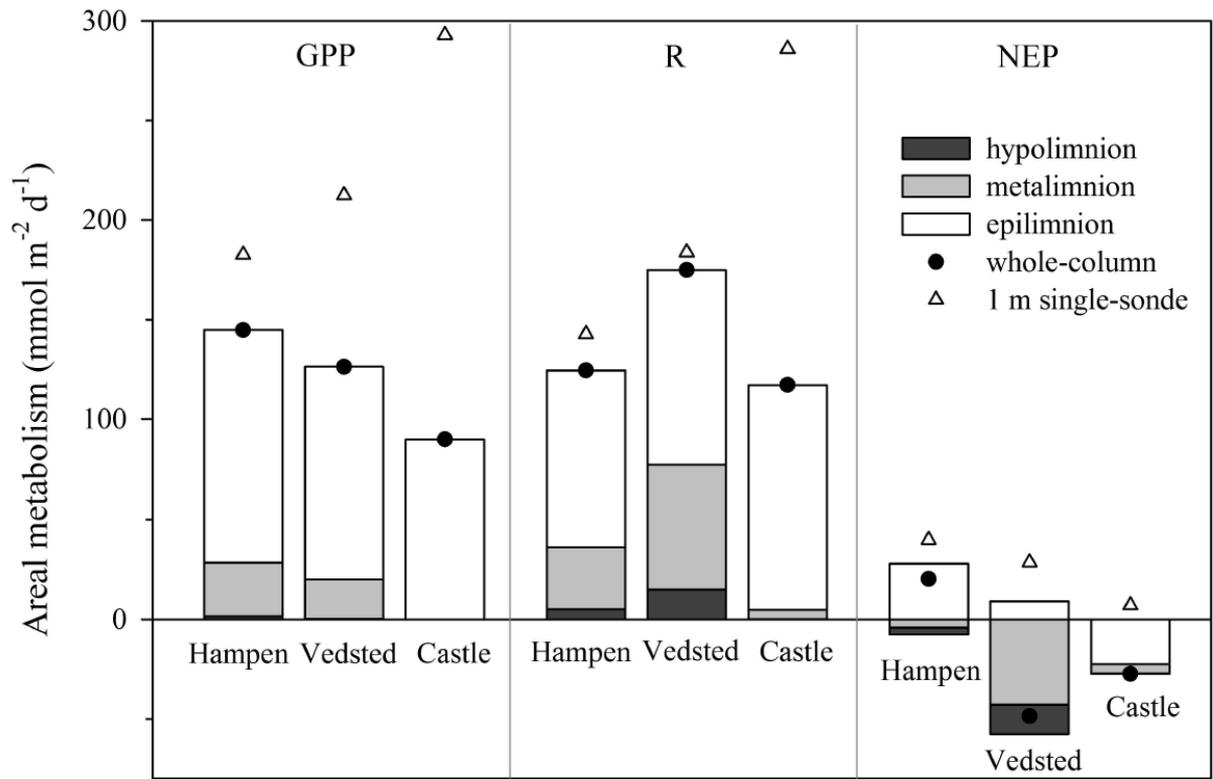
**Figure 1.** Seasonal variation in the physical structure of the water column and the depth of the photic zone ( $Z_{eu}$ , 1% of surface irradiance). Horizontal bars identify the studied periods in each lake. DOY – day of the year.



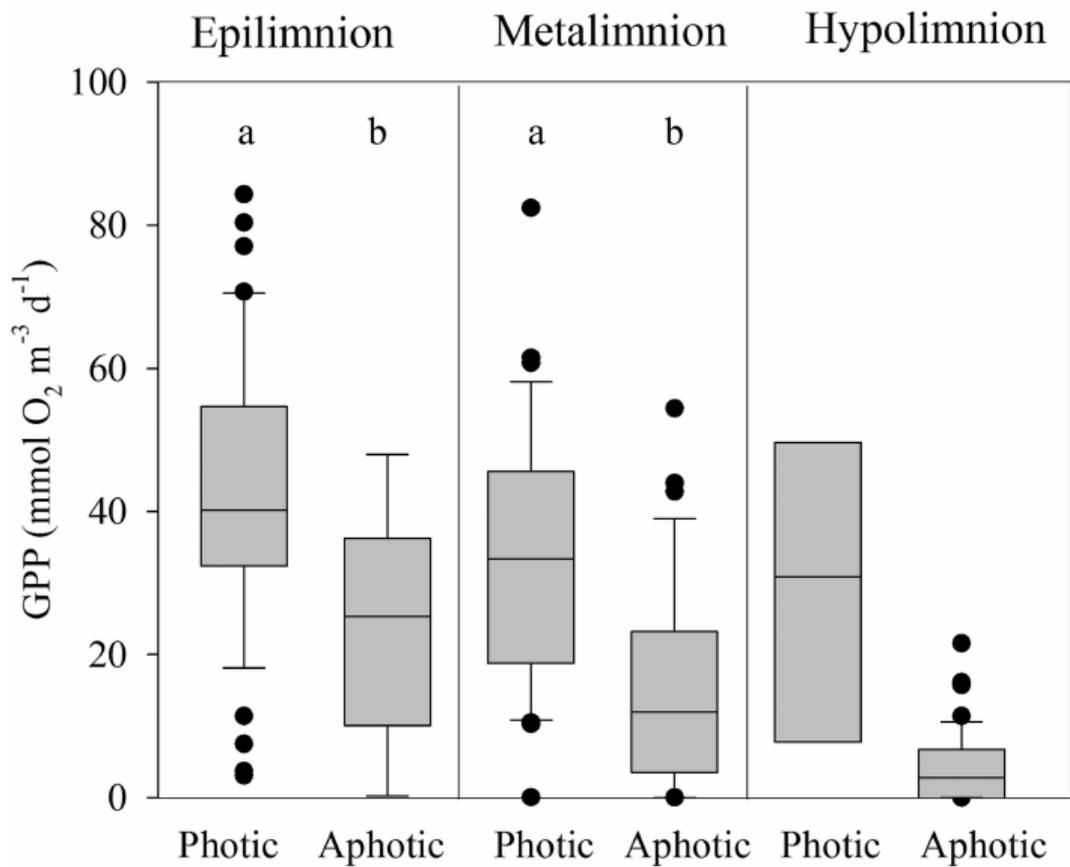
**Figure 2.** Depth specific daily rates ( $\text{mmol O}_2 \text{ m}^{-3} \text{ d}^{-1}$ ) of gross primary production (GPP), respiration (R), and net ecosystem production (NEP) in (A-C) Hampen, (D-F) Vedsted, and (H-J) Castle Lake. The upper and lower limits of the metalimnetic zone are shown as solid white lines and the depth of the photic zone is shown as a dotted line. Time is shown as day of year (DOY) which is not the same for the three lakes (*see* Table 1).



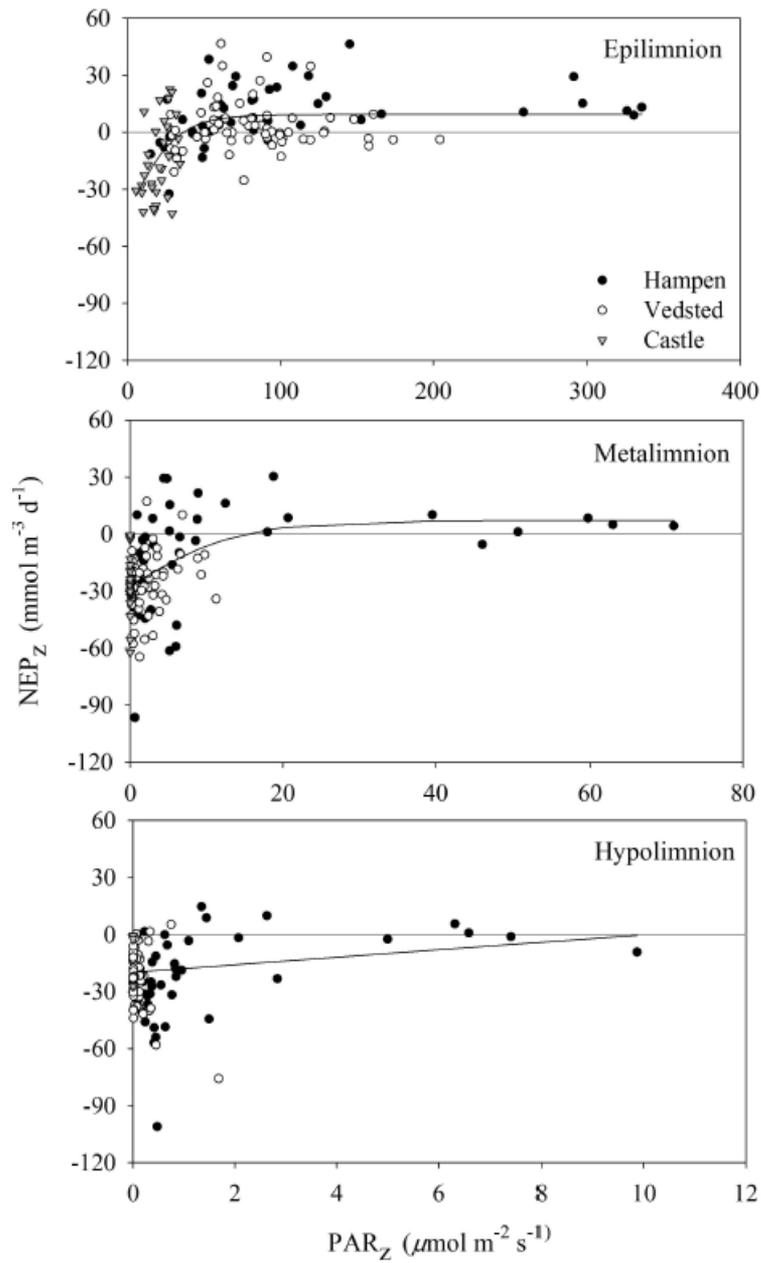
**Figure 3.** Depth profiles of specific NEP, GPP, and R ( $\text{mmol m}^{-3} \text{d}^{-1}$ ) for the three studied lakes (mean  $\pm$  SD). The depths of the metalimnion and of the photic zone are shown on the left boxes through a shading gradation (from black to white) that represents the frequency of a given depth of being within the metalimnion or the photic zone, respectively (the horizontal lines there correspond to the 90% confidence interval).



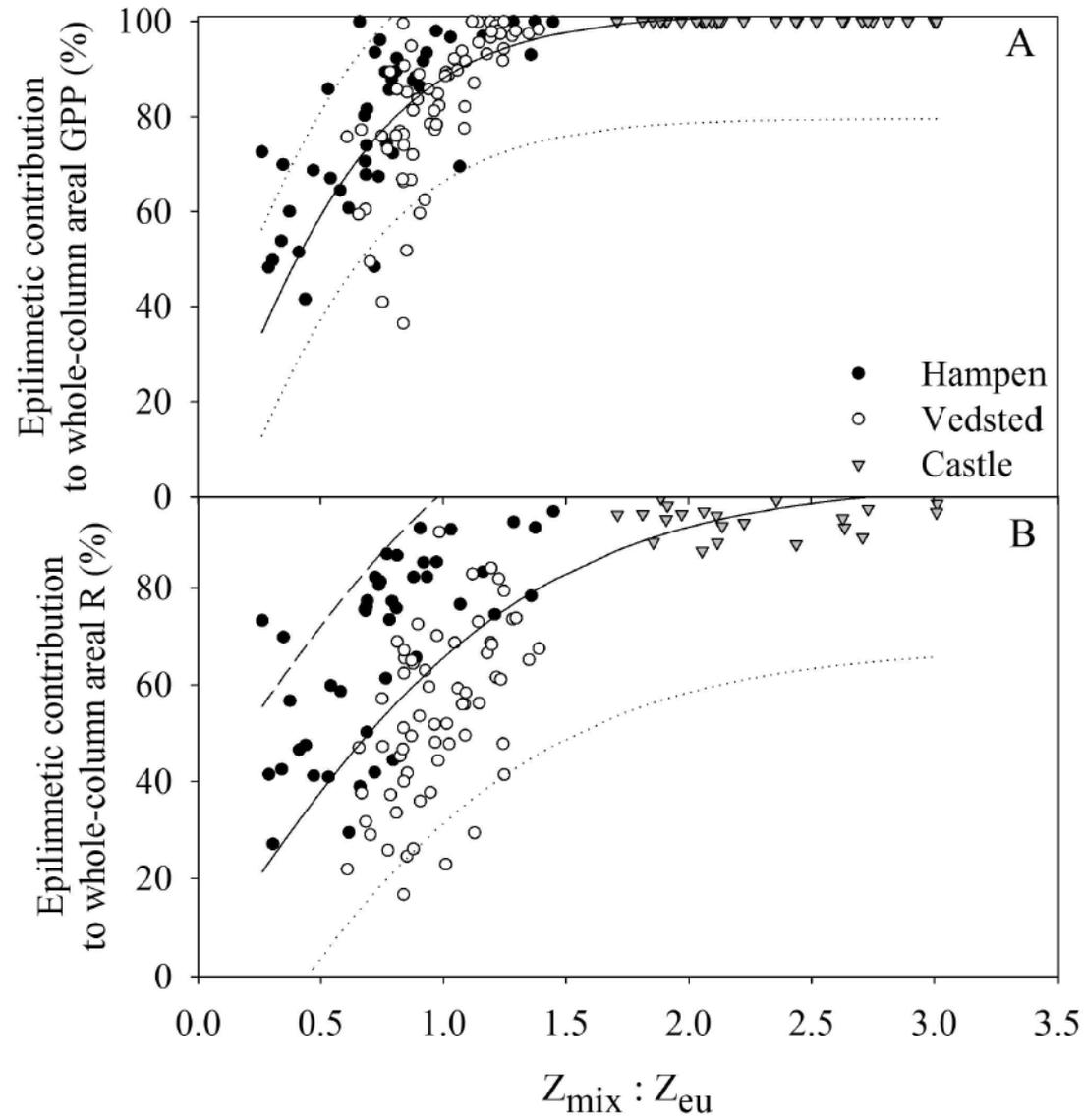
**Figure 4.** Mean areal rates (GPP, R, and NEP, in  $\text{mmol m}^{-2} \text{d}^{-1}$ ) of the depth layers (bars) and integrated whole-column rates (circles) for the three studied lakes. The estimates derived from single-sonde (1 m depth) measurements are also shown (triangles).



**Figure 5.** Median and 25th, 75th percentile boxes with 10th, 90th percentile whiskers of GPP ( $\text{mmol O}_2 \text{ m}^{-3} \text{ d}^{-1}$ ) for the epi-, meta, and hypolimnion of Hampen Lake depending on whether the layers are or not in the photic zone. Dots are outliers and the lower case letters denotes the different groups according to Tukey post test. Hypolimnion data was left out of the analysis as they did not pass the test of variance homogeneity.



**Figure 6.** Relationship between daily specific NEP ( $\text{mmol m}^{-3} \text{d}^{-1}$ ) and mean daily available PAR in each depth layer. The line corresponds to the fitted model at a daily scale, pooling the three lakes (*see* Table 4).



**Figure 7.** Relative contribution of epilimnetic metabolism to whole-column areal (A) GPP and (B) R in relation with the ratio  $Z_{\text{mix}} : Z_{\text{eu}}$ . The lines correspond to a fitted hyperbolic function with 95% confidence intervals (dotted).