Chapter 3

MEDITERRANEAN CLIMATE RIVERS OVER THE WORLD: convergences and divergences between regions¹.

INTRODUCTION

Mediterranean regions and climate

There are five regions over the world present a mediterranean climate (called med-regions). They are located between 32°-40° N and S of the Equator, in the west side of continents (Aschmann, 1973a): Mediterranean Basin, California, Chile, South Africa and Australia (the South-West and South). All these areas represent between 1% and 4% (di Castri, 1981; Mooney, 1982) of the earth surface (Figure 1). However, there are discrepancies among geographers, physics or biologists in the description and delimitation of extension of mediterranean climates (Nahal, 1981).

The mediterranean climate is defined in terms of precipitation (di Castri, 1973a) and temperature (Aschmann, 1973a), and it is characterized by a high seasonality (Paskoff, 1973; Daget & Michel-Villag, 1975) summarized as hot and dry summers and cool and wet winters (Köppen, 1923). Winter temperatures are mild (7-13°C) with frosts and snow infrequent,

 $^{^1}$ Mediterranean is written with "M" when is referred to the Mediterranean Basin, and with "m" when it speaks of the climate.

whereas summers are hot with a mean temperature of 14-25°C (Paskoff, 1973). On the other hand, precipitation it is highly variable along and between years (Paskoff, 1973; Nahal, 1981), implying a high climatic heterogeneity either along time and space (Mount, 1995). The annual range of precipitation goes from 275 to 900mm/y, although several authors define the clime only using the minimum (see Conacher, 1995; Conacher & Sala, 1998). Most of rainfall falls during winter months, and this is one of the key elements to distinguish between mediterranean, temperate and subtropical climates (Dept. Land Affairs, 1994). According to Aschmann (1973a) more than 65% of annual rainfall should fall in winter, although in some areas this is not as strict (Köppen, 1923; Specht, 1979) with important spring and autumn rains (Southern California and Southern Europe —Miller, 1983). During summer, a drought period is present with length and intensity variable between years (di Castri, 1973a, 1981).



Figure 1. Map of the distribution of mediterranean climate over the world.

Although there are many common climatic characteristics among med-regions, local climate is modified by orography, altitude, orientation from coast line, continent and oceanic influences... (Aschmann, 1973a; Conacher & Conacher, 1998), and therefore several microclimatic patterns can be established (Nahal, 1981; Daget *et al.*, 1988). Köppen (1923) distinguishes between hot and cold summers, and Emberger (1930, 1955, and 1971) defines 6 types of subclimates according to its humidity and the winter severity: subdesertic or periarid, xeromediterranean or arid, termomediterranean or semiarid, mesomediterranean or

subhumid, submediterranean or humid and subxeric or perhumid. However, from these subclimates Di Castri (1973c) distinguished an "eumediterranean" subclimate which would be equivalent to the semiarid and subhumid subclimates, with a precipitation range of 275-640mm/y.

Mediterranean climate is located between the temperate and dry climate areas (see Dallman, 1998), and both regions have influenced strongly the mediterranean climate and its present biota (di Castri, 1981; Herrera, 1995). Relationships with these adjacent climates are variable between regions (di Castri, 1981). For example, summer precipitation is higher in South Africa and northern Mediterranean Basin because of their proximity to tropical and temperate regions, and lower in Chile because of Atacama Desert influences (di Castri, 1981).

Mediterranean biota and humans

Mediterranean climate originated in the Pliocene, 3.2 My ago (Alxelrod, 1973; Suc, 1984), and therefore is a young climate in geological terms, younger than part of the biota found in these med-regions (Raven, 1973). As a consequence, some lower vertebrate, invertebrate and plant distributions are homogeneous, reflecting the patterns established during the Mesozoic, before the breakup of the continents and the formation of the mediterranean climate (Deacon, 1983; Herrera, 1995), or the later connections between regions (Nilsen, 1978; Cooke, 1972). Once the climate was formed, similar selection forces acted in the configuration of the mediterranean biota (Johnson, 1973), and therefore, a high similarity between regions should be expected (Mooney, 1982). However, mediterranean areas in the world show an important variability in ecosystems, because of geographic, microclimatic, topographic, physiographic, litologic and historic aspects have contributed to the present faunal and floral distributions (di Castri, 1981). Several authors suggest 4 origins of the mediterranean biota (di Castri, 1991):

- Species developed *in situ* before the mediterranean climate formation.
- Species developed after the mediterranean climate was established.
- Species developed outside the mediterranean regions but that settled in these areas posteriorly.
- Invasive species from human impact.

When biota is compared among med-regions, all these possible origins must be considered and specially the historical or environmental factors (di Castri & Hadley, 1985; di Castri, 1991). When first explorers arrived into the Cape region, central Chile, California and Australia, noted a high resemblance between these regions with the Mediterranean Basin (di Castri, 1981). However, the biogeographical concept of mediterranean biome was established a century later by Grisebach (1872), Drude (1890) and Schimper (1898), based on the similarities between vegetation in mediterranean areas. First comparative studies in these med-regions derived from botanical aspects (Spetch, 1979; Specht & Rayron, 1957; Specht, 1973; Cody & Mooney, 1978) revealing a high similarity in plant morphology and structure (Mooney and Dunn, 1970), with a sclerophyllous and evergreen vegetation (Kummerow, 1973; Mooney, 1977, 1982; Rundel, 1988). Duration of the summer drought and the cold in winter seem the ecological factors more likely to provide this similarity (Aschmann, 1973a; Nahal, 1981; Orshan, 1983; Miller, 1983). Both phenomena imply a continuous and predictable natural disturbance, where the evolution took place (Stanford & Ward, 1983), and developing plant communities with common characteristics with different names: chaparral in California, maquia or matorral in the Mediterranean Basin, matorral in Chile, fynbos in South Africa and health or mallee in Australia (Naveh & Whittaker, 1979). Some studies indicate that this type of vegetation and structure is not exclusive from the mediterranean region, with some extensions through areas with high summer rainfall in the east of Australia (Specht, 1979), Mexico (Muller, 1939) or east Africa (Rundel, 1988).

In spite of these similarities, some differences are present because of local factors as human impact, soil nutrients (Specht, 1979; Mooney, 1982; Specht & Moll, 1983), humidity (Beard, 1983), natural history (di Castri, 1973b) or landscape orography (Cody, 1973; Mooney, 1977; di Castri, 1981). However, as a general rule, mediterranean regions are richer in species that the adjacent ones (e.g., temperate or dry) (Raven 1973; Deacon, 1983), with a high endemism rate (Cowling, 1992) and very heterogeneous in space and time in terms of community composition and structure (di Castri, 1973, 1981).

Comparisons of floral and faunal communities between mediterranean regions are not easy because in some cases a high number of taxa with different ages of origin coexist (di Castri & Mooney, 1973). Despite of those difficulties, affinities in faunal communities between the med-regions have been described in numerous studies. Most of them focused on lizards (Sage, 1973; Fuentes, 1976), birds (Cody, 1973; Herrera, 1995) or terrestrial arthropods (di Castri & Mooney, 1973; Majer & Greenslade, 1988; Stamou, 1998), but few are performed in aquatic arthropods, although some suggestions have been made (Gasith & Resh, 1999).

In the mediterranean climate areas, the water high seasonal stress condition implies that organisms present some specific adaptations. Consequently, most of the affinities it can be found between these areas are related to physiological, morphological or behavioral adaptations (e.g. in artropods —Stamou, 1998) to avoid the severity of drought and its seasonality and interannual variability. In that sense, resilience and resistance are common attributes found in the elements of the mediterranean biota (Grubb & Hopkings, 1986), being the first more frequent (Fox & Fox, 1986). These adaptations give higher probability of permanence to mediterranean communities to the seasonal predictable natural disturbances (Aschmann, 1973; Orshan, 1983).

In addition to this harsh natural conditions, mediterranean climate areas have been exposed since many years ago to human disturbances (Mooney, 1982), although the human impact differs among med-regions because of the different time of human colonization (Aschmann, 1973b; Fox & Fox, 1986). Probably, fire is the main disturbance in these ecosystems (Miller, 1983; Keely, 1986), altering nutrient availability and influencing on speciation (Cowling, 1987) with fire-dependents and resilient plants (Kruger, 1979ac; Fox & Fox, 1986; Keely, 1986). For instance, in the South African fynbos several plants need the fire to reproduce; in Australia some tree are adapted to fire and its frequency and intensity; and even in Chile with a weak fire history, some adaptations to fire are present (Grove & Rackham, 2001).

Trabaud (1981) points out that med-regions are unique to have been affected by human activities for ages, as the introduction of non-native plants and animals, agriculture, cattle, urbanization, salinization... (Conacher & Sala, 1998). Because of the impact of all these activities, med-regions are characterized by lost of natural vegetation, soil salinization, water pollution and high erosion (Conacher & Sala, 1998). These consequences can be more or less important in each region, and topography, lithology or the climate itself can accelerate the erosion and the land degradation (see Conacher & Sala, 1998). Hence, an increase of temperature by climate change could have important consequences in the mediterranean ecosystems, enhancing the water stress in the arid and semiarid areas or prolonging the growth season in the alpine regions (Le Houérou, 1990).

Mediterranean rivers

Rivers influenced by this climatic heterogeneity in temperature and precipitation regimes are considered mediterranean rivers (Gasith & Resh, 1999). Because in the mediterranean climate several microclimates can be distinguished (Nahal, 1981; Daget *et al.*, 1988), we understand that several river typologies can also distinguish in med-regions: from mountain permanent

stream in the most humit areas to ephemeral ones in the most arid ones. In Plates 1 to 5 the main features of med-rivers in the five areas in the world are summarized.

River discharge is related to rainfall patterns (depending on the vegetal cover, temperature, evaporation, relief and soils) (Beckinsale, 1969; Alexander, 1985), and therefore, mediterranean rivers are characterized by yearly variable discharge regimes, with a maximum peak in winter a minimum flow in summer (Bradbury, 1981; Gasith & Resh, 1999). Moreover, an interannual variability is also present carrying out changes in river flow and freshwater communities along time (McElravy *et al.*, 1989; Resh *et al.*, 1990).

Frequently, rainfall in mediterranean rivers is presented as intense storms with a variable frequency, which implies frequent floods in the rainy season, especially in small and steep basins (Paskoff, 1973; Camarasa & Segura, 2001). On the other hand, in the dry season, because of the lack of precipitation and the high evapotranspiration, temporality in small streams can be common (Thrower & Bradbury, 1973; McElravy *et al.*, 1989; Conacher & Conacher, 1998; Gasith & Resh, 1999; del Rosario & Resh, 2000; Camarassa & Segura, 2001), except in cases when discharge is maintained by the influence of ground water (Vidal-Abarca *et al.*, 1996) o by snow-melt when streams are feed by high mountains (Beckinsale, 1969).

As a result of this discharge pattern, aquatic ecosystems are highly subject to biannual hydrologic disturbances, characterized by the alternation of floods and droughts (Molina *et al.*, 1994). This general pattern may be considered as predictable but with a variable frequency and intensity (Gasith & Resh, 1999), being droughts more predictable than floods (Vidal-Abarca, 2001). Permanent rivers may be found in areas with a relative high and predictable rainfall, whereas intermittent or ephemeral ones are located in areas with a low and uncertain rainfall (arid and semiarid zones) and without connection to main aquifers, being completely rainfall dependents (Vidal-Abarca, 1990). Permanency or temporality are functions of precipitation, microclimate patterns and freatic level, although other structural and local factors (e.g., substrate) can be important (see Bonada *et al.*, Chapter 5).

The effect of this variability in discharge over to biological community has been studied by several authors (see Gasith & Resh, 1999). In general, numerous adaptations to droughts (see Williams, 1985) and floods (Statzner *et al.*, 1988; Resh & Solem, 1996) have been found. In early summer, as long as the river is drying up and habitat change, there is a variation in composition and abundance in aquatic biota (Gasith & Resh, 1999). Later, autumn and winter floods connect the isolated reaches in the temporary rivers and provoke spates in temporal

and permanent streams washing out downstream algae and macroinvertebrate communities. After the flood, communities may be reestablished relatively fast (Badri *et al.*, 1987; Vidal-Abarca *et al.*, 1992; Pires *et al.*, 2000). According to the variability of the discharge along the year the rivers and streams can be classified into three groups (Bonada *et al.*, Chapter 5): permanents with running water in summer, intermittents with isolated pools in summer and ephemerals that are totally dried up after few months (even weeks) of the flow cessation.

This variety in characteristics in mediterranean rivers, provided by natural disturbances, is the origin of a high floristic and faunistic diversity (Prat *et al.*, 2000a) and endemism (Gasith & Resh, 1999). However, these ecosystems are very sensitive to a loss of diversity, because the seasonal variability of the mediterranean climate makes the fluvial systems more vulnerable to pollution (Gasith, 1992; Davies *et al.*, 1993; Sabater *et al.*, 1995; Gasith & Resh, 1999). Consequently, Mediterranean rivers are highly affected by human disturbances yielding a very poor ecological status (Prat, 1993; Prat & Munné, 2000). Moreover, in nowhere else in the world the aquatic fauna is declining as rapidly than in mediterranean regions (Moyle & Leidy, 1992) because of the non-native introductions, high population density, limited availability of water and unsatisfactory management.

Johnson (1973) postulated that although similar biota between isolated areas is possible because of a common ancestor, usually the environmental conditions have actuated more or less equally over the evolution of the biota, and so it is not likely that the convergence characteristics are reached by chance. Therefore, because the mediterranean climate shares similar kind of environments, macroinvertebrate communities should also be similar (Gasith & Resh, 1999), as the vegetation (Mooney and Dunn, 1970) or lizard communities are (Fuentes, 1976). Several ecological freshwater studies have been done comparing different biomes, climates or continents (Vinson & Hawkins, 1998) looking for convergences and divergences (Moyle & Herbold, 1987; Lamoroux *et al.*, 2002), but few of them are focused on comparisons between mediterranean rivers (see King *et al.*, 1988), although the numerous local studies performed in these areas (see Gasith & Resh, 1999). Studies about ecological convergences in mediterranean ecosystems are required to determine the applicability to another area of the ecological principles derived from local studies (Kruger *et al.*, 1983).

Because communities have different structure, composition and responses may change at different scales of observation (Kotliar & Wiens, 1990; Tonn *et al.*, 1990; Wiens *et al.*, 1993; Peckarsky *et al.*, 1997). Choosing the appropriate scale of study have been a problem for many stream ecologists (Peckarsky *et al.*, 1997) and thereby, several authors recommend performing

studies at multiple scales of observation to understand patterns and processes of biological communities (Peckarsky *et al.*, 1997; Poff, 1997) and facilitate convergences and divergences between regions (Tonn *et al.*, 1990; Samuels & Drake, 1997). Thus, the aim of this paper is to present convergences and divergences among mediterranean climate rivers over the world at two scales of observation: regional and local (including reach and habitat). The objectives of the study are:

1.- To compare faunistic composition of macroinvertebrates at family level between all medregions, in order to find if differences because of historical reasons may be important for ecological purposes.

2.- To study if the effect of temporality and habitat characteristics over the macroinvertebrate community are similar in four med-regions, despite of biogeographical differences.

3.- To establish convergences and divergences between the med-regions according to different aspects of the aquatic ecosystems.

METHODOLOGY

The faunistic assemblage in mediterranean rivers

In this Chapter, the assemblage of macroinvertebrates in a region is defined as the total taxa that could be found in any stream of the area at family level at any time (accumulated assemblage). Family level was preferred to simplify comparisons and to avoid data "noise" from genus or species level because of their high biogeographic variability. Moreover, in some areas the genus and species of some families (e.g. Chironomidae) are poorly known or difficult to identify. In some cases, as Oligochaeta and Hydracarina, the order level was used instead of family because taxonomic difficulties. The list of families recorded in each med-region has been obtained using personal data, references and non-published reports of freshwater researchers in each area. Chris Madden from the Australian Water Quality Center and Ricardo Figueroa from the Universidad de la Concepción provided unplublished data from South Australia and Chile, respectively.

Effect of the temporality and habitat on macroinvertebrates

Sites studied

Between 1999 and 2002 the former author made several studies in the mediterranean areas of California, South Africa, SWAustralia and the Mediterranean Basin. In all cases, the sampling period was spring, when the rivers were flowing to enable comparisons between areas.

	Med-Basin	California	South Africa	SW Australia
Total reference sites	42	10	12	9
Total R/L sites (reference)	8	10	12	9
Permanent reference sites	32	8	8	6
Temporary reference sites	10	2	4	3

Table 1. Number of sampling sites in med-regions.

As possible, many types of rivers in headwaters and downstream reaches were sampled, although sampling sites did not cover all the extension of each mediterranean region. Several permanent and temporary reference sites were selected and sampled in all four med-regions. Sites were identified as temporary if the river usually remains intermittent (disconnected pools) or dries up in summer or at least it was dry the summer before the study, whereas sites known to have continuous flow along year were qualified as permanents. Table 1 summarizes the number of sampling sites for each mediterranean sampling area.

Sites in Mediterranean Basin: Spain

Data from spring 1999 was obtained from the GUADALMED project (see Chapter 1). From the 157 sites sampled by the Project, 42 reference localities were used according to the criteria established by Bonada *et al.* (in press b). Sites are distributed along the mediterranean coast of Spain, including the Balearic Islands, and they belong to 11 basins: Besòs Llobregat, Mijares, Turia, Júcar, Segura, Aguas, Almanzora, Guadalfeo, Adra and Pollença (Figure 2). Information about the characteristics of these basins, typology of sites, riparian vegetation and habitat characteristics is found in Robles *et al.* (in press), Bonada *et al.* (in press a), Suárez *et al* (in press) and Pardo *et al* (in press). In general terms, sampling sites cover wide types of rivers, as was showed by Bonada *et al.* (in press) from small alpine rivers in Catalonia and Granada to karstic systems in Catalonia, Mallorca or Almería and ramblas in Murcia and Almería. Reference localities with disconnected pools in spring 1999 were non-considered in the analysis. Rainfall accumulated one year before sampling was 507.6 mm in an area near sampling sites (Vallès Oriental, Catalonia).

Sites in California

10 sampling sites in Northern California (Figure 2) were sampled in spring 2002, belonging to 3 of the 5 ecoregions established in California by Bailey (1995): 3 located in the west side of Sierra Nevada in the "Sierran steppe, mixed and coniferous forest"

ecoregion; 4 in "Northern California Coastal redwood forest and steppe" and 3 in "Southern California coastal chaparral woodland, coniferous forest and meadow". Localities in the Sierra Nevada are reaches with fast flowing waters, in forested basins, medium slopes and substrates with boulders and cobbles. The riparian vegetation is dominated by *Alnus* sp., *Salix* sp. (3 species), *Populus* sp. with some *Pinus lambertiana* and *Pseudotsuga menziesii* from the adjacent coniferous forest. The algae *Nostoc* sp. is abundant and frequently is present in symbiosis with Chironomidae larvae. In coastal ranges, rivers have high slopes with coarse substrate except in the longest rivers (as San Geronimo River) where gravels, sand and bedrock are important. Riparian vegetation is composed by *Quercus lobata, Alnus* sp., *Corylus cornuta, Sequoia sempervirens* and *Umbellularia californica*. Through the south and near the coast, sampled rivers are short, small and steep. Substrates are similar to the ones in the northern coasts. The riparian vegetation is composed by *Quercus lobata, Platanus racemosa, Juglans hindsii, Populus* sp., *Salix* sp., *Salix* sp., *Corylus cornuta* and *Umbellularia* californica. Rainfall accumulated one year before sampling was 572.64 mm in an area near sampling sites (Napa valley).

Sites in South Africa

South African survey was performed in spring 2000. Sampling sites were located in three different places (Figure 2): 3 in the Table Mountain ranges, 8 in the area of influence of the Franschhoek and Drakenstein Mountains, and 1 in the low Palmiet close to the sea, in the area called Kogelberg State Forest Reserve. Table Mountain sites have a high slope, with boulders and cobbles dominant, whereas low slopes characterize the rest of sites. In all cases, dominant vegetation is mountain fynbos, although near the river individuals of the riparian shrubs *Metrosideros angustifolia* or *Brabejum stellatifolium* are present. In headwaters, the macrophyte *Isolepes* sp. is abundant with mosses and Palmiet (*Prionium serranum*). Rainfall accumulated one year before sampling was 569.2 mm in an area near sampling sites (Cape Town).

Sites in SW Australia

A total of 9 sites in three catchments in the south of Perth were sampled in spring 1999: Serpentine River, Drakes Brook and Brunswick River (Figure 2). They are located on the Darling Plateau, where the karri forest is present providing a very forested basins by several *Eucaliptus* species and with a shrubby undergrowth. Near the river, individuals of *Melaeuca* sp. are present. Gravels, sand and silt, with bedrock and boulders scarce, dominate the substrate of the rivers. Rainfall accumulated one year before sampling was 774.6 mm in an area near sampling sites (Perth).

92



Figure 2. Sampling sites in California, Spain, South Africa and SWAustralia.

Sampling procedure

All reference sites were sampled using the Protocol 2 developed by the GUADALMED project (see Bonada *et al.*, Chapter 1; Jáimez-Cuéllar *et al.*, in press). This Spanish project attempts to assess the ecological status of the Mediterranean rivers according to the guidelines of the Water Frame Directive (WFD) (European Parliament and Council, 2000). The first stage of the

Project was to establish a standardized methodology to be used as a Rapid Bioassessment Protocol (RBP). The protocol was initially developed by Prat *et al.* (1997, 1999, 2000b), and has been tested and established later for all sampling sites in GUADALMED Project. Although GUADALMED Protocol was conceived as a bioassessment method, the fine mesh size used (250 μ m) and the absence of sampling restrictions comparing with other procedures, allow us the use of this Rapid Bioassessment Protocol (and specially Protocol 2) in macroinvertebrate community studies (Bonada *et al.*, Chapter 2).

In each site, several physical and chemical parameters were measured *in situ* as pH, oxygen (mg/l and %), temperature, conductivity and discharge. Two indexes developed in the GUADALMED Project were applied to check for the applicability of Spanish methodology in other mediterranean areas. The first one is the QBR, an index that to assess riparian vegetation quality (see Munné *et al.*, 1998, 2002; Suárez and Vidal-Abarca, 2000 and Suarez *et al.*, in press). The second one is the IHF designed to characterize habitat diversity (see Pardo *et al.*, in press), a method with some similarities with the RBP's method used in EEUU (see Barbour *et al.*, 1999).

For macroinvertebrates, the GUADALMED methodology consists in collecting samples separately from riffles (R) and lentic (L) habitats with a circular kick net of 250 μ m mesh size. Each sample is examined in the field and successive samples in both habitats are taken until no more families are found, to collect a representative sample of the community. Several invertebrates seen in the field but not taken in the sample were also recorded, as the large Heteroptera and Coleoptera. Samples were preserved in alcohol 70% and sorted in the lab, to avoid errors in identifying taxa. Macroinvertebrates were identified at family level (except Oligochaeta and Hydracarina) and a rank of abundances was recorded for each taxa: 1 from 1-3 individuals, 2 from 4-10, 3 from 11-100 and 4 for more than 100 individuals. Thus, data were semiquantitative and all the statistical analysis has been done according to it.

Data Analysis

Macroinvertebrate community assemblage in all med-regions

A Non-metric Multidimensional Scaling (NMDS) was applied to the presence/absence matrix with all taxa in med-regions. This ordination method preserves the distances between objects, plotting dissimilar objects far from the similar ones (Legendre & Legendre, 1998). On the other hand, the NMDS method is a method not based on eigenvalues, and the final axes are arbitrary without enclosing the explained variability. Because of the presence/absence data,

Jaccard coefficient was selected as a binary method to calculate distances between variables. This index is focused in checking similarities between variables, as double 0 are excluded. The statistical program used was the PCORD (McCune & Mefford, 1999).

Differences and similarities between all reference sampling sites in med-regions Physical, chemical and structural features

Once reference sites were established, a Principal Component Analysis (PCA) was performed to check for differences in environmental variables measured. Prior to analysis, these variables were standardized, tested for autocorrelation with r-Pearson. Skewness values for each variable were also calculated to avoid variables that could bias the analysis (Legendre & Legendre, 1998). PCA is a multivariate ordination analysis based on the Euclidean distance that let to study relationship between objects (sites) and descriptors (variables) in a reduced space. The CANOCO Program vs.4 (ter Braak, 1998) was used to compute the analysis.

Macroinvertebrate's abundances

Common taxa in all sampled regions were used to check for differences between the relative abundance between reference sites in the 4 sampled med-regions. Because in each med-region a different proportion of temporary sites respect permanents were sampled, only permanent sites were used to check similarities and differences in abundances. A non-parametric Kruskal-Wallis test was computed with a relative abundance (in %) data matrix and using sites in med-regions as replicates because its reference condition (Norris, 1995). Previously, data was tested to normality with the Shapiro-Wilk's test (that have a good power properties compared with a wide range of alternative tests), and to homogeneity of variances with the Barlett's test. The STATISTICA Program was used to perform the analysis (Stat Soft, 1999).

Effect of temporality

To test if differences between permanent and temporary sites were higher than between medregions, a Bray-Curtis cluster with flexible clustering (β =-0.25) was performed using all common taxa. For each med-region, macroinvertebrate relative abundance matrix in permanent and temporary sites was compared using a MRPP analysis (Multi-response Permutation Procedures). This method is a nonparametric method for testing multivariate differences among pre-defined groups (permanent or temporary sites), providing the statistic A and a p-value obtained by permutation (999 runs) as result. Because its non-parametric condition this method is more appropriated than MANOVA in comparisons of data matrixes that involve species abundances including many zero values.

Several metrics were tested to check for differences between flow conditions: number of taxa, relation EPT/OCH, number of Diptera taxa and number of non-insect taxa. Differences were tested with a Kruskal-Wallis test (after test for normality and homogeneity of variances). Further, an IndVal method (Dufrêne & Legendre, 1997) was performed to explain the meaning of the differences found between permanent and temporary sites and to determine the indicator community of each flow condition in med-regions. This method checks the indicator value for each taxon in the defined environmental conditions (permanent and temporary sites). A Monte Carlo permutation test with 9999 permutations was used to test for significance of each indicator value. This analysis has several advantages over the widely used TWINSPAN (Hill, 1979) as the possibility to obtain an indicator value for each taxon that is independent of abundances of other species (Dûfrene & Legendre, 1997). Only the taxa with more than an indicator value (IV-value) of 25 were kept (Dûfrene & Legendre, 1997). In order to find the degree of similarity between two med-regions, the percentage of common taxa in those two regions that have a similar pattern was calculated.

The statistical program PCORD (McCune & Mefford, 1999) was used to compute MRPP, IndVal and Bray-Curtis analysis whereas STATISTICA was employed for the non-parametric tests.

Effect of habitat

Similar analysis performed among riffle and pool habitats were applied to permanent and temporary sites. Thereby, a Bray-Cluster was also performed with common taxa, to test if biogeographical differences were more important than the habitat effect. Further, MRPP test, IndVal method and % of similarity between pair-regions were performed to check for differences among riffles and pools habitats and their indicator community in all mediterranean regions.

RESULTS

Macroinvertebrate community assemblage

A total of 212 different taxa were used to perform the analysis between med-regions. Most of taxa were at family level except Oligochaeta and Hydracarina that were used at a higher taxonomical level. Figure 3 shows the result from NMDS analysis. Macroinvertebrate

community is highly similar between the northern hemisphere med-regions of California and Med-Basin, with a 94% of similarity, while in the southern hemisphere there is an important variability in the macroinvertebrates community. South African community is the nearest to the northern hemisphere, sharing 65% of the taxa, while the Chilean area seems to be the less similar. The two Australian med-regions have 35% of similarity with the South African, California and Med-Basin communities, and 88% between them.



Figure 3. NMDS Plot using Jaccard distance.

The unique and common taxa are variable between regions (Table 2), being South Africa the region with a major number of unique taxa (13 families), followed by South Australia and Chile, both with 11 taxa. California and SWAustralia have few unique taxa, although some coldstenoterm plecopterans, as Peltoperlidae o Pteronarcyidae, are characteristic from North America. The med-region of SWAustralia has a scarce unique community, and only the caddisfly Plectotarsidae seems to be present in this area and absent in SAustralia. On the other hand, South Africa and Chile has important unique taxa from a Gondwanic origin, as Barbarochthonidae, Petrothrincidae Hydrosalpingidae and in South Africa and Austronemouridae or Diamphinoidae in Chile.

Table 2. Exclusive and ubiquitous taxa in the six med-regions.

Exclusive families		Ubiquitous families			
MEdBasin	SAustralia	Aeshnidae	Psychodidae		
Aphelocheiridae	Conoesucidae	Ancylidae	Simuliidae		
Beraeidae	Eusiridae	Baetidae	Sphaeriidae		
Bithvnellidae	Hemicorduliidae	Caenidae	Stratiomvidae		
Ferrissiidae	Nannochoristidae	Ceratopogonidae	Svrphidae		
Georrisidae	Ochteridae	Chironomidae	Tabanidae		
Limnebiidae	Paracalliopidae	Coenagrionidae	Tipulidae		
Neritidae	Scatopsidae	Corixidae	-		
Polymitarcidae	Syllidae	Culicidae			
Potamanthidae	Talitridae	Dixidae			
Prosopistomatidae	Tasimiidae	Dytiscidae			
-	Telephlebiidae	Elmidae			
	-	Empididae			
California	SWAustralia	Ephydridae			
Ameletidae	Plectrotarsidae	Gerridae			
Peltoperlidae		Gomphidae			
Pteronarcyidae		Gordiidae			
Neiridae		Gyrinidae			
		Haliplidae			
Chile	SouthAfrica	Hydracarina			
Aeglidae	Barbarochthonidae	Hydrophilidae			
Amelotopsidae	Chlorophydae	Hydropsychidae			
Amnicolidae	Dipseudopsidae	Hydroptilidae			
Austronemouridae	Hydrosalpingidae	Leptoceridae			
Chilinidae	Munnidae	Leptophlebiidae			
Coloburiscidae	Petrothrincidae	Lestidae			
Diamphipnoidae	Phreatoicidae	Libellulidae			
Eomeropidae	Pisuliidae	Lymnaeidae			
Mesobdellidae	Potamonautidae	Notonectidae			
Molanidae	Protojaniridae	Oligochaeta			
Samastacidae	Spelaeogriphidae	Physidae			
	Sternophysingidae	Planorbidae			
	Teloganodidae	Polycentropodidae			

A total of 40 taxa are shared between all med-regions (Table 2), being dipterans the group with more common families. Ephemeroptera as Baetidae, Caenidae and Leptophlebiidae are also found in all med-regions, as well as the caddisflies Hydropsychidae, Hydroptilidae, Polycentropodidae and Leptoceridae, or other Coleoptera, Heteroptera and Odonata. On the other hand, many taxa are shared between several med-regions. For instance, because of their strong similarity, Med-Basin and California have 16 taxa in common (excluding the ubiquitous ones), as the stoneflies Nemouridae, Leuctridae, Capniidae, Chloroperlidae and Perlodidae, or the caddisfly Rhyacophilidae (absent in the southern hemisphere). Other taxa are present in all the med-regions except in one, as Pyralidae only absent in South Africa, Corduliidae or Ecnomidae in California, Glossosomatidae in SWAustralia or Athericidae, Dugesiidae and Hydraenidae in SAustralia. Moreover, this last continent lacks of families abundant in other regions as Blephariceridae, Dryopidae or Sericostomatidae (see Annex 1).

Crustaceans are the group more convergent between regions in southern hemisphere, with some variability. From the common taxa between SAustralia and SWAustralia, 5 are crustaceans (Ceinidae, Hymenosomatidae, Oniscidae, Parastacidae and Perthiidae), at the same time that Janiridae and Paramelitidae are shared with South Africa. This taxonomical group has numerous exclusive families in Southern Hemisphere (4 in SAustralia, 5 in South Africa and 2 in Chile). In northern hemisphere, there are few crustacean families, and only the Astacidae seems to be exclusive in Med-Basin and California. Other similarities in the crustacean taxa are between some taxa that has been introduced in several med-regions as Cambaridae from North America, introduced in Med-Basin and South Africa, or Grapsidae introduced in California and Med-Basin regions.

Another taxonomical group shared between med-regions in southern hemisphere is Odonata, with South Africa and SAustralia having 2 families in common (Chorolestidae and Protoneuridae), 1 between Chile and SWAustralia (Petaluridae), and 2 between both Australian med-regions (Synthemidae and Megapodagrionidae). Although there are a high similarity between Odonata and Crustacea in southern hemisphere, few convergences exist in other groups (excluding the ubiquitous taxa), as Ephemeroptera, with only the Onicigastridae shared by Chile and SAustralia. Plecoptera shows a typical distribution around med-regions, with more families in northern hemisphere (until 8) than in the South one: 4 in SAustralia, 7 in Chile, 1 in SWAustralia and 1 in South Africa. Notonemouridae is exclusive from South Africa, Chile and SAustralia, and absent in SWAustralia. In this last region, the Gripopterygidae are present and only shared with SAustralia and Chile. On the other hand,

Chile and SAustralia have 2 common stoneflies (Austroperlidae and Eustheniidae) absent in other med-regions.

Many connections are found between north and southern hemisphere, as Noteridae or Platycnemidae in Med-Basin and South Africa, or Corylidae in South Africa, Chile and California. A relationship between northern hemisphere and South Africa or Chile is noticed, with some coldstenoterm families in fast flowing waters as Heptageniidae, Lepidostomatidae or Goeridae only in South Africa, California and Med-Basin; or as for example Perlidae, Psychomyiidae and Limnephilidae in the northern hemisphere and Chile. Other linkages between both Hemispheres are recent because the human introductions of Mollusca (Physidae in Australia, Hydrobiidae in Med-Basin, Corbiculiidae in California) or Crustacea (Cambaridae in Med-Basin and Grapsidae in California and Med-Basin).

When the list of taxa is analyzed for each med-region according to several multimetric indexes (Figure 4a-h), a differential composition among med-regions is found (Figure 4d-h). Communities from northern hemisphere have higher richness in Ephemeroptera, Plecoptera and Trichoptera, while in the southern hemisphere Coleoptera, Heteroptera, Diptera and non-Insect are more abundant. Taxa richness (as number of taxa) is slightly higher in Med-Basin (125), California (119), South Africa (104) and SAustralia (120), compare with Chile (92) and SWAustralia (98). However, even though this similarity in richness, taxonomical composition between med-regions is different (Figure 4b-h). Number of EPT appears to be high in Med-Basin, California, Chile and SAustralia (Figure 4b), because a major number of Plecoptera (Figure 4f), while is low on South Africa and SWAustralia with only one stonefly family (Notonemouridae and Gripopterygindae, respectively) (Annex 1). On the other hand, the low value of EPT/OCH in Australia would indicate that although having a high EPT value, the OCH are more important than in other med-regions (Figure 4h). In general, in southern hemisphere the percentage of non-insect taxa is higher, because of the presence of many crustacean families (Figure 4d). This phenomenon is more important in Australia and South Africa than in Chile, because in this last region the community is mainly composed by Ephemeroptera, Plecoptera and Trichoptera, being more similar to the northern hemisphere one at order level (Figure 4e-g).



Figure 4. Multimetric indexes between mediterranean regions.

Differences between all sampling sites in med-regions

Physical, chemical and structural differences

Pearson correlations between environmental variables, presented significant correlation between pH-Temperature (p=0.001), Temperature-Conductivity (p=0.034), Temperature-Discharge (p=0.03), Temperature-IHF (p=0.008) and QBR-Conductivity (p=0.002) (Table 3). Because the correlation coefficients were low (r<0.4) all variables were used in further analysis. Skewness values associated to each variable showed that Conductivity (s=7.449) and Discharge (s=7.569) had a high data asymmetry and consequently were deleted from the PCA analysis (Table 3). Only QBR, pH, temperature and IHF were used to avoid a bias in the results.

Table 3. Pearson correlations and skewness values for all measured environmental features. In bold, significant correlations (p<0.05) and high skewness values. Temp: temperature; Cond: conductivity; Dis: discharge.

		pН	Temp	Cond	QBR	IHF	Dis	Skewness
pН	r-Pearson		-0.379	0.137	-0.226	0.183	-0.133	
	p-value		0.001	0.245	0.053	0.118	0.260	-1.310
Temp	r-Pearson	-0.379		0.246	-0.067	-0.305	0.252	
	p-value	0.001		0.034	0.571	0.008	0.030	0.004
Cond	r-Pearson	0.137	0.246		-0.359	-0.146	-0.069	
	p-value	0.245	0.034		0.002	0.214	0.561	7.449
QBR	r-Pearson	-0.226	-0.067	-0.359		0.101	0.122	
	p-value	0.053	0.571	0.002		0.391	0.299	-1.854
IHF	r-Pearson	0.183	-0.305	-0.146	0.101		-0.114	
	p-value	0.118	0.008	0.214	0.391		0.335	-0.249
Dis	r-Pearson	-0.133	0.252	-0.069	0.122	-0.114		
	p-value	0.260	0.030	0.561	0.299	0.335		7.569

The first two PCA axis from the analysis made with selected variables explain 68.74% of total variability of the physical-chemical data and results are plotted in Figure 5. First axis is highly correlated positively with pH (r=0.5891) and IHF (r=0.4995) and negatively with temperature (r=-0.6315), whereas QBR is located positively and near the second axis (r=0.8313). Thereby, X₂-axis would be related to a good riparian status, and the X₁-axis with abiotic factors as pH, temperature and habitat diversity availability. A clear segregation of sites is noticed between med-regions, and different physical-chemical conditions can be associated with each one.

Reference sites in Med-Basin have a higher pH and IHF compare with SWAustralia and South Africa, but a lower riparian quality. Californian sites are located near Med-Basin with more neutral pH but variable conditions of riparian vegetation and habitat diversity. Finally, samples from South Africa and SWAustralia are associated with low pH, high temperature, good riparian quality and medium to lower habitat diversity.



Figure 5. PCA plot with environmental variables in reference sites.

Relative abundance of common taxa between sampled med-regions

In Table 4, results of the non-parametric ANOVA (model Kruskal-Wallis by ranks) indicate that in permanent sites, the relative abundance of taxa is variable between med-regions, with a 51% of common taxa showing significative differences in abundance between regions (pvalue<0.05 in Table 4). The rest of taxa have a similar abundance between regions, including the abundants Caenidae, Hydropsychidae, Hydroptilidae, Ostracoda, Corixidae, several Odonata and several infrequent taxa (Thyaridae, Sphaeriidae, Sciomyzidae, Dolichopodidae...).

Table 4. Kruskal-Wallis non-parametric tests with relative abundances of common taxa in MedBasin, California, South Africa and SWAustralia permanent sites. *p<0.05 and **p<0.001.

	Chi-squared (K-W test)	p-valu	e
Aeshnidae	12.46	0.006	**
Ancylidae	8.4	0.038	*
Anthomyidae	1.58	0.662	
Athericidae	11.4	0.009	**
Atyidae	8.15	0.043	*
Baetidae	11.79	0.008	**
Caenidae	3.5	0.32	
Ceratopogonidae	14.5	0.002	**
Chironomidae	16.12	0.001	**
Coenagrionidae	8.43	0.037	*
Corixidae	1.05	0.787	
Culicidae	0.7	0.873	
Dixidae	12.25	0.006	**
Dolichopodidae	2.18	0.535	
Dugesiidae	4.16	0.244	
Dytiscidae	6.79	0.078	
Elmidae	11.12	0.011	*
Empididae	15.79	0.001	**
Ephidridae	0	1	
Gerridae	17.36	0	**
Gomphidae	4.46	0.215	
Gyrinidae	4.04	0.256	
Haliplidae	6.22	0.101	
Helodidae	14.16	0.002	**
Hydracarina	11.44	0.009	**
Hydraenidae	20.19	0	**
Hydrobiidae	9.25	0.026	*
Hydrometridae	2.18	0.535	
Hydrophilidae	2.02	0.567	
Hydropsychidae	1.66	0.644	
Hydroptilidae	0.62	0.89	
Leptoceridae	23.62	0	**

Leptophlebiidae	17.11	0	**
Lestidae	2.3	0.511	
Libellulidae	1.6	0.658	
Lymnaeidae	7.42	0.059	
Naucoridae	2.18	0.535	
Nepidae	6.4	0.091	
Notonectidae	9.49	0.023	*
Oligochaeta	15.16	0.001	**
Ostracoda	6	0.111	
Philopotamidae	11.58	0.009	**
Physidae	10.41	0.015	*
Planorbidae	5.71	0.126	
Pleidae	2.18	0.535	
Polycentropodidae	8.43	0.037	*
Psychodidae	1.64	0.648	
Sciomyzidae	0	1	
Simuliidae	19	0	**
Sphaeriidae	5.96	0.113	
Stratiomyidae	9.25	0.026	*
Tabanidae	8.43	0.037	*
Thyaridae	1.56	0.459	
Tipulidae	13.79	0.003	**
Veliidae	11.89	0.007	**

Figure 6 plot the average of relative abundance patterns for all significant taxa (p<0.005) between sampled med-regions permanent sites. Oligochaeta, Baetidae, Chironomidae and Simuliidae are very abundant in all med-regions although several differences and similarities are present between them. For most groups of taxa, greater differences in the abundance pattern are shown in the interhemispheric comparisons, whereas the intrahemispheric show higher similarities (Figure 6). The sampled area in Med-Basin is characterized by a higher number of Mollusca, Coleoptera, Chironomidae, Notonectidae and Baetidae, but a lower Leptophlebiidae, Hydracarina and Simuliidae compared with California (Figure 6). Fewer similarities in relative abundance are shown between South Africa and SWAustralia with higher numbers of Leptoceridae in contrast to MedBasin and California.



Figure 6. Mean of the abundances relatives for permanent sampled sites in MedBasin, California, South Africa and SWAustralia. Only the significant taxa are plotted (p<0.005).

However, southern hemisphere med-regions present lower abundances of Mollusca, Heteroptera, Polycentropodidae and some Diptera compared with northern hemisphere. South Africa is characterized by a high relative abundance of Elmidae, Helodidae and Hydraenidae compared with the other med-regions, but a low abundance in Mollusca. In SWAustralia, Oligochaeta, Aeshnidae, Leptophlebiidae, Chironomidae, Simuliidae and Ceratopogonidae are more abundant than other med-regions. Consequently, other taxa are less frequent as Coleoptera, Heteroptera, Hydracarina, Polycentropodidae, Philopotamidae and Baetidae (Figure 6).

Effect of temporality

Although similarities and differences between some med-regions were shown by common taxa abundances for permanent sites, responses of communities to temporality may also differ between areas. Results from Bray-Curtis Cluster made with common taxa in all med-regions (Figure 7), indicate differences between regions are higher than between permanent and temporary sites. Each region is grouped together, and in each cluster permanent and temporary sites are included. MedBasin is the first region to be clustered, followed by SWAustralia. California and South Africa appear closer, indicating a similar response of communities to temporality. However, when each region is analysed independent of the other, and all taxa (common and exclusive) are analysed, differences in macroinvertebrate community between permanent and temporary sites (p<0.01) are found in all regions except for SWAustralia (p>0.05), as is show in Table 5.

	PERMANE	NT vs TEMPORAL	RIFFLES	vs POOLS
	А	p-value	А	p-value
Med-Basin	0.036	0.0000014**	0.03	0.017*
California	0.0664	0.0084**	0.1182	0**
SouthAfrica	0.0666	0.0094**	0.0839	0**
SWAustralia	0.011	0.3192	0.0789	0**

Table 5. MRPP results between different categories and regions. *p<0.05 and **p<0.001.

The macroinvertebrate assemblage in SWAustralia and South Africa has a number of taxa, dipterans, EPT/OCH similar between permanent and temporary sites (Figure 8 and Table 6).

For the rest of med-regions number of taxa and % of non-insect taxa is similar between permanent and temporary sites, but there are differences in the EPT/OCH ratio and the number of dipterans. MedBasin has a significantly higher EPT/OCH in permanent sites than temporary, whereas in California, even though this ratio is also higher in permanent localities, it is not significant. Number of dipterans is significantly lower in permanent sites from California, but similar to temporary in the rest of sampling areas.



Figure 7. Bray-Curtis cluster with common matrix and permanent and temporary reference sites.

When the analysis are performed comparing all regions separately in permanent and temporary sites (Table 6), only number of taxa and EPT/OCH ratio in permanent sites are significantly different between med-regions. For both metrics, California has higher values in permanent sites compared with the rest of med-regions. No differences exist between regions for all metrics in temporary sites, although box-plot graphs indicate lower higher number of taxa and dipterans in California but a low % of non-insect taxa respect the other regions.



Figure 8. Box-plot between permanent and temporary reference sites.

Table 6. Kruskal-Wallis non-parametric tests between med-regions and temporality. PERM=permanent sites and TEMP=temporary sites. * p<0.05 and **p<0.001.

		Taxa	EPT/OCH	Dipterans	%Non-insect
PERM among	Chi-squared K-W test	15.52	9.7916	4.7769	3.666
regions	p-value	0.0014**	0.0204*	0.1889	0.2998
TEMP among	Chi-squared K-W test	4.96	6.899	0.3886	3.6909
regions	p-value	0.1747	0.0752	0.274	0.2968
ModPosin	Chi-squared K-W test	0	4.72	0.364	2.1
Meabasiii	p-value	1	0.0297*	0.5459	0.1473
California	Chi-squared K-W test	0.4761	2.5	5.833	1.666
	p-value	0.4902	0.1139	0.0157*	0.1967
South Africa	Chi-squared K-W test	1.5	0.6857	2	1.5
South Allica	p-value	0.2207	0.4076	0.1573	0.2207
SW Australia	Chi-squared K-W test	1.2857	0.9	0	0.9
Sw Australia	p-value	0.2568	0.3428	1	0.3428

The values of IV-values for each taxon obtained with IndVal analysis are shown in Table 7. Firstly, it is interesting to note, that although no significant differences are found between permanent and temporary sites in SWAustralia (Table 5), several taxa appear to have a high IV-value. Overall, the general pattern observed indicates that in springtime, permanent community present more taxa characteristic from riffles, whereas temporary sites have a pool of more tolerant taxa. Thereby, EPT taxa are more common in permanent localities and OCH in temporary ones. Several convergences are observed among regions, with MedBasin and California sharing taxa as Rhyacophilidae, Perlidae and Ephemerellidae in permanent sites. Other plecopterans as Nemouridae have a high IV-value in temporary sites in MedBasin, but in permanent localities in California. The SWAustralia stonefly Gripopterygidae and Notonemouridae in South Africa have a high IV-value in temporary sites. Ephemeroptera are also more common in permanent sites than in temporary ones with Ephemerellidae in MedBasin and California and Teloganodidae in South Africa, or Heptageniidae in MedBasin and South Africa. On the other hand, Leptophlebiidae only is a family associated to permanent localities in SWAustralia whereas is more typical from temporary sites in MedBasin and South Africa. Some caddisflies are also more abundant in permanent than in temporary sites as the reophilic families Hydropsychidae (MedBasin, South Africa and SWAustralia) or Rhyacophilidae (MedBasin and California) or the close family Hydrobiosidae (SWAustralia) or its equivalent in South Africa (the predator Corydalidae). Hydroptilidae and the mayfly Caenidae are permanent families in MedBasin and SWAustralia but associated to temporary habitats in California. Other caddis exclusive form South Africa have found in permanent sites as the lotic families Barbarocthonidae and Petrothrincidae, whereas Pisuliidae is indicator of temporary localities. Leptoceridae, a very abundant caddisfly in southern hemisphere, is characteristic from permanent sites there, but from temporary reaches in California. Similar pattern is observed in Hydropsychidae, Elmidae and Heptageniidae, characteristic from permanent sites in MedBasin and South Africa but from more temporary sites in California.

Table 7. IndVal results between permanent and temporary sites for each med-region. Line separates taxa significant at p<0.05 from not significant. Non-significant taxa are from high indicator values (IV) to low. Only taxa with IV higher than 25 are presented (Dûfrene & Legendre, 1997).

MEDBASIN					
Permanent	IV	p-value	Temporary	IV	p-value
Elmidae	65	0	Hydrometridae	72.5	0
Baetidae	69	0	Veliidae	41.4	0
Hydropsychidae	70	0	Hydrophilidae	38.1	0.03
Heptageniidae	56	0.02	Planorbidae	33.1	0.04
Gomphidae	44	0.02	Culicidae	28.2	0.04
Perlidae	49	0.02	Dytiscidae	54.1	0.05
Calopterygidae	38	0.04	Chironomidae	52.9	0.26
Hydroptilidae	48	0.05	Ostracoda	42.2	0.09
Simuliidae	49	0.07	Oligochaeta	41.8	0.43
Rhyacophilidae	48	0.07	Gerridae	37.8	0.21
Ephemerellidae	39	0.15	Stratiomyidae	34.7	0.29
Caenidae	39	0.18	Leptophlebiidae	33.8	0.12
Hydracarina	38	0.54	Notonectidae	33.2	0.17
Hydrobiidae	35	0.11	Ceratopogonidae	32	0.79
Gammaridae	34	0.13	Hydraenidae	31.4	0.23
Helodidae	31	0.09	Lymnaeidae	30.5	0.18
Limnephilidae	30	0.97	Nemouridae	29.9	0.45
Brachycentridae	28	0.13			
Ephemeridae	25	0.16			

CALIFORNIA					
Permanent	IV	p-value	Temporary	IV	p-value
Ceratopogonidae	81	0.029	Culicidae	100	0.029
Limnephilidae	100	0.029	Stratiomyidae	95.2	0.053
Perlidae	85.7	0.029	Caenidae	88.9	0.073
Rhyacophilidae	85.2	0.029	Corixidae	85.7	0.074
Nemouridae	85.7	0.055	Brachycentridae	80	0.078
Ephemerellidae	84.6	0.058	Physidae	76.2	0.108
Nematoda	75	0.138	Hydroptilidae	75.7	0.053
Sialidae	75	0.164	Philopotamidae	63.2	0.202
Tipulidae	64.7	0.071	Oligochaeta	58.3	0.35
Empididae	61.9	0.286	Dixidae	57.1	0.589
Dytiscidae	52.5	0.629	Chloroperlidae	57.1	0.618
Calamoceratidae	50	0.453	Ostracoda	55.8	0.683
Sphaeridae	50	0.457	Hydracarina	54.9	0.352
Peltoperlidae	50	0.475	Lepidostomatidae	54.1	0.58
Asellidae	37.5	0.55	Hydropsychidae	52.2	1
Cordulegasteridae	37.5	0.681	Heptageniidae	51.3	0.933
Perlodidae	37.5	0.705	Elmidae	51.1	1
Odontoceridae	37.5	0.709	Belostomatidae	50	0.2
Psephenidae	37.5	0.912	Ephydridae	50	0.2
Glossosomatidae	28.4	1	Haliplidae	50	0.2
Psychodidae	27.8	1	Sciomyzidae	50	0.2
Corydalidae	25	1	Coenagrionidae	50	0.212
	25	1	Hydraenidae	50	0.212
			Naucoridae	50	0.212
			Tabanidae	50	0.212
			Polycentropodidae	50	0.465
			Chironomidae	50	1
			Gerridae	50	1
			Psychomyiidae	44.4	0.2
			Hydrophilidae	40	0.222
			Tricorytidae	40	0.238
			Lestidae	40	0.377
			Gomphidae	36.4	0.601
			Planorbidae	36.4	0.602
			Dugesiidae	35.3	0.593
			Sericostomatidae	35.3	0.729
			Siphlonuriidae	33.3	0.381
			Pteronarcydae	33.3	0.381
			Leptoceridae	33.3	1
			Oribatidae	28.6	1
			Uenoidae	28.6	1
			Aeshnidae	28.6	1
			Cambaridae	25	1
			Helicopsychidae	25	1

SOUTH AFRI	CA				
Permanent	IV	p-value	Temporary	IV	p-value
Barbarocthonidae	88	0.01	Paramelitidae	86.7	0.01
Leptoceridae	71	0.02	Hydracarina	56.2	0.46
Hydropsychidae	67	0.08	Nematoda	51.9	0.81
Blepharoceridae	66	0.11	Oligochaeta	51.7	0.58
Petrothrincidae	63	0.12	Leptophlebiidae	51.1	1
Elmidae	57	0.06	Chironomidae	50	1
Teloganodidae	57	0.18	Notonemouridae	47.8	1
Baetidae	56	0.17	Helodidae	47.8	1
Hydraenidae	56	0.56	Dytiscidae	42.9	0.17
Simuliidae	54	0.07	Ostracoda	37.5	1
Glossosomatidae	53	0.31	Dugesiidae	30.8	0.85
Tipulidae	52	0.37	Hirudinidae	25	0.32
Coenagrionidae	50	0.19	Pisuliidae	25	0.34

SWAUSTRALIA							
Permanent	IV	p-value	Temporary	IV	p-value		
Hydropsychidae	68.4	0.13	Griptopterygidae	61.5	0.25		
Baetidae	66.7	0.172	Hydracarina	53.3	0.235		
Leptophlebiidae	61.5	0.23	Oligochaeta	51.6	1		
Simuliidae	55.6	0.358	Chironomidae	50	1		
Aeshnidae	55.6	0.48	Perthiidae	44.4	0.393		
Ceratopogonidae	54.5	0.635	Psychodidae	44.4	0.48		
Leptoceridae	52	1	Dytiscidae	44.4	0.523		
Corduliidae	50	0.371	Ostracoda	44.4	0.643		
Athericidae	50	0.387	Tipulidae	38.1	0.519		
Hydroptilidae	50	0.396	Coenagrionidae	33.3	0.305		
Hydrobiosidae	50	0.412	Gyrinidae	33.3	0.325		
Empididae	47.6	0.461	Stratiomyidae	33.3	0.325		
Parastacidae	44.4	0.543	Nematoda	29.6	1		

Temporary sites are characterized by high abundance of OCH taxa, together with some other taxa as Mollusca, Oligochaeta and Crustacea. Dytiscidae is a temporary family in MedBasin, South Africa and SWAustralia but not in California, whereas Haliplidae only have a high IVvalue in temporary sites of California. On the other hand, Gyrinidae is an indicator family only of SWAustralia and MedBasin. Heteroptera are exclusively temporary families, but Odonata have permanent and temporary taxa. Lestidae, Gomphidae and Coenagrionidae are temporary families in California. Coenagrionidae also is a temporary family in SWAustralia but a permanent in South Africa. Similar divergences are observed with Gomphidae, a significant Odonata family found in permanent sites in MedBasin but with a high IV-value in temporary sites of California. Crustaceans and Mollusca appear to be more indicator taxa from temporary than permanent sites, with Planorbidae present in MedBasin and California, or Lymnaeidae common in MedBasin and Physidae in California. Other taxa are present in temporary sites in all med-regions, as Ostracoda or Oligochaeta.



Figure 9. Percentages of similarity between med-regions considering common abundant and representative (IV-value) taxa for each pair of med-regions, in permanent and temporary sites.

Figure 9 displays percentages of similarities in macroinvertebrate responses under permanency and temporality between each pair of med-regions. These similarity values were obtained using common taxa with high IV-values (very abundant and representative) present between pair of med-regions. When comparing two med-regions if a common family was representative from different conditions (permanent or temporary) in one region respect to the others, it was recorded as a disimilarity, whereas when a family display a similar affinity in both regions it was recorded as a similarity. Thereby, and according to Figure 9, the highest percentage of similarity is shown between South Africa and SWAustralia, indicating that common taxa with high indicator values display a similar pattern in both regions. Consequently, 94.4% of the common and abundant taxa between South Africa and SWAustralia present a similar response to the permanency and temporality. On the other hand, MedBasin and SWAustralia also show a high similarity, compared with the rest of paired-comparisons. In fact, California and MedBasin, even though sharing more families with a high IV-value than the other regions only 54.2% of the taxa display a similar segregation in permanent and temporary sites.

Habitat effects

In Figure 10 the cluster using all common taxa in sampled med-regions is presented. Differences between regions are higher than between habitats in South Africa and SWAustralia





but not for MedBasin and California. In northern hemisphere med-regions, common taxa have a similar composition and abundance in each habitat independently of the region, especially in pools. Riffles in MedBasin, although they are closer to the Californian ones, have a higher variability in composition and abundance and are closer to SWAustralia and South Africa than to lentic habitats of Med-Basin and California. However, for each region independently and all taxa, MRPP results (Table 5) show that in studied sites all med-regions have significant differences between the macroinvertebrate assemblage of riffles and pools. Number of taxa is similar between R and L in all regions (Figure 11 and Table 8), but some differences are found with the rest of metrics. EPT/OCH ratio is significant higher in R than L in all med-regions except for South Africa (p>0.005) (Table 8). SWAustralia and South Africa have a significant higher number of dipterans in R compared with MedBasin and California. Percentage of noninsect taxa only displays significant differences among habitats in South Africa where a lower number of non-insect taxa are observed in R than L.



Figure 11. Box-Plot between riffles and pools for reference sites.

		Таха	EPT/OCH	Dipterans	%Non-insect
R among regions	Chi-squared K-W test	15.76	19.51	5.25	11.91
	p-value	0.0013**	0.0002**	0.1539	0.0077**
T	Chi-squared K-W test	19.48	2.6	17.07	15.46
	p-value	0.0002**	0.4575	0.0007**	0.0015**
ModPosin	Chi-squared K-W test	0	4.666	1	0.253
MeuDasin	p-value	1	0.0308*	0.3173	0.6143
California	Chi-squared K-W test	0.202	9.8989	1.8181	3.2
Camorina	p-value	0.6531	0.0017**	0.1775	0.0736
South Africa	Chi-squared K-W test	2.7428	2.666	8.7111	6
South Annea	p-value	0.0977	0.1025	0.0032**	0.0143*
SW Australia	Chi-squared K-W test	0.9	7.244	5.555	2
Sw Australia	p-value	0.3428	0.0071**	0.0184*	0.1573

Table 8. Kruskal-Wallis non-parametric tests between med-regions and habitat. R=riffles and L=pools. * p<0.05 and **p<0.001.

When habitats are considered separately, several differences between med-regions are found in all metrics. Thus, California has a higher number of taxa in R and L separately, a higher EPT/OCH ratio in riffles and a higher number of dipterans in L, compared with the rest of med-regions. Significant differences found in the percentage of non-insect fauna are because of the lower values found in SWAustralia in both habitats separately, in contrast with the observed in the rest of sampled areas (Table 8).

Convergences and divergences between habitats in med-regions are given in detail in Table 9, where IndVal results are presented. The significant families and the ones with a high IV-value but non-significant have to be understood in terms of presence and relative abundance of each taxa in each region. Hydropsychidae and Simuliidae are significantly abundant in riffles in all med-regions, whereas other taxa are exclusive from one, two or three regions. All the plecopterans found in MedBasin and California and the respective families found in southern hemisphere (Notonemouridae and Gripopterygidae) have higher IV-value in riffles than in pools. Within the Ephemeropterans, Ephemerellidae is typical from riffles in California and MedBasin, and the close family Teloganodidae is in South Africa. Heptageniidae is a common family found in riffles in MedBasin and California, but without habitat preferences in South Africa and absent in SWAustralia. Differences in habitat preferences are found in

Table 9. IndVal results between riffles and pool habitats for each med-region. Line separates taxa significant at p<0.05 from not significant. Non-significant taxa are from high indicator values (IV) to low. Only taxa with IV higher than 25 are presented (Dûfrene & Legendre, 1997).

MEDBASIN					
Riffles	IV	p-value	Pools	IV	p-value
Hydropsychidae	63	0.03	Gerridae	87.5	0
Simuliidae	63	0.03	Dytiscid	62.5	0.02
Ephemerellidae	60	0.07	Hydrometridae	45.5	0.21
Ceratopogonidae	55	0.21	Culicidae	37.5	0.2
Chironomidae	54	0.56	Hydrophilidae	37.5	0.2
Nemouridae	51	0.12	Coenagrionidae	37.5	0.21
Elmidae	51	0.11	Hydracarina	37.5	1
Chloroperlidae	50	0.08	Leptophlebiidae	33.7	0.89
Erpobdellidae	50	0.08	Polycentropodidae	27.3	0.6
Hydroptilidae	50	0.08	Nepidae	25	0.46
Perlodidae	50	0.09	Notonectidae	25	0.47
Limnephilidae	49	0.5	Hydrobiidae	25	0.47
Perlidae	44	0.12			
Betidae	43	0.61			
Heptageniidae	42	0.33			
Oligochaeta	40	0.8			
Tabanidae	38	0.2			
Leuctridae	38	0.2			
Hydraenidae	33	0.34			
Sericostomatidae	31	0.33			
Ostracoda	29	0.77			
Caenidae	29	0.8			
Empididae	30	0.43			
Rhyacophilidae	27	0.54			
Glossosomatidae	25	0.47			
Limoniidae	25	0.47			

CALIFORNIA					
Riffles	IV	p-value	Pools	IV	p-value
Hydropsychidae	93.3	0.001	Gerridae	84.4	0.002
Simuliidae	87.8	0.001	Dytiscidae	74.1	0.005
Tipulidae	73.5	0.005	Sialidae	60	0.007
Rhyacophilidae	72	0.008	Leptophlebiidae	64.3	0.027
Perlidae	68.2	0.016	Ceratopogonidae	60	0.077
Baetidae	56.2	0.017	Ostracoda	58.4	0.068
Chloroperlidae	66.7	0.031	Lepidostomatidae	56.8	0.07
Heptageniidae	60	0.085	Chironomidae	50	1
Ephemerellidae	58.2	0.139	Limnephilidae	46.2	0.296
Oligochaeta	57.4	0.188	Calamoceratidae	40	0.09
Elmidae	55.8	0.187	Gomphidae	40	0.106
Nemouridae	53.6	0.269	Oribatidae	40	0.121
Hydracarina	51.9	0.744	Sericostomatidae	35.6	0.179
Empididae	48.7	0.516	Physidae	34.3	0.555
Philopotamidae	45.5	0.07	Corixidae	33.3	0.228
Polycentropodidae	43.6	0.064	Sphaeridae	32	0.299
Hydroptilidae	42	0.43	Nematoda	31.2	0.646
Glossosomatidae	36.7	0.443	Cordulegasteridae	30	0.204
Psephenidae	35	0.707	Aeshnidae	30	0.232
Peltoperlidae	34.3	0.247	Dixidae	30	0.998
Dugesiidae	32	0.225			
Uenoidae	31.1	0.341			
Brachycentridae	26.7	0.522			

SOUTH AFRICA								
Riffles	IV	p-value	Pools	IV	p-value			
Simuliidae	78	0	Ostracoda	55.3	0.06			
Hydraenidae	74	0	Oligochaeta	53.8	0.23			
Philopotamidae	67	0	Nematoda	44.9	0.66			
Teloganodidae	62	0.01	Ceratopogonidae	39.7	0.68			
Notonemouridae	67	0.01	Ecnomidae	30	0.64			
Blepharoceridae	55	0.01	Coenagrionidae	28.6	0.22			
Hydropsychidae	53	0.02						
Leptophlebiidae	63	0.02						
Baetidae	58	0.08						
Elmidae	57	0.11						
Helodidae	55	0.17						
Chironomidae	52	0.15						
Hydracarina	51	0.53						
Empididae	49	0.15						
Tipulidae	48	0.06						
Leptoceridae	44	0.81						
Glossosomatidae	38	0.12						
Corydalidae	33	0.1						
Potamonantidae	25	0.21						

SWAUSTRALIA					
Riffles	IV	p-value	Pools	IV	p-value
Simuliidae	90.3	0.001	Leptoceridae	59.3	0.146
Hydropsychidae	88.9	0.001	Ostracoda	47.6	0.122
Ceratopogonidae	62.5	0.08	Dytiscidae	35.6	0.279
Griptopterygidae	60.4	0.12	Corduliidae	35.4	0.557
Oligochaeta	60.3	0.134	Gomphidae	29.6	0.621
Leptophlebiidae	54.5	0.232	Caenidae	25.9	0.372
Chironomidae	50.9	1	Ecnomidae	25	0.583
Empididae	46.3	0.116			
Aeshnidae	38.1	0.581			
Hydrobiosidae	33.3	0.206			
Athericidae	33.3	0.349			
Nematoda	33.3	0.588			
Baetidae	25	0.353			

Leptophlebiidae that have a high IV-value in riffles in California, South Africa and SWAustralia but not in MedBasin, where is classified as a pool-specie. The Coleopterans Elmidae and Hydraenidae are present in riffles in the regions where they were found: MedBasin, California and South Africa. The Dipteran Ceratopogonidae is characteristic from riffles in MedBasin and SWAustralia but from pools in California and South Africa. The highly abundant Chironomidae is a riffle-family in MedBasin, South Africa and SWAustralia but not in California. Several non-cased caddisflies as Rhyacophilidae in MedBasin and California, Polycentropodidae in California, Hydrobiosidae in SWAustralia are common in riffles jointly with the stony-case caddisflies as Glossosomatidae. Other cased caddisflies, as Sericostomatidae, have a high IV-value in riffles in MedBasin but in pools in California. On the other hand, Leptoceridae have a high IV-value in riffles in South Africa but in pools in SWAustralia.

Odonata, Heteroptera and Coleoptera are dominant taxa in pools. Dytiscidae, Gerridae, Leptophlebiidae and Culicidae are tolerant lentic families shared between MedBasin and California. At the same time, Dytiscidae is also characteristic and abundant in SWAustralia. California have a longer list of families with a high IV-value compared with MedBasin, with Gomphidae, Cordulegasteridae and Aeshnidae, together with the woody-cased caddisflies Lepidostomatidae, some Limnephilidae and Calamoceratidae. Some Mollusca very abundant in pools in MedBasin as Hydrobiidae are not significantly present in California but others with a high IV-value as Physidae were collected. In southern hemisphere fewer number of Mollusca taxa was found. The Odonata Coenagrionidae is a shared and abundant family in pools in MedBasin and South Africa, whereas Gomphidae is in California and SWAustralia. Ostracoda have been collected in all med-regions, but were more significantly found in pools in California, South Africa and SWAustralia.

Comparisons between common and characteristic community according to habitat preferences between pairs of med-regions were performed and the results are plot in Figure 12. Higher similarities between riffles and pools are observed between regions than the between permanent and temporary sites. All med-regions have convergent responses of the common taxa in front of reophilia, with 80% of similarity between California and South Africa and SWAustralia. MedBasin also have high similarities with the rest of med-regions, with over than 64% with South Africa, and around 70% with California and SWAustralia. On the other hand, the percentage of similarity between South Africa and SWAustralia appears to be lower than the observed when permanent and temporary conditions are compared.



Figure 12. Percentages of similarity between med-regions considering common abundant and representative (IV-value) taxa in riffles and pools habitats.

DISCUSSION

Traditionally, convergence in ecology has had a deterministic framework, where communities developed under similar conditions would have a common structure (Cody & Mooney, 1978). However, several approaches argue that historical factors and intrinsic indeterminism can imply divergence even under similar environments (Samuels & Drake, 1997). The scale of study is important to check for global convergences and divergences between regions, as some scales of observation can obscure others (Samuels & Drake, *op. cit.*). Overall, our study has focused at two levels with the mediterranean climate as a common denominator: regional and local.

Regional scale: historical factors

Appropriate taxonomical sufficiency in aquatic studies has been highly discussed in stream ecology, and especially in bioassessment (Resh & Unzicker, 1975; Cranston, 1990; Marchant *et al.*, 1995; Bowman & Bailey, 1997). Although the use of a low taxonomical resolution implies a gain in precision and information in the responses (Furse *et al.*, 1984; Resh *et al.*, 1995), at large scale, when different regions in the world are compared in taxonomical composition and ecological responses, higher taxonomical levels can be more useful as

speciation have acted in a different way in each region. On the other hand, several authors have demonstrated that ecological patterns observed at higher taxonomical levels are equivalent to the ones at genus or species (e.g. Furse *et al.*, 1984; Ferrano & Cole, 1992; Rutt *et al.*, 1993; Marchant *et al.*, 1995). However, because the high speciation present in some families (e.g. Leptoceridae in southern hemisphere —de Moor, 1988, 1997) some divergences observed between med-regions could be attributed to dissimilar ecological responses because of the existence of different genus and species between regions. When detailed taxonomical information is not available, comparisions at genus or species level are impossible. In this case, as in our study, studies at higher taxonomic level may provide interesting insigths about the importance of historical and ecological factors (Ricklefs, 1987).

According to the results obtained, when macroinvertebrate assemblages at family level are studied in the 5 med-regions in the world, the historical factor seems very important in determining macroinvertebrate community of streams in med-regions (see Figure 13 for summary). Convergences found between med-regions in macroinvertebrate community are unlikely related to the climate because of the short history of the mediterranean climate (Alxelrod, 1973; Suc, 1984), and seems to be related to previous historical factors. Notable differences are found between northern and southern hemisphere, as the early breakup between Laurasia and Gondwanaland during the Cretaceous (70-135 myBP). As in plant communities (Deacon, 1983), macroinvertebrates from boreal med-regions are highly convergent, because the connections between both regions until the Eocene by Bering bridge (Deacon, 1983), whereas macroinvertebrates in southern hemisphere are more divergent between med-regions because the early break up of Gondwanaland in early Cretaceous. This strong convergence observed between MedBasin and California implies lower exclusivity of taxa in both regions at family level. On the other hand, some exclusive families in MedBasin or California have a wider distribution than med-region it self. For instance, Prosopistomatidae or Polymitarcidae from MedBasin are also found in African areas (Edmunds, 1972), and Pteronarcyidae or Peltoperlidae in California have been recorded in other Neartic and Asian regions (Zwick, 2000).

High convergences have been reported in plants between California and Chile because the exchange of taxa through the Panama istm, 3 My ago (Naveh & Wittaker, 1979; Di Castri, 1991). Contrary to that, our study shows that Chilean macroinvertebrate community have a weak convergence with California med-region at family level, what could be related with the faster dispersion of plants in front of invertebrates (Cox & Moore, 1993) and the lowest richness by the insular situation of the country between Andes, Atacama desert and Pacific



Figure 13. Convergence graph among mediterranean regions. The line width is a representation of the degree of similarity between mediterranean areas.

ocean (Armesto *et al.*, 1997). However, several taxa originated in northern hemisphere are present in Chile and absent in other southern med-regions, as Limnephilidae, Psychomyiidae, Perlidae or Notonemouridae (nowadays extinct in northern hemisphere —Zwick, 2000). These cold-temperate fauna could have arrived to Chile through Panama istm and be established because of the local conditions provided by Andes ranges. On the other hand, as in plant community (Naveh & Wittaker, 1979), we found that Chile have a weak convergence between med-regions in southern hemisphere in macroinvertebrates. However, more families are shared with Australia than South Africa because of the connections between the Neotropical and Austral areas until late Cretaceous (Brundin, 1965; Harrison, 1978).

As it has been seen in plants (Cowling, 1992), macroinvertebrate assemblage in South African med-region is more convergent to tropical adjacent areas and northern hemisphere than to Chile or Australia. This relationship with the boreal area would indicate an active migration during Miocene or Pliocene, of organisms originated in the Paleartic and Neartic area through the south by the African uplands, as for example some coldstenotermous macroinvertebrates (Heptageniidae, Notonemouridae, Lepidostomatidae and Goeridae). Moreover, some convergences between South Africa and the Holartic area as the Tricorythidae originated in Africa (Edmunds, 1972), would explain that exchange in a different direction were also probable. However, not all taxa originated in northern hemisphere or tropical adjacent areas are found in the med-region. For instance, Perlidae, Sialidae, Oligoneuridae, Palaemonidae, Calamoceratidae or Psephenidae lack in Western Cape, although that does not imply a lost of convergence between northern hemisphere areas. In spite of that, we have found few similarities between South Africa and Australia (Notonemouridae, Chlorolestidae, Protoneuridae with SAustralia, and Athuridae and Amphisopodidae with SWAustralia). Day et al. (2001) pointed out that more convergences in crustaceans should have existed before the end of Cretaceous when an important drought affected South Africa, extinguishing many taxa.

Although closer in space, SWAustralia and SAustralia are more divergent in macroinvertebrates than California and MedBasin. This fact can be explained by the presence of an extreme arid period around 18000 yBP ago (DeDeckker, 1986) that affected SWAustralia region, implying several taxonomical extinctions and isolating this area from the east by 1000 km of desert. Consequently, further colonizations of macroinvertebrates from the east coast to the southwest were difficult and a low number of fauna in the southwest has been noticed by Bunn & Davies (1990).

Chapter 3

In spite of these observed divergences between macroinvertebrate assemblage in med-region, numerous ubiquitous taxa are present, what implies the existence of similar evolutionary mechanisms of dispersion, extinction and adaptation of the taxa along time (Deacon, 1983). Because of the dispersion and colonization of one organism depends on the environmental conditions and life cycles (Cox & Moore, 1993), organisms with short life cycles, continuous reproductions and aerial phases should be easily dispersed, and therefore they will have a wider distribution (Williams & Feltmate, 1993). Baetidae, Caenidae, Leptophlebiidae, Leptoceridae, Hydropsychidae, Oligochaeta and almost all dipterans are ubiquitous taxa, and most of them have been considered as organisms easily to disperse because their morphological and reproductive traits (for example see Gray, 1981; Gray & Fisher, 1981; Fisher *et al.*, 1982). Moreover, most of these taxa are characterized by having a very old origin (e.g., Baetidae, Caenidae and Leptophlebiidae) in contrast with others with a local distribution (e.g., Teloganodidae — Edmunds, 1972).

Historical factors are important to understand taxonomical convergences and divergences, but the viability and success of one taxa in one new region will depend on the local and environmental conditions (Resh & Solem, 1996). In that sense, Ball (1975) distinguishes between an analytical (Historical Biogeography) and an empirical biogeography (Ecological Biogeography). Thus, the abundance of EPT in Northern hemisphere and Chile respect other areas could be interpreted by the mountain river typology of some rivers flowing from high mountains near the coast (Sierra Nevada in California, Andes in Chile and Sierra Nevada, Pyrenees, Apennines, Atlas,... in Med-Basin), that let the establishment of species adapted to steep, cold and fast flowing rivers. On the other hand, even though some mountainous and high gradient rivers with cold waters are present in South Africa and SAustralia, the low EPT/OCH is explained by the poor contribution of Plecoptera in these regions because of biogeographical factors (Zwick, 2000), instead of environmental ones. This difficulty to discern between ecological and historical factors has been emphasized by Endler (1982), but both have to be present to understand community structure and composition (Ricklefs, 1987; Menge & Olson, 1990).

Local scale: Ecological factors and the spatio-temporal variability

In our study and according to the measured factors in reference conditions (without human disturbance), pH is a key variable to diferenciate med-regions. Calcareous geology in mediterranean basin (di Castri, 1981) is the responsible of a high pH in its rivers and streams (Toro *et al.*, (in press)), whereas heavily washed soils in South Africa and SWAustralia (Specht & Moll, 1983), provide a high acidity in reference conditions. In South Africa, fynbos vegetation

accentuates this phenomenon supplying highly concentrations of humic acids to the water (Midgley & Schafer, 1992). This gradient of med-regions in pH is followed by temperature in a contrary sense, related with local and microclimatic factors. For example, sites located near high mountains in California or MedBasin should have a lower temperature than the ones in the plains or coastal areas (Hornbeck *et al.*, 1983; Robles *et al.* (in press)).

This variability of physical and chemical factors and the high diversity in landscape topography between med-regions, influence the presence and abundance of several macroinvertebrate's taxa. Thus for instance, Mollusca are infrequent in South Africa because the acidic and oligotrophic conditions (Harrison & Agnew, 1962; Brown, 1978), but are abundant in Med-Basin localities with a calcareous geology (Martínez-López *et al.*, 1988; Gallardo-Mayenco *et al.*, 1994; Habdija *et al.*, 1995). However, mixed with local factors, historical events can also influence the abundance of some taxa respect others. For example, Leptoceridae and Leptophlebiidae are very abundant in South Africa and SWAustralia, because of its large diversification in southern hemisphere (de Moor, 1988, 1997; Edmunds, 1972). Consequently, a combination of local and historical factors appear, once again, as the responsible of the convergences and divergences observed in med-regions.

Disturbance (e.g., floods and droughts) is a key element to understand biodiversity in streams (Resh *et al.*, 1988; Vinson & Hawkins, 1998) independently of historical events. The response of a system to disturbance depends on the degree of stability (Wishart, 1998) and its frequency and intensity (Resh *et al.*, 1988). We should expect that annual discharge variation in mediterranean rivers (Gasith & Resh, 1999) could imply similar responses to the effect of temporality between med-regions except in extreme cases (e.g., in a more semiarid or arid areas), where the high unpredictability in the intensity and frequency of the occurrence of rainfall and discharge (Vidal-Abarca, 1990) would emphasize the differences between temporary and permanent sites.

Several comprehensive studies about the effect of drougths have been done in different regions in the world (see Boulton & Suter, 1986; Williams, 1987; Boulton & Lake, 1992ab). This flow disturbance induces an adapted macroinvertebrate community (Williams & Hynes, 1977; Williams, 1987; Delucchi & Peckarsky, 1989; Sommerhäuser *et al.*, 1997) in terms of resilience and resistance (Stanley & Fisher, 1992) with the former more important in temporary streams (Grimm & Fisher, 1989). Studies performed in mediterranean rivers also suggest that, temporary rivers and streams have macroinvertebrates of smaller size with multivoltine life cycles (Bonada *et al.*, Chapter 5), but at the same time adapted taxa have also developed.

Disagreements exist comparing differences between richness in permanent and temporary sites. For instance, Wrigth et al. (1984) or del Rosario & Resh (2000) found lower richness in temporary than permanent sites, whereas Legier & Talin (1973), Boulton & Suter (1986) or Miller & Golladay (1996) report similar number of taxa. In our case, similar richness has been recorded for all regions among permanent and temporary sites in springtime. But despite the similar species richness between permanent and temporary sites in all regions, several differences appear, specially when SWAustralia is compared with other regions. Bunn & Davies (1990) pointed out that the fauna in SWAustralia is depauperate compared with the one found in the east side of the continent, because the isolation of the SW and the previous history, that caused that permanent taxa with longer life cycles (e.g., Oligoneuridae, Siphlonuridae, Austroperlidae, Eustheniidae, Calocidae or Helicophidae) disappeared in the past, and further colonizations were difficult. Then, the taxa typical from permanent sites are absent in this area, what would explain the similarities between permanent and temporary sites in this med-region, although the higher precipitaion occurred in the area one year before sampling, respect the other med-regions, could be also another reason. On the other hand, Boulton & Lake (1992) studying several temporary sites in Australia and their nearby permanent ones found a significant faunal overlap, not found in other studies in northern hemisphere (Williams & Hynes, 1976; Casey & Ladle, 1979; Wrigth et al., 1984). In that sense, Wishart (1998) suggest that there are different responses to temporality between northern and southern hemisphere, with higher differences between permament and temporary communities in northern because of more stable rivers are present.

Looking for differences between several insect orders, we found riffle taxa (EPT) with higher abundance values in permanent sites, and a mixing of riffle and lentic communities (OCH and some EPT) in temporary ones. That would suggest that differences between permanent and temporary sites are consequence of a different proportion of riffles and lentic habitats. Legier & Talin (1973) in a study in several French mediterranean temporary and permanent rivers pointed out that some lotic taxa were absent in temporary sites, as Rhyacophilidae, Heptageniidae, Goeridae or Perlidae, and a similar pattern we have found in Med-Basin, California and South Africa. On the other hand, Boulton & Lake (1992a) studyiong several tmeporary sites, found more taxa and individuals in pools than in riffles. Consequently, in springtime even though the temporary rivers still present some riffles and pools, a major presence of pools taxa indicates a restructuring community adapting to intermittency and ephemerally of waters in summertime.

From our results, differences in taxa richness and abundance between riffles and pools (R vs. L) seem to be more important than temporality and consequently more convergent between med-regions. In fact, Delucchi (1988) noticed higher differences between riffles and pools than temporary and permanent sites, which agree with our results.

All sampled med-regions have a similar number of taxa between riffle and lentic habitats, as has been shown in elsewhere (e.g., Scullion et al., 1982; Logan & Brooker, 1983, but see Brown & Brussock, 1991; Boulton & Lake, 1992a or McCulloch, 1986). However, composition differs between habitats looking at order or family assemblage, as has been reported by many studies (e.g., Logan & Brooker, 1983). Riffles in mediterranean rivers hold a higher number of Ephemeroptera, Plecoptera and Trichoptera in front of Odonata, Coleoptera and Heteroptera, because of their different hydraulic requirements (Statzner et al., 1988). This pattern has also been observed in other areas far away from mediterranean regions, as in a more temperate area in England (Scullion et al., 1982) or more arid in North America (McCulloch, 1986). Strong similarities are also observed looking at the exclusive fauna found in riffles and pools, when our results are compared with other studies around the world (e.g., see Rabeni & Minshall, 1977; Armitage et al., 1974; Scullion et al., 1982; McCulloch, 1986; Malmqvist et al., 1993), what would indicate that differences between habitats in macroinvertebrate community are independent of mediterranean climate. However, climatic patterns can influence in the proportion of riffles and pools in a reach along the year, loosing riffles as the drought is coming (Gasith & Resh, 1999; Bonada et al., Chapter 4).

Regional and local richness in mediterranean rivers

In any region, richness of plants or animals is a product of the balance between processes acting at different scales (Ricklefs, 1987; Tonn *et al.*, 1990). Regional richness (by history factors) have been obtained using all accumulated taxa recorded in each med-region, whereas local richness (by ecological factors) may be referred to the number of taxa collected per sampling site.

Traditionally, stream ecologists have focused on the Thienemann's principles (Thienemann, 1954) where richness is a function of frequency and magnitude of physical disturbance, indicating a highest richness at intermediate levels of disturbance (e.g., Minshall, 1988; Sousa, 1994). Poff & Ward (1989) classified North American streams according to degrees of

intermittency, flood frequency and predictability and flow predictability suggesting highest richness under high flow predictability by low in high flood and drought frequency. Mediterranean rivers, are characterized by predictable disturbances (floods and droughts) (Gasith & Resh, 1999) what would imply a higher regional richness compare with other climatic regions where floods and droughts are more extreme and less predictible. Consequently, although the intensity and frequency of this discharge disturbance is impredectible (Vidal-Abarca, 1990; Gasith & Resh, 1999) regional richness found in medregions (as number of family taxa) appears is very high, except for SWAustralia, where historical events constrained a major number of taxa (Bunn & Davies, 1990). Values of richness obtained from this study are difficult to compare with other climatic areas because no comparative studies between climates have been performed. However, some extensive studies performed elsewhere may be useful. Thus, for example, an extensive study in NWAustralia, under two climates, 90 macroinvertebrate families were collected, 77 in the tropical wet-dry climate and 73 and 61 in the desert one (Kay et al., 1999), values lower than the found in med-regions. Studies in plant biology agree with this high biodiversity in med-regions (Raven 1973; Deacon, 1983), what exhibit that under intermediate stress conditions a high biodiversity can be hold. The explanation may be found in the predictability of events (floods and drougths) in the mediterranean climate (Gasith & Resh, 1999) and the presence of resilient and resistant adaptations to these disturbances by evolutionary forces acting since the begining of the earth history.

According to Ricklefs (1987) the relationship between regional and local richness indicate the relative importance of regional and local processess influencing local richness. Thus, if medregions with different regional richness have similar local richness, local processes are important, whereas if a positive relationship between regional and local richness is present, regional factors dominate. Figure 14 displays the relationship between regional and local richness in the sampled med-regions. The relationship is positive between SWAustralia, South Africa and California. Regional processes are important to determine local diversity in these areas because their different biogeographic history. However, although MedBasin presents the highest regional richness, it displays a highly variable intermediate local richness, indicating that there is a regulation of local richness by local factors. This phenomena does not agree with Shmida (1981) who comparing vegetation of chaparral and coastal sage between Israel and southern California. MedBasin region present a changing topography, landscapes and microclimates (see Plate 2) that originate numerous river types (e.g., ramblas, short and nival-influence rivers, karsts,...). In this situation, different local factors (biotic or abiotic) can act more or less severely modulating richness and making it highly variable between sites. On the other hand, this high variability in river tipology and several historical events may be the responsibles to the highest value of regional richness found in MedBasin. Thereby, the presence of such tipologies could allow the presence and survival of some specific taxa and at the same time modulate local richness.



Figure 14. Relationship between regional and local richness. In Figure 14a, model from Ricklefs (1987) is presented. In Figure 14b the application of the model to collected data in med-regions and the approximate curve is plot.

Other convergences and divergences between med-regions

Convergences and divergences between med-regions have been established looking at the community structure. However, communities may display different structure in distant regions, but present similar patterns in feeding or behavioural strategies. Because unavailable data in some med-regions, we were not able to compare trophic relationships under similar climatic conditions. However, some observations can be made from our study. For example, in southern Hemisphere, where the predators Rhyacophilae lack by historical factors, they a replaced by other macroinvertebrates with similar feeding requirements, as the Corydalidae in South Africa. However, in some cases, some feeding strategies are not replaced because again, local factors constrain them. That is the case of South Africa, where its oligotrophic and acidic waters (King *et al.*, 1979) constrain the presence of grazers in headwaters, appearing downstream.

Concluding remarks

Historical and ecological factors have been used in our study to check for differences between macroinvertebrate communities in med-regions. In all cases, northern hemisphere regions have strong convergences compared with the rest of med-regions, and a closer similarity with South Africa than SWAustralia is noticed. Scale of study and the interchange of information between scales have been a problem for ecologists (Levin, 1992). Difficulties are found to define and establish what are large and small scales (Peckarsky et al., 1997). We argue that under the same climatic constrains both historical and local factors are important to macroinvertebrate communities. Convergences or divergences in historical and ecological factors among med-regions let us to understand convergences or divergences in macroinvertebrate community, although other factors as the incidence of non-native macroinvertebrates could affect similarities and differences between med-regions. Consequently, in reference conditions, history, climate, landscape, temporality, habitat and biotic interactions can be considered as filters acting as mechanistic factors in a hierarchical direction, allowing the presence and enhancing the abundance of several species in front of others (Tonn et al., 1990; Poff, 1997). Comparisons in common taxa between med-regions suggest that historical and ecological factors could be more important than climate to determine the abundance of several taxa in a specific habitat or flow condition. However, when such factors let the abundance of some taxa in several regions (e.g., the high pH in California and MedBasin enhance the presence of Mollusca), responses to habitat and temporality between med-regions are highly convergent. From our data, more similarities have been observed between pairs of med-regions in habitat than in flow conditions, suggesting that habitat has more effect on the macroinvertebrate community than temporality, as has been pointed out by Delucchi (1998).

Responses to temporality and habitat in mediterranean rivers suggest the existence of an habitat template (Southwood, 1977; Poff & Ward, 1990; Townsend and Hildrew, 1994) modulated by mediterranean climate, where the evolution acted establishing adaptative convergences in macroinvertebrates between med-regions, whereas divergences may be related to topographic, litologic or historical factors. Further studies focused on biological or ecological species traits should be done to provide more information about similar response under common flow disturbances and habitat heterogeneity independently of similarities and differences between macroinvertebrate assemblages. Moreover, comparative information about the annual and interannual changes of macroinvertebrate community between med-regions would help us to understand better the identity and uniqueness of mediterranean rivers.

REFERENCES

- ALBA-TERCEDOR, J.; GONZÁLEZ, G. & PUIG, M. A. (1992). Present level of knowledge regarding fluvial macroinvertebrate communities in Spain. *Limnetica*, 8: 231-241.
- ALEXANDER, W. J. R. (1985). Hydrology of low latitude Southern Hemisphere land masses. *Hydrobiologia*, 125: 75-83.
- ARMESTO, J. J.; LEÓN-LOBOS, P. & KALIN-ARROYO, M. K. (1997). Los bosques templados del sur de Chile y Argentina: Una isla biogeográfica. In: Armesto, J. J.; Villagrán, C. & Kalin-Arroyo, M. K. (eds.). *Ecología de los bosques nativos de Chile*. Editorial Universitaria. Santiago de Chile. 23-28pp.

ARMITAGE, P. D.; MACHALE, A. M. & CRISP, D. C. (1974). A survey of stream invertebrates in the Cow Green basin (Upper Teesdale) before inundation. *Freshwat. Biol.*, 4: 369-398. ASCHMANN, H. (1973).

a. Distribution and peculiarity of Mediterranean Ecosystems. In: Di Castri, F. & Mooney, H. A. (eds). *Mediterranean Type Ecosistems: Origin and structure*. Springer-Verlag, New York. 405 pp.

b. Man's impact on the several regions with Mediterranean climates. In: Di Castri, F. & Mooney, H. A. (eds). *Mediterranean Type Ecosistems: Origin and structure*. Springer-Verlag, New York. 405 pp.

ATTIWILL, P. M.; GUTHRIE, H. B.; LEUNING, R. (1978). Nutrient cycling in Eucaliptus oblique (L'Herit.) forest. I. Litter production and nutrient return. *Aust. J. Bot.*, 26: 79-91.

- AXELROD, D. I. (1973) History of the mediterranean ecosystem in California. In: Di Castri, F.
 & Mooney, H. A. (eds). *Mediterranean Type Ecosistems: Origin and structure*. Springer-Verlag, New York. 405 pp.
- BADRI, A.; GUIDICELLI, J. & PRÉVOT, G. (1987). Effets d'une crue sur la communauté d'invertébrés benthiques d'une rivière méditerranéenne, Le Rdat (Maroc). Acta Oecologica, 8(4): 481-500.
- BAILEY, R. G. (1995). Description of the ecoregions of the United States. Forest Service, US Department of Agriculture. Publication 1391. Washington DC.

- BALL, I. R. (1975). Nature and formulation of Biogeographic hypothesis. Syst. Zool., 24: 407-430.
- BARBOUR, M. T.; GERRITSEN, J.; SNYDER, B. D. & STRIBLING, J. B. (1999). Rapid Bioassessment Protocols for Use in Streams and Wadeable Rivers: Periphyton, Benthic Macroinvertebrates and Fish, Second Edition. EPA 841-B-99-002. US EPA; Office of Water; Washington, D.C.
- BEARD, J. S. (1983). Ecological control of the vegetation of south-western Australia: moisture versus nutrients. In: Kruger, F. J.; Mitchell, D. T. & Jarvis, J. U. M. (eds.). Mediterranean-type Ecosystems. The Role of Nutrients. Springer-Verlag, Berlin.
- BECKINSALE, R. P. (1969). River Regimes. In: Chorley, R. J. Water, Earth, and Man: A synthesis of hydrology, geomorphology, and socio-economic geography. Methuen & CO Ltd. 455-471pp.
- BENEDETTO, L. (1974). Clave para la identificación de los Plecópteros Sudamericanos. Stud. N. the Neotr. Fauna, 9: 141-170.
- BILANDZIJA, J.; FRANKOVIC, M. & KAUCIC, D. (1998). The Croatian Adriatic coast. In: Conacher, A. J. & Sala, M. Land degradation in Mediterranean environments of the world: Nature and extent cause and solutions. John Wiley & Sons. 491 pp.
- BONADA, N.; PRAT, N.; MUNNÉ, A.; RIERADEVALL, M.; ALBA-TERCEDOR, J.; ÁLVAREZ, M.; AVILÉS, J.; CASAS, J.; JÁIMEZ-CUÉLLAR, P.; MELLADO, A.; MOYÀ, G.; PARDO, I.; RAMÓN, G.; SUÁREZ, M. L.; TORO, M.; VIDAL-ABARCA, M. R.; VIVAS, S. & ZAMORA-MUÑOZ, C. (in press).

a. Ensayo de una tipología de las cuencas mediterráneas del proyecto GUADALMED1 siguiendo las directrices de la Directiva Marco del Agua. *Limnética*.

b. Criterios para la selección de condiciones de referencia en los ríos mediterráneos. Resultados del proyecto GUADALMED1. *Limnética*.

BOULTON, A. J. & BROCK, M. A. (1999). Australia Freshwater Ecology. Processes and Management. Gleneagles Publishing. Adelaide. 300pp.

BOULTON, A. J. & LAKE, P. S. (1992)

a. The ecology of two intermittent streams in Victoria, Australia. II. Comparisions of faunal composition between habitats, rivers and years. *Freshwat. Biol.*, 27: 99-121.

b. The ecology of two intermittent streams in Victoria, Australia. III. Temporal changes in faunal composition. *Freshwat. Biol.*, 134: 27-52.

- BOULTON, A. J. & SUTER, P. J. (1986). Ecology of temporary streams an Australian perspective. In: de Decker, P. & Williams, W. D. (eds.). Limnology in Australia. CSIRO/Dr. W. Junk. Melbourene/Dordrecht. 313-327.
- BRADBURY, D. E. (1981). The physical geography of the mediterranean lands. In: di Castri, F.;
 Goodall, D. W. & Specht, R. L. Ecosystems of the world 11: Mediterranean-Type Shrublands. Elsevier Scientific Publishing Company. 643 pp.
- BROWN, C. A. (1978). Freshwater molluscs. In: Werger, M. J. A. (ed.). *Biogeography and Ecology of Southern Africa*. Dr. Junk Publishers. The Hague. 1153-1180pp.
- BROWN, A. V. & BRUSSOCK, P. P. (1991). Comparision of Benthic invertebrates between riffles and pools. *Hydrobiologia*, 220: 99-108.
- BROWN, C. A. & DALLAS, H. F. (1995). *Eerste River, Western Cape: Situation Assessment of the Riverine Ecosystem.* CSIR-Stellenbosch. Final Report.
- BRUNDIN, L. (1965). On the real nature of transanteartic relationships. *Evolution*, 19: 496-505.
- BUNN, S. E. & DAVIES, P. M. (1990). Why is the stream fauna of south-western Australia so impoverished? *Hydrobiologia* 194: 169-176.
- BUNN, S. E. (1988). Life histories of some benthic invertebrates from streams of the Northern Jarrah Forest, Western Australia. *Aust. J. Mar. Freshwater Res.*, 39: 785-804.
- BUNN, S. E., EDWARD, D. H. & LONERAGAN, N. R. (1986). Spatial and temporal variation in the macroinvertebrate fauna of streams of the northern jarrah forest, Western Australia: community structure. *Freshwat. Biol.*, 16: 67-91.
- CAMARASA, A. M. & SEGURA, F. (2001). Flood events in Mediterranean ephemeral streams

(ramblas) in Valencia region, Spain. Catena, 45: 229-249.

- CAMPOS, H. (1985). Distribution of the fishes in the Andean rivers in the South Chile. Arch. Hydrobiol., 104(2): 169-191.
- CASEY, H. & LADLE, M. (1976). The chemistry and biology of the South Winterbourne, Dorset, England. *Freswat. Biol.*, 6: 1-12.
- CHURCHWARD, H. M. & DIMMOCK, G. M. (1989). The soild and landforms of the nothern jarrah forest. In: B. Dell, J. J. Havel & N. Malajezuk (eds). *The Jarrah Forest. A Complex Mediterranean Ecosystem*. Kluwer Academic Publishers. Dordrecht: 13-21.
- CODY, M. & MOONEY, H. (1978). Convergence versus nonconvergerce in mediterranean climate ecosystems. Ann. Rev. Ecol. And Syst., 9: 265-321.
- CODY, M. L. (1973). Parallel evolution and bird niches. In: Di Castri, F. & Mooney, H. A. (eds). *Mediterranean Type Ecosistems: Origin and structure*. Springer-Verlag, New York. 405 pp.
- CONACHER, A. & CONACHER, J. (1998). Southern Australia. In: Conacher, A. J. & Sala, M. Land degradation in Mediterranean environments of the world: Nature and extent cause and solutions. John Wiley & Sons. 491 pp.
- CONACHER, A. J. & CONACHER, J. (1998). Introduction. In: Conacher, A. J. & Sala, M. Land

degradation in Mediterranean environments of the world: Nature and extent cause and

solutions. John Wiley & Sons. 491 pp.

CONACHER, A. J. & SALA, M. (1998) Land degradation in Mediterranean environments of the world: Nature and extent cause and solutions. John Wiley & Sons. 491 pp.

CONACHER, A. J. (1995). Definition of Mediterranean climates. Geojournal, 36: 298.

- COOPER, S. D.; DUDLEY, T. L. & HEMPHIL, N. (1986). The biology of chaparral streams in Southern California. In: DeVries (ed.). Proc. Chaparral Ecosystem Research Conference. Report n°62. California Water Resources Center. Davis: 139-151.
- COWLING, R. M. (1987). Fire and its role in coexistence and speciation in Gondwanan shrublands. *South African Journal of Science*, 83: 106-112.
- COWLING, R. M. (ed.) (1992). The ecology of fynbos: nutrients, fire and diversity. Oxford university Press, Cape Town.
- COX, C. B. & MOORE, P. D. (1993). *Biogeography: An ecological and evolutionary approach*. Fifth Edition. Blackwell Science. 326 pp.
- DAGET, P. H. & MICHEL-VILLAG, J. P. (1975). Délimitation de la region méditerranéenne selon les régimes des précipitations. In: Proceedings of Symposium Israel-France: ecological research and development of arid zones with winter precipitation. Special Publication, 39. Volcani Centre Bet-Dagan, Israel. 3-13pp.
- DAGET, PH.; AHDALI, L. & DAVID, P. (1988). Mediterranean bioclimate and its variation in the paleartic region. In: Specht, R. L. (ed.). *Mediterranean-type ecosystems: A data source book.* Kluwer Academic Publishers. Dordrecht. 248pp.
- DALLMAN, P. R. (1998). *Plant life in the world's mediterranean climates*. California Native Plant Society. University of California Press. Berkeley and Los Angeles. 257pp.
- DAVIES, B. R.; O'KEEFFE, J. H. & SNADDON, C. D. (1993). A synthesis of the ecological functioning, conservation and management of South African river ecosystems. Water Res. Commission Rep. TT62/93. South Africa: Pretonia.
- DAY, J. A.; STEWART, B. A.; DE MOOR, I. J. & LOUW, A. E. (2001). Guides to the Freshwater invertebrates of Southern Africa, 4: Crustacea III. Water Research Commission, 141 pp.
- DE MOOR, I. J. & BRUTTON, M. N. (1988). Atlas of alien and translocated indigenous aquatic animals in Southern Africa. Port Elizabeth, South African National Scientific Programmes. Report nº144: 1-310.
- DE MOOR, F. C. (1988). Pristine river systems and caddisflies in South Africa. *The Naturalist*, 32(2): 30-37.

DE MOOR, F. C. (1992).

a. Factors influencing the distribution of Trichoptera in South Africa. *Proc.* 7th Int. Symp. Trichoptera: 51-58.

b. Factors influencing the establishment of aquatic insect invaders. *Trans. Roy. Soc. S. afr.*, 48(1): 141-158.

- DE MOOR, F. C. (1997). An unusual caddisfly larva from South Africa, a posible member of the Triplectidinae (Trichoptera: Leptoceridae). *Proc.* 8th Int. Symp. Trichoptera, 323-330.
- DEACON, H. J. (1983). Comparative evolution of Mediterranean-type ecosystems: A sourthern perspective. In: Kruger, F. J.; Mitchell, D. T. & Jarvis, J. U. M. (eds.). *Mediterranean-type Ecosystems. The Role of Nutrients.* Springer-Verlag, Berlin.
- DEDECKKER, P. (1986). What happened to the Australian aquatic biota 18000 years ago?. In: DeDeckker, P. & Williams, W. D. (eds.). *Limnology in Australia*. CSIRO. Melbourne. 487-496.
- DEL ROSARIO, R. B. & RESH, V. H. (2000). Invertebrates in intermittent and perennial streams: Is the hyporheic zone a refuge from drying?. J. N. Am. Benthol. Soc., 1984): 680-696.
- DELL, B.; HOPKINS, A. J. M. & LAMONT, B. B. (1986). Introduction. In: Dell, B.; Hopkings, A. J. M. & Lamont, B. B. (eds.). *Resilience in mediterranean-type ecosystems*. Dr. W. Junk Publishers. 168 pp.
- DELUCCHI, C. M. (1988). Comparison of community structure among streams with different temporal flow regimes. *Can. J. Zoology*, 66: 579-586.
- DELUCCHI, C. M. & PECKARSKY, B. L. (1989). Life history patterns of insects in an intermittent and permanent stream. J. N. Am. Benthol. Soc., 8(4): 308-321.
- DEPARTMENT OF LAND AFFAIRS. (1994). Illustrated atlas of Southern Africa. Institute of Geographical Analysis. 232 pp.
- DI CASTRI, F. & HADLEY, M. (1985) Enhancing the credibility of ecology: can research be made more comparable and predictive?. *GeoJournal*, 11: 321-338.
- DI CASTRI, F. & MOONEY, H. A. (1973). Animal biogeography and ecological niche. In: Di Castri, F. & Mooney, H. A. (eds). *Mediterranean Type Ecosistems: Origin and structure*. Springer-Verlag, New York. 405 pp.
- DI CASTRI, F. (1973)

a. Climatographical comparisions between Chile and the Western Coast of North America. In: Di Castri, F. & Mooney, H. A. (eds). *Mediterranean Type Ecosistems: Origin and structure*. Springer-Verlag, New York. 405 pp.

b. Soil animals in latitudinal and topographical gradients of Mediterranean ecosystems. In: Di Castri, F. & Mooney, H. A. (eds). *Mediterranean Type Ecosistems: Origin and structure*. Springer-Verlag, New York. 405 pp.

c. Mediterranean-type shrublands of the world. In: Di Castri, F. & Mooney, H. A. (eds). *Mediterranean Type Ecosistems: Origin and structure*. Springer-Verlag, New York. 405 pp.

- DI CASTRI, F. (1981). Mediterranean-type shrublands of the world. In: di Castri, F.; Goodall,
 D. W. & Specht, R. L. *Ecosystems of the world 11: Mediterranean-Type Shrublands*. Elsevier Scientific Publishing Company. 643 pp.
- DI CASTRI, F. (1991). An ecological overview of the five regions of the world with a mediterranean climate. In: Groves, R. H. & di Castri, F. (eds). *Biogeography of Mediterranean Invasions*. Cambridge University Press. Cambridge. 485pp.
- DOADRIO, I. (2001). Atlas y libro rojo de los peces continentals de España. Museo Nacional de Ciencias Naturales. Madrid. 364pp.
- DOLÉDEC, S. (1989). Seasonal dynamics of benthic macroinvertebrate communities in the Lower Ardèche River (France). *Hydrobiologia*, 182: 73-89.

DRUDE, O. (1890). Hanbuch der Pflanzengeographie. J. Enfelhorn. Stuttgart.

DUDLEY, T. L.; COOPER, S. D. & HEMPHILL, N. (1986). Effects of macroalgae on a stream invertebrate community. J. N. Am. Benthol. Soc., 5(2): 93-106.

- DUFRÊNE, M. & LEGENDRE, P. (1997). Species assemblages and indicator species: the need for a flexible asymmetrical approach. *Ecological Monographs*, 67:345-366.
- EDMUNDS, G. F. Jr. (1972). Biogeography and evolution of Ephemeroptera. Annual Review of Entomolgy, 17: 21-42.
- EMBERGER, L. (1930). La végétation de la región méditerranéenne, essai d'une classification des groupements végétaux. *Rev. Gén. Bot.*, 42: 641-662, 705-721.
- EMBERGER, L. (1955). Une classification biogéographique des climates. *Rev. Trav. Lab. Bot. Fac. Sci., Montpellier,* 7: 3-43.
- EMBERGER, L. (1971). Travax de Botanique et d'Ecologie. Masson, Paris. 520pp.
- ENDLER, J. A. (1982). Problems in distinguishing historical from ecological factors in biogeography. *Amer. Zool.*, 22: 441-452.
- ERMAN, D. C.; ANDREWS, E. D. & YODER-WILLIAMS, M. (1988). Effect of winter floods on fishes in the Sierra Nevada. *Can. J. Fish. Aquat. Sci.*, 45: 2195-2200.
- EUROPEAN PARLIAMENT AND THE COUNCIL (2000). Establishing a framework for Community actions in the field of water policy. PECONS 3639/00.
- FERRANO, S. P. & COLE, F. A. 1992. Taxonomic level sufficient for assessing a moderate impact on macrobenthic communities in Puget Sound, Washington, USA. Can. J. Fish. Aquat. Sci., 49: 1184-1188.
- FIGUEROA, R. & ARAYA, E. (2002). Macroinvertebrados acuáticos del estero Nonguen. Publicacition in http://www.eula.cl/educa.htm. 15pp.
- FIGUEROA, R.; ARAYA, E. & VALDOVINOS, C. (2000). Deriva de macroinvertebrados bentónicos en un sector ritron: río Rucue, Chile Centro-Sur. *Bol. Soc. Biol.*. *Concepción*, 71: 23-32.
- FISHER, S. G.; GRAY, L. J.; GRIMM, N. B. & BUSCH, D. E. (1982). Temporal succession in a desert stream ecosystem following flash flooding. *Ecol. Monogr.*, 52: 93-110.
- FLINT, O. S. (1974). Checklist of the Trichoptera, or Caddisflies, of Chile. Rev. Chilena Ent., 8: 83-93.
- FOX, B. J. & FOX, M. D. (1986). Resilience of animal and plant communities to human disturbance. In: Dell, B.; Hopkings, A. J. M. & Lamont, B. B. (eds.). Resilience in mediterranean-type ecosystems. Dr. W. Junk Publishers. 168 pp.
- FUENTES, E. (1976). Ecological convergence of lizard communities in Chile and California. *Ecology*, 57: 3-17.
- FUGGLE, R. F. & ASHTON, E. R. (1979). Climate. In: Day, J. A.; Siegfried, W. R.; Louw, G. N. & Jarman, M. L. (eds.). *Fynbos Ecology: A preliminary syntesis*. South African National Scientific Programmes Report, 40. Council for Scientific and Industrial Research, Pretoria.
- GALLARDO-MAYENCO, A.; FRESNEDA, J. & TOJA, J. (1992). Distribución de los coleópteros acuáticos (Insecta: Coleoptera) en dos cuencas del sur de la P. Ibérica. Relaciones con algunos factores del medio. *Limnética*, 11(1): 19-28.
- GALLARDO-MAYENCO, A.; PRENDA, J. & PUJANTE, A. (1994). Influence of some environmental factors on the freshwater macroinvertebrates distribution in two adjacent river basins under mediterranean climate. *Arch. Hydrobiol.*, 131(4): 449-463.
- GALLARDO-MAYENCO, A.; PRENDA, J. & TOJA, J. (1998). Spatio-temporal distribution and ecological preferences of coexisting Hydropsychid species (Trichoptera) in two Mediterranean River basins (S Spain). *Internat. Rev. Hydrobiol.*, 83(2): 124-134.
- GARDNER, C. A. (1942). The vegetation of western Australia. With special reference to the climate and soils. *Journal of the Royal Society Western Australia*, 28: 11-87.
- GASITH, A. & RESH, V. H. (1999). Streams in Mediterranean climate region: Abiotic influences and biotic responses to predictable seasonal events. *Annu. Rev. Ecol. Syst.*, 30: 51-81.
- GASITH, A. (1992). Conservation and management of the coastal streams of Israel: an assessment of stream status and prospect of rehabilitation. In: *River Conservation and Management*, Eds. P. J. Boon, P. Calow, G. E. Petts, pp. 51-64. New York: Wiley. 470 pp.

- GENTILLI, J. (1989). Climate of the jarrah forest. In: B. Dell, J. J. Havel & N. Malajezuk (eds). The Jarrah Forest. A Complex Mediterranean Ecosystem. Kluwer Academic Publishers. Dordrecht: 23-40.
- GRAÇA, M. A. S.; FONSECA, D. M. & CASTRO, S. T. (1989). The distribution of macroinvertebrate communities in two Portuguese rivers. *Freshwat. Biol.*, 22: 297-308.
- GRAY, L. J. (1981). Species composition and life histories of aquatic insects I a lowland Sonoran desert stream. *Am. Mid.Nat.*, 106: 229-242.
- GRAY, L. J.; FISHER, S. G. (1981). Postflood recolonisation pathways of macroinvertebrates in a lowland Sonran desert stream. *Am. Mid. Nat.*, 106: 249-257.
- GRIMM, N. B. & FISHER, S. G. (1989). Stability of periphiton and macroinvertebrates to disturbance by flash floods in a desert stream. J. N. Am. Benthol. Soc., 8(4): 293-307.
- GRISEBACH, A. (1872). Die Vegetation der Erde nach ihrer klimatischen Anordnung. W. Engelmann, Leipzig. 603pp.
- GROVE, A. T. & RACKHAM, O. (2001). *The nature of Mediterranean Europe*. Yale University Press, New Haven & London. 384 pp.
- GRUBB, P. J. & HOPKINS, A. J. M. (1986). Resilience at the level of the plant community. In: Dell, B.; Hopkings, A. J. M. & Lamont, B. B. (eds.). *Resilience in mediterranean-type* ecosystems. Dr. W. Junk Publishers. 168 pp.
- GUIDICELLI, J.; DAKKI, M. & DIA, A. (1985). Caractéristiques abiotiques et hydrobiologiques des eaux courantes méditerranéennes. *Verh. Internal. Verein. Limnol.*, 22: 2094-2101.
- HABDIJA, I.; LAJTNER, J. & BELINIĆ, I. (1995). The contribution of gastropod biomass in macrobentic communities of a karstic river. *Int. Revue ges. Hydrobiol.*, 80: 103-110.
- HAJEK, E. R. (1991). Medio ambiente en Chile. In: Hajek, E. R. (ed.). La situación ambiental en América Latina. CIEDLA. Buenos Aires: 237-294.
- HALL, A. V. (1979). Invasive Weeds. In: Day, J. A.; Siegfried, W. R.; Louw, G. N. & Jarman, M. L. Fynbos Ecology: A preliminary syntesis. South African National Scientific Programmes Report, 40. Council for Scientific and Industrial Research, Pretoria.
- HARRISON, A. D. & AGNEW, J. D. (1962). The distribution of invertebrates endemic to acid streams in the western and Southern Cape Province. *Ann. Cape Prov. Mus.*, 2: 273-291.
- HARRISON, A. D. & BARNARD, K. H. (1972). The stream fauna of an isolated mountain massif, Table Mountain, Cape Town, South Africa. *Trans. Roy. Soc. S. Afr.*, 40(3): 135-153.
- HARRISON, A. D. (1978). Freshwater invertebrates (except molluscs). In: Werger, M. J. A. (ed.). Biogeography and Ecology of Southern Africa. Dr. Junk Publishers. The Hague. 1139-1152.
- HART, S. D. & HOWMILLER, R. P. (1975). Studies on decomposition of allochthonous detritus in two Southern California streams. *Verh. Int. Verein. Limnol.*, 19: 1665-1674.
- HEDIN, L. O. & CAMPOS, H. (1991). Importance of small streams in understanding and comparing watershed ecosystem processes. *Rev. Chil. Hist. Nat.*, 64(3): 583-596.
- HERRERA, C. M. (1995). Plant-vertebrate seed dispersal systems in the Mediterranean: Ecological, evolutionary, and historical determianents. *Annu. Rev. Ecol. Syst.*, 26: 705-727.
- HILL, M. O. (1979). TWINSPAN: a FORTRAN Program for Arranging Multivariate Data in An Ordered Two-way Table by Classification of the Individuals and Attributes. Ecology and Systematics, Cornell University. Ithaca. New York. 49 pp.
- HOLSTEIN, G. (1984). California riparian forests: Deciduous islands in an evergreen sea. In: Warner, R. E. & Hendrix, K. M. (eds.). California Riparian Systems: Ecology, Conservation and Productive Management. University of California Press. 1035 pp.
- HORNBECK, D.; KANE, PH. & FULLER, D. N. (1983). California patterns: a geographical and historical atlas. Mayfield Publishing Company. 117 pp.
- ILLIES, J. (1964). Südamerikanische Perlidae (plecoptera), besonders aus Chile und Argentinien. *Beiträge Neot. Fauna*, 3(3): 207-233.
- ILLIES, J. (1965). Neue Plecopteren aus Chile und Argentinien. Mitt. Schweiz. Entomol. Gesells., 37(3): 151-156.

- JÁIMEZ-CUÉLLAR, P.; VIVAS, S.; BONADA, N.; ROBLES, S.; MELLADO, A.; ÁLVAREZ, M.; AVILÉS, J.; CASAS, J.; ORTEGA, M.; PARDO, I.; PRAT, N.; RIERADEVALL, M.; SÁINZ-CANTERO, C.; SÁNCHEZ-ORTEGA, A.; SUÁREZ, M. L.; TORO, M.; VIDAL-ABARCA, M. R.; ZAMORA-MUÑOZ, C. & ALBA-TERCEDOR, J. (in press). Protocolo Guadalmed (PRECE). Limnética.
- JOHNSON, A. W. (1973). Historical view of the concept of ecosystem Convergence. In: Di Castri, F. & Mooney, H. A. (eds). *Mediterranean Type Ecosistems: Origin and structure*. Springer-Verlag, New York. 405 pp.
- KAHRL, W. L.; BOWEN, W. A.; BRAND, S.; SHELTON, M. L.; FULLER, D. L. & RYAN, D. A. (1978). The California water atlas. The Governor's Office of Planning and Researcj. Scramento. 118pp.
- KEELY, J. E. (1986). Resilience of mediterranean shrub communities to fires. In: Dell, B., Hopkins, A. J. M. & Lamont, B. B. (eds.). *Resilience in mediterranean-type ecosystems*. W. Junk, The Hague: 95-112 pp.
- KING, J. M. (1982). An ecological study of the macroinvertebrate fauna of the Eerste River, Western Cape Province, South Africa. Thesis. University of Cape Town. 222 pp.
- KING, J. M.; DAY, J. A. & VAN DER ZEL, D. W. (1979). Hydrology and hydrobiology. In: Day, J. A.; Siegfried, W. R.; Louw, G. N. & Jarman, M. L. Fynbos Ecology: A preliminary syntesis. South African National Scientific Programmes Report, 40. Council for Scientific and Industrial Research, Pretoria.
- KING, J. M.; DAY, J. A.; HURLY, P. R.; HENSHALL-HOWARD, M-P. & DAVIES, B. R. (1988). Macroinvertebrate communities and environment in a Souther African mountain stream. Can. J. Fish. Aquat. Scie., 45: 2168-2181.
- KÖPPEN, W. (1923). De klimate der Erde. Bornträger, Berlin. 369 pp.
- KOTLIAR, N. B. & WIENS, J. A. (1990). Multiple scales of patchiness and patch structure: A hierarchical framework for the study of heterogeneity. *Oikos*, 59: 253-260.
- KRUGER, F. J. (1979).

a. Plant Ecology. In: Day, J. A.; Siegfried, W. R.; Louw, G. N. & Jarman, M. L. *Fynbos Ecology: A preliminary syntesis.* South African National Scientific Programmes Report, 40. Council for Scientific and Industrial Research, Pretoria.

b. Introduction. In: Day, J. A.; Siegfried, W. R.; Louw, G. N. & Jarman, M. L. *Fynbos Ecology: A preliminary syntesis*. South African National Scientific Programmes Report, 40. Council for Scientific and Industrial Research, Pretoria.

c. Fire. In: Day, J. A.; Siegfried, W. R.; Louw, G. N. & Jarman, M. L. *Fynbos Ecology: A preliminary syntesis*. South African National Scientific Programmes Report, 40. Council for Scientific and Industrial Research, Pretoria.

- KRUGER, F. J.; MITCHELL, D. T. & JARVIS, J. U. M. (1983). *Mediterranean-type Ecosystems*. The Role of Nutrients. Springer-Verlag, Berlin.
- KUMMEROW, J. Comparative anatomy of sclerophylls of Mediterranean climatic areas. In: Di Castri, F. & Mooney, H. A. (eds). *Mediterranean Type Ecosistems: Origin and structure*. Springer-Verlag, New York. 405 pp.
- LAKE, P. S.; BARMUTA, L. A.; BOULTON, A. J.; CAMPBELL, I. C. & ST CLAIR, R. M. (1986). Australian streams and Northern Hemisphere stream ecology: comparisons and problems. *Proc. Ecol. Soc. Aust.*, 14: 61-82.
- LAMOUROUX, N.; POFF, N. L. & ARGERMEIER, P. L. (2002). Intercontinental convergence of stream fish community traits along geomorphic and hydraulic gradients. *Ecology*, 83(7): 1792-1807.
- LE HOUÉROU, H. N. (1990). Global change: Vegetation, ecosystems, and land use in the southern Mediterranean basin by the mid twenty-first century. *Isr. J. Bot.*, 39: 481-508.
- LEGENDRE, P. & LEGENDRE, L. (1998). *Numerical ecology*. Developments in environmental modelling 20. Elsevier. The Netherlands. 853pp.
- LEGIER, P. & TALIN, J. (1973). Comparaison de ruisseaux permanents et temporaries de la Provence calcarie. *Annls. Limnol.*, 9(3): 273-292.
- LEVIN, S. A. (1992). The problem of pattern and scale in ecology. Ecology, 73: 1943-1967.

- LOGAN, P. & BROOKER, M. P. (1983). The macroinvertebrate faunas of riffles and pools. *Water Res.*, 17(3): 263-270.
- LOUNACI, A.; BROSSE, S.; THOMAS, A. & LEK, S. (2000). Abundance, diversity and community structure of macroinvertebrates in an Algerian stream: the Sébaou wadi. *Annls. Limnol.*, 36(2): 123-133.
- MAAMRI, A.; CHERGUI, H. & PATTEE, E. (1994). Allochthonous input of coarse particulate organic matter to a Moroccan mountain stream. *Acta Oecologica*, 15(4): 495-508.
- MAJER, J. D. & GREENSLADE, P. (1988). Soil and litter invertebrates. In: Specht, R. L. (ed.). Mediterranean-type ecosystems: A data source book. Kluwer Academic Publishers. Dordrecht. 248pp.
- MALMQVIST, B.; NILSSON, A. N.; BAEZ, M.; ARMITAGE, P. D. & BLCKBURN, J. (1993). Stream macroinvertebrate communities in the island of tenerife. *Arch. Hydrobiol.*, 128(2): 209-235.
- MARCHANT, R., BARMUTA, L. A., & CHESSMAN, B. C. (1995). Preliminary study of the ordination of macroinvertebrate communities from running waters in Victoria, Australia. *Aust. J. Mar. Freshwat. Res.*, 45: 945-962.
- MARTÍNEZ-LÓPEZ, F.; ALFONSO, S. & PUJANTE, A. (1988). La malacofauna de la cuenca del río Mijares (Castellón, España): Distribución y aspectos ecológicos. *Iberus*, 8(2): 85-101.
- MENGE, B. A. & OLSON, A. M. (1990). Role of scale and environmental factors in regulation of community streucture. *TREE*, 5: 52-57.
- MCCULLOCH, D. L. (1986). Benthic macroinvertebrate distributions in the riffle-pool communities of two east texas streams. *Hydrobiologia*, 135: 61-70.
- MCCUNE, B. & MEFFORD, M. J. (1999). PC-ORD for Windows: Multivariate Analysis of Ecological Data v. 4.20. MjM Software, Gleneden Beach, Oregon.
- MCELRAVY, E. P.; LAMBERTI, G. A. & RESH, V. H. (1989). Year-to-year variation in the aquatic macroinvertebrate fauna of northern California stream. J. N. Am. Benthol. Soc., 8(1): 51-63.
- MIDGLEY, J. & SCHAFER, G. (1992). Correlates of water colour in streams rising in southern Cape catchments vegetated by fynbos and/or forest. *Water South Afri*ca, 18(2): 93-100.
- MILLER, P. C. (1983). Canopy structure of Mediterranean-type shrubs in relation to heat and moisture. In: Kruger, F. J.; Mitchell, D. T. & Jarvis, J. U. M. (eds.). *Mediterranean-type Ecosystems. The Role of Nutrients.* Springer-Verlag, Berlin.
- MILLER, A. M. & GOLLADAY, S. W. (1996). Effects of spates and drying on macroinvertebrate assemblages of an intermittent and a perennial prairie stream. J. N. Am. Benthol. Soc., 15(4): 670-689.
- MINSHALL, G. W. (1988). Stream ecosystem theory: a global perpective. J. North. Am. Benthol. Soc., 7(4): 263-288.
- MOLINA, C.; VIDAL-ABARCA, M. R. & SUÁREZ, M. L. (1994). Floods in arid south-east spanish areas: a historical and environmental review. In: Rossi, g. (ed.). *Coping with floods*. Kluwer academic publishers. The netherlands: 271-278.
- MOONEY, H. A. & DUNN, E. L. (1970). Convergent evolution of mediterranean-climate evergreen sclerophyll shrubs. *Evolution*, 24: 292-303.
- MOONEY, H. A. (1982). Mediterranean-type Ecosystems: Research progress and opportunities. South African Journal of Science, 78: 5-7.
- MOONEY, H. A., (1977). *Convergent Evolution in Chile and California*. Mediterranean Climate Ecosystems. Dowden, Hutchinson and Ross. Stroudsburg, Penn. 224pp.
- MORENO, J. L.; VIDAL-ABARCA, M. R. & SUÁREZ, M. L. (2001). Ecosistemas acuáticos de la Ramblas Murcianas. *Quercus*, 184: 28-33.
- MOUNT, J. F. (1995). *California Rivers and streams: The conflict between fluvial process and land use*. University of California Press, Berkeley. 359 pp.
- MOYLE, P. B. & HERBOLD, B. (1987). Life-history patterns and community structure in stream fishes of western North America: comparisions with eastern North America and Europe. In: Matthews, W. J. & Heins, D. C. (eds.). *Community and evolutionary ecology*

of North American stream fishes. University if Oklahoma Press. Oklahoma. USA. 25-32pp.

- MOYLE, P. B. & LEIDY, R. L. (1992). Loss of biodiversity in aquatic ecosystems: evidence from fish faunas. In: Fiedler, P. L. & Jain, S. K. (eds.). *Conservation biology: the theory and prectice of nature conservation, preservation, and management.* Chapman and Hall. New York.
- MOYLE, P. B. (1995). Conservation of native freshwater fishes in the mediterranean-type climate of California, USA: a review. *Biological Conservation*, 72: 271-279.
- MOYLE, P. B. & LIGHT, T. (1996). Fish invasions in California: do abiotic factors determine success: *Ecology*, 77(6): 1666-1670.
- MULLER, C. H. (1939). Relations of the vegetation and climatic types in nuevo León, Mexico. *Am. Mid. Nat.*, 21: 687-729.
- MUNNÉ, A., SOLÀ, C., y PRAT, N. (1998). QBR: Un índice rápido para la evaluación de la calidad de los ecosistemas de ribera. *Tecnología del Agua*, 175: 20-37.
- MUNNÉ, A., PRAT, N., SOLÀ, C., BONADA, N., & RIERADEVALL, M. (in press). A simple field method for assessing the ecological quality of riparian habitat in rivers and streams. QBR index. *Aquatic Conservation: Marine and Freshwater Ecosystems*.
- NAHAL, I. (1981). The Mediterranean climate from a biological viewpoint. In: di Castri, F.; Goodall, D. W. & Specht, R. L. Ecosystems of the world 11: Mediterranean-Type Shrublands. Elsevier Scientific Publishing Company. 643 pp.
- NAVEH, Z. & WHITTAKER, R. H. (1979). Structural and floristic diversity of shrublands and woodlands in Northern Israel and other Mediterranean areas. *Vegetatio*, 41(3): 171-190.
- NORRIS, R. H. 1995. Biological monitoring: the dilemma of data analysis. J. N. Am. Benthol. Soc., 14(3): 440-450.
- ORME, A. & ORME, A. J. (1998). Greater California. In: Conacher, A. J. & Sala, M. Land degradation in Mediterranean environments of the world: Nature and extent cause and solutions. John Wiley & Sons. 491 pp.
- ORSHAN, G. (1983). Approaches to the definition of mediterranean growth forms. In: Kruger, F. J.; Mitchell, D. T. & Jarvis, J. U. M. (eds.). *Mediterranean-type Ecosystems. The Role* of Nutrients. Springer-Verlag, Berlin. 86-100pp.
- PARDO, I.; ÁLVAREZ, M.; CASAS, J.; MORENO, J. L.; VIVAS, S.; BONADA, N.; ALBA-TERCEDOR, J.; JÁIMEZ-CUÉLLAR, P.; MOYÀ, G.; PRAT, N.; ROBLES, S.; SUÁREZ, M. L.; TORO, M. & VIDAL-ABARCA, M. R. (in press). El hábitat de los ríos mediterráneos. Diseño de un índice de diversidad de hábitat. *Limnética*.
- PASKOFF, R. P. (1973). Geomorphological processes and characteristics land-forms in the Mediterranean Regions of the World. In: Di Castri, F. & Mooney, H. A. (eds). *Mediterranean Type Ecosistems: Origin and structure*. Springer-Verlag, New York. 405 pp.
- PECKARSKY, B. L.; COOPER, S. D. & McINTOSH, A. R. (1997). Extrapolating from individual behaviour to populations and communities in streams. J. N. Am. Benthol. Soc., 16(2): 375-390.
- PIRES, A. M.; COWX, I. G. & COELHO, M. M. (2000). Benthic macroinvertebrate communities of intermittent streams in the middle reaches of the Guadiana Basin (Portugal). *Hydrobiologia*, 435: 167-175.
- POFF, N. L. & WARD, J. V. (1989). Implications of streamflow variability and predictability for lotic community structure: a regional analysis of streamflow patterns. *Can. J. Fish. Aquat. Sci.*, 46: 1805-1818.
- POFF, N. L. & WARD, J. V. (1990). Physical habitat template of lotic systems: recovery in the context of historical pattern of spatiotemporal heterogeneity. *Environmental Management*, 14: 629-645.
- POFF, N. L. (1997). Landscape filters and species traits: towards mechanistic understanding and prediction in stream ecology. J. N. Am. Benthol. Soc., 16(2): 391-409.

- PRAT, N. 1993. El futuro de los ríos españoles: secos o contaminados. Quercus. Febrero: 22-24
- PRAT, N. & MUNNÉ, A. (2000). Water use and quality and stream flow in a mediterranean stream. *Wat. Res.*, 34(15): 3876-3881.
- PRAT, N.; MUNNÉ, A.; RIERADEVALL, M.; SOLÀ, C. & BONADA, N. (2000). ECOSTRIMED: Protocol to stablish the ecological status of Mediterranean rivers and streams. Diputació de Barcelona. Àrea de Medi Ambient (Estudis de la Qualitat Ecològica dels Rius, 8). 94 pp.
- PRAT, N.; RIERADEVALL, M.; MUNNÉ, A.; SOLÀ, C. & BONADA, N. (1999). La qualitat ecológica del Llobregat, el Besòs i el Foix. Informe 1997. Diputació de Barcelona. Àrea de Medi Ambient (Estudis de loa Qualitat Ecològica dels Rius, 6). 154 pp.
- PRAT, N.; RIRERADEVALL, M.; MUNNÉ, A.; SOLÀ, C. & CHACÓN, G. (1997). La qualitat ecològica del Llobregat, el Besòs i el Foix. Informe 1996. Diputació de Barcelona. Àrea de Medi Ambient (Estudis de loa Qualitat Ecològica dels Rius, 2). 153 pp.
- PRAT, N.; MUNNÉ, A.; RIERADEVALL, M.; CARCELLER, F.; FONS, J.; CHACÓN, G.; IBÁÑEZ, J.; FONT, X.; CARMONA, J. M. & ROMO, A. (2000a). Biodiversity of a Mediterranean stream drainage network. Verh. Internat. Verein. Limnol., 27: 135-139.
- RABENI, C. F. & MINSHALL, G. W. (1977). Factors affecting microdistribution of stream benthic insects. *Oikos*, 29: 33-43.
- RAVEN, P. H. (1973). The evolution of Mediterranean floras. In: Di Castri, F. & Mooney, H. A. (eds). Mediterranean Type Ecosistems: Origin and structure. Springer-Verlag, New York. 405 pp.
- RESH, V. H.; BROWN, A. V.; COVICH, A. P.; GURTZ, M. E.; LI, H. W.; MINSHALL, G. W.; REICE, S. R.; SHELDON, A. L.; WALLACE, J. B. & WISSMAR, R. (1988). The role of disturbance in stream ecology. J. N. Am. Benthol. Soc., 7: 433-455.
- RESH, V. H. & SOLEM, J. O. (1996). Phylogenetic relationships and evolutionary adaptations of aquatic insects. In: Merrit, R. W. & Cummins, K. W. (eds.). An introduction to the Aquatic Insects of North America, 3rd ed. Kendall/Hunt Publishing Company. Dubuque, Iowa. 98-107pp.
- RESH, V. H.; JACKSON, J. F. & MCELRAVY, E. P. (1990). Disturbance, annual variability, and lotic benthos: examples from a California stream influenced by a mediterranean climate. *Memorie Inst, Ital. Idrobiol.*, 47: 309-329.
- RESH, V. H. & UNZICKER, J. D. 1975. Water quality monitoring and aquatic organisms: the importance of species identification. *J. Water Pollut. Control Fed.*, 47: 9-19.
- RICKLEFS, R. E. (1987). Community diversity: relative roles of local and regional processes. *Science*, 235: 167-171.
- RIERADEVALL, M. & PRAT, N. (1997). La qualitat biològica de les aigües del Parc Natural de Sant Llorenç del Munt i Serra de l'Obac. Diputació de Barcelona. Report. 75pp.
- RIERADEVALL, M.; BONADA, N. & PRAT, N. (in press). Community structure and water quality in the mediterranean streams of a natural park (St. Llorenç del Munt, NE Spain). *Limnética.*
- ROBLES, S.; TORO, M.; NUÑO, C.; AVILÉS, J.; ALBA-TERCEDOR, J.; ÁLVAREZ, M.; ROBLES, S.; CASAS, J.; JÁIMEZ-CUÉLLAR, P.; MELLADO, A.; MUNNÉ, A.; PARDO, I.; PRAT, N.; SUÁREZ, M. L. & VIDAL-ABARCA, M. R. (in press). Descripción de las cuencas mediterráneas seleccionadas en el proyecto GUADALMED. *Limnética*.
- ROBY, K. B. & AZUMA, D. L. (1995). Changes in a reach of a Northern California streams following wildfire. *Environmental Management*, 19(4): 591-600.
- RUNDEL, P. W. (1988). Leaf structure and nutrition in mediterranean-climate sclerophylls. In: Specht, R. L. (ed.). *Mediterranean-type ecosystems: A data source book*. Kluwer Academic Publishers. Dordrecht. 248pp.
- RUTT, G. P.; PICKERING, T. D. & REYNOLDS, N. R. M. 1993. The impact of livestock farming on Welsh streams: The development and testing of a rapid biological method for use in the assessment and control of organic pollution from farms. *Envir. Pollut.*, 81: 217-228.

- SABATER F., GUASCH, H., MARTÍ, E., ARMENGOL, J. & SABATER, S. (1995). The river Ter: Mediterranean river case-study in Spain. In: *River and stream ecosystems*, Eds. C. E. Cushing, K. W. Cummins & G. W. Minshall, pp. 419-438. Elsenvier. Amsterdam. 817 pp.
- SAGE, R. D. (1973). Ecological convergence of the lizerd faunas of the Chaparral Communities in Chile and California. In: Di Castri, F. & Mooney, H. A. (eds). *Mediterranean Type Ecosistems: Origin and structure*. Springer-Verlag, New York. 405 pp.
- SAMUELS, C. L. & DRAKE, J. A. (1997). Divergent perspectives on community convergence. *TREE*, 12(11): 427-432.
- SCHIMPER, A. F. W. (1898). Pflanzengeographic ayf physiologischer Grundlage. G. Fisher, Jena. 877pp.
- SCHWARZ, A. E. & SCHWOERBEL, J. (1997). The aquatic processing of sclerophyllous and malacophyllous leaves on a Mediterranean island (Corsica): spatial and temporal pattern. Annl. Limnol., 33(2): 107-119.
- SCULLION, J.; PARISH, C. A.; MORGAN, N. & EDWARDS, R. W. (1982). Comparison of benthic macroinvertebrate fauna and substratum composition in riffles and pools in the impounded River Elan and the unregulated River Wye, mid-Wales. *Freshwat. Biol.*, 12: 579-595.
- SHMIDA, A. (1981). Mediterranean vegetation in California and Israel: similarities and differences. *Isr. J. Bot.*, 30: 105-123.
- SMITH, M. J., KAY, W. R., EDWARD, D. H. D., PAPAS, P. J., RICHARSON, K. ST J., SIMPSON, J. C., PINDER, A. M., CALE, D. J., HORWITZ, P. H. J., DAVIS, J. A., YUNG, F. H., NORRIS, R. H. & HALSE, S. A. (1999). AusRivAS: using macroinvertebrates to assess ecological condition of rivers in Western Australia. *Freshwat. Biol.*, 41: 269-282.
- SOMMERHÄUSER, M.; ROBERT, B. & SCHUHMACHER, H. (1997). Fligth periods and life history strategies of caddisflies in temporary and permanent woodland brooks in the Lower Rhine area (Germany). In: Holzental, R. H. & Flint, O. J. (eds.). Proceedings of the 8th International Symposium on trichoptera. Ohio Biological Survey, Columbus, Ohio. 425-433.
- SOTO, D. & CAMPOS, H. (1997). Los lagos oligotróficos dels bosque templado húmedo del sur de Chile. In: Armesto, J. J.; Villagrán, C. & Kalin-Arroyo, M. K. (eds.). *Ecología de los bosques nativos de Chile*. Editorial Universitaria. Santiago de Chile. 317-334.
- SOUSA, W. P. (1984). The role of disturbance in natural communities. Ann. Rev. Ecol. Syst., 15: 353-391.
- SOUTH AFRICAN WEATHER BUREAU (1996). The weather and climate of the extreme South-Western Cape. Department of Environmental Affairs and Tourism. 33 pp.
- SOUTHWOOD, T. R. E. (1977). Habitat, the templet for ecological strategies?. J. Anim. Ecol., 46: 337-365.
- SPECHT, R. L. (1973). Structure and functional response of ecosystems in the mediterranean climate of Australia. in: di Castri, F. & Mooney, H. A. (eds.). *Mediterranean Type Ecosystems. Origin and Structure*. Springer-Verlag. Berlin. 113-120pp.
- SPECHT, R. L. (1979). Heathlands and related shrublands of the world. In: Specht, R. L. (ed.). *Ecosystems of the world*, 9A. *Heathlands and related shrublands. Descriptive studies*. Elsevier, Amsterdam. 1-18pp.
- SPETCH, R. L. & MOLL, E. J. (1983). Mediterranean-type heathlands and sclerophyllous shrublands of the world: an overview. In: Kruger, F. J.; Mitchell, D. T. & Jarvis, J. U. M. (eds.). *Mediterranean-type Ecosystems*. The Role of Nutrients. Springer-Verlag, Berlin.
- SPECTH, R. L. & Rayson, P. (1957). Dark Island heath (Ninethy-Mile Plain, South Australia). I. Definition of the ecosystem. *Aust. J. Bot.*, 5: 52-85.
- STAMOU, G. P. (1998). Arthropods of Mediterranean-Type Ecosystems. Springer-Verlag. Berlin. 135pp.
- STANFORD, J. A. & WARD, J. V. (1983). Insect species diversity as a function of environmental variability and disturbance in stream systems. In: Barnes, J. R. &

Minshall, G. W. (eds.). Stream Ecology: Application and testing of general ecological theory. 399pp.

- STANLEY, E. H. & FISHER, S. G. (1992). Intermittency, disturbance, and stability in stream ecosystems. In: Robarts, R. D. & Bothwell, M. L. (eds.). Aquatic Ecosystems in semiarid regions: implications for resource management. N.H.R.I. Symposium Series 7, Environmental Canada. Saskatoon. 271-280pp.
- STATSOFT, INC. (1999). STATISTICA for Windows (Computer Program Manual). Tulsa, UK.
- STATZNER, B.; GORE, J. A. & RESH, V. H. (1988). Hydraulic stream ecology: observed patterns and potential applications. J. N. Am. Benthol. Soc., 7(4): 307-360.
- SUÁREZ, M. L. & VIDAL-ABARCA, M. R. (2000). Aplicación del índice de calidad del bosque de ribera, QBR (Munné *et al.*, 1998) a los cauces fluviales de la cuenca del río Segura. *Tecnología del Agua*, 201: 33-45.
- SUÁREZ, M. L.; VIDAL-ABARCa, M. R.; SÁNCHEZ-MONTOYA, M.; ALBA-TERCEDOR, J.; ÁLVAREZ, M.; AVILÉS, J.; BONADA, N.; CASAS, J.; JÁIMEZ-CUÉLLAR, P.; MUNNÉ, A.; PARDO, I.; PRAT, N.; Rieradevall, M.; SALINAS, M. J.; TORO, M. & VIVAS, S. (in press.). Las riberas de los ríos mediterráneos y su calidad: El uso del índice QBR. *Limnética*.

SUC, J. P. (1984). Origin and evolution of the Mediterranean vegetation and climate in Europe. *Nature*, 307: 429-432.

Nature, 307: 429-432.

- TER BRAAK, J. F. (1998). CANOCO for Windows vs. 4.0. Centre for Biometry Wageningen. CPRO-DLO. The Netherlands.
- THIENEMANN, A. (1954). Eir drittes biozonotisches Grundprinzip. Arch. Hydrobiol., 49(3): 421-422.
- THROWER, N. J. W. & BRADBURY, D. E. (1973). The physiography of the mediterranean lands with special emphasis on California and Chile. In: Di Castri, F. & Mooney, H. A. (eds). *Mediterranean Type Ecosystems: Origin and structure*. Springer-Verlag, New York. 405 pp.
- TONN, W. M.; MAGNUSSON, J. J.; RASK, M. & TOIVONEN, J. (1990). Intercontinental comparison of small-lake fish assemblages: the balance between local and regional processes. Am. Nat., 136(3): 345-375.
- TORO, M.; ROBLES, S.; AVILÉS, J.; NUÑO, C.; VIVAS, S.; BONADA, N.; PRAT, N.; ALBA-TERCEDOR, J.; CASAS, J.; GUERRERO, C.; JAIMEZ-CUÉLLAR, P.; MORENO, J. L.; MOYÁ, G.; RAMÓN, G.; SUÁREZ, M. L.; VIDAL-ABARCA, M. R. (in press.). Calidad de las aguas de los ríos mediterráneos del proyecto Guadalmed. Caracteríticas físico-químicas. *Limnética.*
- TOWNSEND, C. R. & HILDREW, A. G. (1994). Species traits in relation to a habitat templet for river systems. *Freshwat. Biol.*, 31: 265-275.
- TRABAUD, L. (1981). Man and Fire: Impacts on mediterranean vegetation. In: di Castri, FR.; Goodall, D. W. & Specht, R. W. (eds.). *Mediterranean-Type Shrublands*. Elsevier, New York. 643pp.
- VALDOVINOS, C. (2001). Riparian leaf litter processing by benthic macroinvertebrates in a woodland stream of central Chile. *Rev. Chil. Hist. Nat.*, 74(2).
- VIDAL-ABARCA, M. R. (1990). Los ríos de las cuencas áridas y semiráridas: una perspectiva ecológica comparativa y de síntesis. *Sentía gerundensis*, 16(1): 219-228.
- VIDAL-ABARCA, M. R. (2001). Los ríos mediterráneos: Características y funciones, una perspectiva ecológica. Life Guadajoz.
- VIDAL-ABARCA, M. R.; SUÁREZ, M. L. & RAMÍREZ-DÍAZ, L. (1992). Ecology of Spanish semiarid streams. *Limnética*, 8: 151-160.
- VIDAL-ABARCA, M. R.; SUÁREZ, M. L. & RAMÍREZ-DÍAZ, L. (1996). Ramblas/Wadis. In: Morillo, C. & González, J. L. (eds.). Management of Mediterranean wetlands. Ministerio de Medio Ambiente. 17-38pp.
- VINSON, M. R. & HAWKINS, C. P. (1998). Biodiversity of stream insects: Variation at local, basin and regional scales. *Annu. Rev. Entomol.*, 43: 271-293.
- WALKER, F. (1952). In: Mabbut, J. A. (ed.). The Cape Peninsula. Cape Town.

- WALTER, H. (1979). Vegetation of the earth and ecological systems of the geo-biosphere. Second edition. Springer Verlag. New York. 274pp.
- WALTER, H. & LIETH, H. (1960). Klimadiagramm-Weltatlas. Fisher, Jena.
- WARDELL-JOHNSON, G. & HORWITZ, P. (1996). Conserving biodiversity and the recognition of heterogeneity in ancient landscapes: a case study from south-western Australia. *Forest Ecology and Management*, 85: 219-238.
- WAYLEN, P.; COMPAGNUCCI, R. & CAFFERA, R. M. (2000). Interannual and interdecadal variability in stream flow from the Argentine Andes. *Phys. Geogr.*, 21(5): 452-465.
- WELLS, M. J. (1991). Introduced plant of the fynbos biome of South Africa. In: Groves, R. H. & di Castri, F. (eds). *Biogeography of Mediterranean Invasions*. Cambridge University Press. Cambridge. 485pp.
- WELLS, M. J.; ENGELBRECHT. V. M.; BALSINHAS, A. A.; STIRTON, C. H. (1983). Weed flora of South Africa 3: more power shifts in the veld. *Bothalia*, 14: 967-970.
- WIENS, J. A.; STENSETH, N. C.; VAN HORNE, B. & IMS, R. A. (1993). Ecological mechanisms and landscape ecology. *Oikos*, 66: 369-380.
- WILLIAMS, W. D. (1985). Biotic adaptations in temporary lentic waters, with special reference to those on semi-arid regions. *Hydrobiologia*, 125: 85-110.
- WILLIAMS, D. D. (1987). *The Ecology of Temporary Waters*. The Blackburn Press. New Jersey. 205pp.
- WILLIAMS, D. D. & FELTMATE, B. W. (1992). Aquatic Insects. CAB International. Wallingford.
- WILLIAMS, D. D. & HYNES, H. B. N. (1976). The ecology of temporary streams. II. General remarks on temporary streams. *Internationale Revue der gesamter Hydrobiologie*, 62: 53-61.
- WISHART, M. J. (1998). Temporal variations in a temporary sandbed stream. MSc. Thesis. University of Cape Town. South Africa.
- WRIGTH, J. F.; HILEY, P. D.; COOLING, D. A.; CAMERON, A. C.; WIGHAM, M. E. & BERRIE, A. D. (1984). The invertebrate fauna of a small chalk stream in Berkshire, England, and the effect of intermittent flow. Arch. Hydrobiol., 99: 176-199.
- ZWICK, P. (2000). Phylogenetic system and zoogeography of the Plecoptera. Ann. Rev. Entomol., 45: 709-746.

		California	MedBasin	Chile	South- Africa	SW Australia	S Australia
PLATIHELMINT	Dugesiidae	1	1	1	1	1	0
	Planariidae	1	1	0	0	0	0
	Temnocephalidae	0	0	1	0	1	1
GORDIIDAE	Gordiidae	1	1	1	1	1	1
OLIGOCHAETA	Oligochaeta	1	1	1	1	1	1
POLIQUETA	Neiridae	1	0	0	0	0	0
	Syllidae	0	0	0	0	0	1
AQUETA	Erpobdellidae	1	1	0	0	0	1
	Glossiphoniidae	1	1	0	0	0	0
	Hirudinidae	1	1	0	1	0	0
	Mesobdellidae	0	0	1	0	0	0
	Richardsonianidae	0	0	0	0	1	1
MOLLUSCA	Amnicolidae	0	0	1	0	0	0
	Ancylidae	1	1	1	1	1	1
	Bithynellidae	0	1	0	0	0	0
	Bithyniidae	0	1	0	0	0	1
	Chilinidae	0	0	1	0	0	0
	Corbiculiidae	1	0	0	1	1	1
	Ferrissiidae	0	1	0	0	0	0
	Hvdrobiidae	1	1	0	1	1	1
	Hvriidae	0	0	1	0	1	1
	Lymnaeidae	1	1	1	1	1	1
	Neritidae	0	1	0	0	0	0
	Physidae	1	1	1	1	1	1
	Planorbidae	1	1	1	1	1	1
	Pomatopsidae	0	0	0	0	- 1	- 1
	Sphaerijdae	1	1	1	1	- 1	- 1
	Thiaridae	- 1	1	0	1	- 1	- 1
	Unionidae	1	1	0	1	0	0
	Valvatidae	- 1	1	0	0	0	0
	Viviparidae	0	1	0	0	0	1
CRUSTACEA	Aeglidae	0	0	1	0	0	0
	Amphisopodidae	0	0	0	1	1	0
	Anthuridae	0	0	0	1	1	0
	Asellidae	1	1	1	0	0	0
	Astacidae	1	1	0	0	0	0
	Atvidae	1	1	0	1	1	1
	Cambaridae	1	1	0	1	0	0
	Ceinidae	0	0	0	0	1	1
	Cirolanidae	0	0	0	0	0	1
	Fusiridae	0	0	0	0	0	1
	Gammaridae	1	1	1	0	0	0
	Gransidae	1	1	0	1	1	0
	Hvallelidae	1	0	1	1	0	0
	Humenosomotidae	1	0	0	0	1	1
	Igniridae	0	0	0	1	1	1
	Mesoveliidae	1	1	0	1	1	1
	Munnidae	1	1	0	1	1	1
	Opiocidae	0	0	0	1	1	1
	Oniscidae	1	1	0	1	1	1
	Delegment	1	1	U	1	1	1
		0	1	0	1	1	1
	Paracamopidae	U	0	0	1	1	1
	ratamen00ae	U	U	0	1	1	1

Annex 1. Presence and absence of taxa in each mediterranean region.

		California	MedBasin	Chile	South- Africa	SW Australia	S Australia
	Parastacidae	0	0	0	0	1	1
	Perthiidae	0	0	0	0	1	1
	Phreatoicidae	0	0	0	1	1	0
	Potamonautidae	0	0	0	1	0	0
	Protojaniridae	0	0	0	1	0	0
	Samastacidae	0	0	1	0	0	0
	Spelaeogriphidae	0	0	0	1	0	0
	Sphaeromatidae	0	0	0	1	1	0
	Sternophysingidae	0	0	0	1	0	0
	Talitridae	0	0	0	0	0	1
ACARI	Hydracarina	1	1	1	1	1	1
	Oribatidae	1	1	0	0	1	1
EPHEMEROPTERA	Ameletidae	1	0	0	0	0	0
	Amelotopsidae	0	0	1	0	0	0
	Baetidae	1	1	1	1	1	1
	Caenidae	1	1	1	1	1	1
	Coloburiscidae	0	0	1	0	0	0
	Ephemerellidae	1	1	0	0	0	0
	Ephemeridae	1	1	0	0	0	0
	Heptageniidae	1	1	0	1	0	0
	Leptophlebiidae	1	1	1	1	1	1
	Oligoneuriidae	1	1	1	0	0	1
	Onicigastridae	0	0	1	0	0	1
	Polymitarcidae	0	1	0	0	0	0
	Potamanthidae	0	1	0	0	0	0
	Prosopistomatidae	0	1	0	0	0	0
	Siphlonuridae	1	1	1	0	0	1
	Teloganodidae	0	0	0	1	0	0
	Tricorythidae	1	0	0	1	0	0
PLECOPTERA	Austronemouridae	0	0	1	0	0	0
	Austroperlidae	0	0	1	0	0	1
	Capniidae	1	1	0	0	0	0
	Chloroperlidae	1	1	0	0	0	0
	Diamphipnoidae	0	0	1	0	0	0
	Eustheniidae	0	0	1	0	0	1
	Gripopterygidae	0	0	1	0	1	1
	Leuctridae	1	1	0	0	0	0
	Nemouridae	1	1	0	0	0	0
	Notonemouridae	0	0	1	1	0	1
	Peltoperlidae	1	0	0	0	0	0
	Perlidae	1	1	1	0	0	0
	Perlodidae	1	1	0	0	0	0
	Pteronarcyidae	1	0	0	0	0	0
LEPIDOPTERA	Nymphulidae	1	0	0	1	0	0
	Pyralidae	1	1	1	0	1	1
MECOPTERA	Nannochoristidae	0	0	1	0	0	0
	Eomeropidae	0	0	1	0	0	0
MEGALOPTERA	Corydalidae	1	0	1	1	0	0
	Sialidae	1	1	1	0	0	0
NEUROPTERA	Chauliodidae	1	0	0	0	1	0
	Osmylidae	0	1	1	0	0	0
	Sisyridae	1	1	0	1	0	1
COLEOPTERA	Brentidae	0	0	0	0	1	1
	Carabidae	0	0	0	0	1	1
	Colymbetidae	1	1	0	0	0	0

Chapter 3

		California	MedBasin	Chile	South- Africa	SW Australia	S Australia
	Crysomelidae	1	0	0	0	1	1
	Dryopidae	1	1	1	1	0	0
	Dytiscidae	1	1	1	1	1	1
	Elmidae	1	1	1	1	1	1
	Georrisidae	0	1	0	0	0	0
	Gyrinidae	1	1	1	1	1	1
	Haliplidae	1	1	1	1	1	1
	Helodidae	1	1	0	1	1	0
	Helophoridae	1	1	0	0	0	0
	Heteroceridae	0	0	0	0	1	1
	Hydraenidae	1	1	1	1	1	0
	Hydrochidae	0	1	0	0	0	1
	Hydrophilidae	1	1	1	1	1	1
	Hygrobiidae	0	1	0	0	0	1
	Limnebiidae	0	1	0	0	0	0
	Limnichidae	1	0	1	1	0	1
	Noteridae	0	1	0	1	0	0
	Psephenidae	1	0	1	0	0	1
	Scirtidae	0	0	0	0	1	1
	Staphylinidae	0	0	0	0	1	1
HETEROPTERA	Aphelocheiridae	0	1	0	0	0	0
	Belastomatidae	1	0	1	1	0	1
	Corixidae	1	1	1	1	1	1
	Gerridae	1	1	1	1	1	1
	Hebridae	1	1	0	1	1	1
	Hydrometridae	1	1	0	1	1	1
	Naucoridae	1	1	0	1	1	1
	Nepidae	1	1	0	1	1	1
	Notonectidae	1	1	1	1	1	1
	Ochteridae	0	0	0	0	0	1
	Pleidae	1	1	0	1	1	1
	Saldidae	1	0	0	0	0	1
	Veliidae	1	1	0	1	1	1
ODONATA	Aeshnidae	1	1	1	1	1	1
	Calopterygidae	1	1	1	0	0	0
	Chlorophydae	0	0	0	1	0	0
	Chorolestidae	0	0	0	1	0	1
	Coenagrionidae	1	1	1	1	1	1
	Cordulegasteridae	1	1	0	0	0	0
	Corduliidae	0	1	1	1	1	1
	Gomphidae	1	1	1	1	1	1
	Hemicorduliidae	0	0	0	0	0	1
	Lestidae	1	1	1	1	1	1
	Libellulidae	1	1	1	1	1	1
	Megapodagrionidae	0	0	0	0	1	1
	Petaluridae	0	0	1	0	1	0
	Platycnemididae	0	1	0	1	0	0
	Protoneuridae	0	0	0	1	0	1
	Synthemidae	0	0	0	0	1	1
	Telephlebiidae	0	0	0	0	0	1
TRICHOPTERA	Atriplectididae	0	0	0	0	1	1
	Beraeidae	0	1	0	0	0	-
	Brachycentridae	1	1	0	0	0	0
	Calamoceratidae	1	1	1	0 0	0	1
	Conoesucidae	0	0	0	0 0	0 0	1

		California	MedBasin	Chile	South- Africa	SW Australia	S Australia
	Dipseudopsidae	0	0	0	1	0	0
	Ecnomidae	0	1	1	1	1	1
	Glossosomatidae	1	1	1	1	0	1
	Goeridae	1	1	0	1	0	0
	Helicophidae	0	0	1	0	0	1
	Helicopsychidae	1	0	1	0	0	1
	Hydrobiosidae	0	0	1	0	1	1
	Hydropsychidae	1	1	1	1	1	1
	Hydroptilidae	1	1	1	1	1	1
	Hydrosalpingidae	0	0	0	1	0	0
	Lepidostomatidae	1	1	0	1	0	0
	Leptoceridae	1	1	1	1	1	1
	Limnephilidae	1	1	1	0	0	0
	Molanidae	0	0	1	0	0	0
	Odontoceridae	1	1	1	0	0	1
	Petrothrincidae	0	0	0	1	0	0
	Philopotamidae	1	1	0	1	1	1
	Philorheithridae	0	0	0	0	1	1
	Pisuliidae	0	0	0	1	0	0
	Plectrotarsidae	0	0	0	0	1	0
	Polycentropodidae	1	1	1	1	1	1
	Psychomyiidae	1	1	1	0	0	0
	Rhyacophilidae	1	1	0	0	0	0
	Sericostomatidae	1	1	1	1	0	0
	Tasimiidae	0	0	0	0	0	1
	Uenoidae	1	1	0	0	0	0
	Barbarochtharidae	0	0	0	1	0	0
DIPTERA	Anthomyidae	1	1	0	1	1	1
	Athericidae	1	1	1	1	1	0
	Blephariceridae	1	1	1	1	0	0
	Ceratopogonidae	1	1	1	1	1	1
	Chironomidae	1	1	1	1	1	1
	Culicidae	1	1	1	1	1	1
	Dixidae	1	1	1	1	1	1
	Dolichopodidae	1	1	0	1	1	1
	Empididae	1	1	1	1	1	1
	Ephydridae	1	1	1	1	1	1
	Limoniidae	1	1	1	0	1	0
	Psychodidae	1	1	1	1	1	1
	Ptycopteridae	0	1	1	1	0	0
	Scatopsidae	0	0	0	0	0	1
	Sciomyzidae	1	1	0	1	1	-
	Simuliidae	- 1	- 1	1	-	- 1	- 1
	Stratiomvidae	1	1	1	1	1	1
	Symbidae	1	1	1	1	1	1
	Tabanidae	1	1	1	1	1	1
	Tanyderidae	0	0	1	1	0	0
	Thaumaleidae	1	1	0	0	1	1
	Tipulidae	1	1	1	1	1	1

Plate 1. Characteristics of the mediterranean region in California.





Climatic features

Location and latitude in almost all California, south of Oregon and north of Mexico (1), 28-44°N (2). Extension of med-region around 250.000 km2 (2) Rank of precipitation from 150mm/y to 1800mm/y (1) with snow in the high mountains of Sierra Nevada. Months with highest rain between November and March (3). Range of altitude between 0 to >4000m. Population of 27.000.000 inhabitans (4)

Basin characteristics

Geology composed by metamorphosed, granites and sedimentary deposits (2).

Landscape and topography composed by Coast, Transverse and Peninsula ranges, Great or Central Valley and a more alpine area in western side of Sierra Nevada, Klamath and Cascade mountains (5&6). Coastal plains scarce.

Vegetation with coniferous trees and meadows in alpine areas, deciduous trees in wet areas, redwood formation in northern California (*Sequoia sempervirens*) and sclerofillous and evergreen trees and shrubs (called "chaparral"), savannas or dry steppe in drier areas (2&4).

Abiotic river properties

Annual discharge peak in winter but also in spring in nival influenced rivers (7). Discharge can be affected by El Niño event (8). **Physical and Chemical properties:** oligotrophic rivers with high pH (9). **River and stream typologies:** short and steepe streams and rivers from coastal ranges longer, steepeness and wider from inland mountains;

karstic rivers in southern California (8). **Riverbed substrate** composed by boulders, cobbles, gravels in headwaters and sands and silt downstream. Some bedrock rivers in

Biotic river properties

Periphyton community highly diverse, abundant and productive.

Macroinvertebrate community highly diverse but lacks of families present in other areas of North America. Some non-native macroinvertebrate's families as Corbiculiidae and Grapsidae.

southern California (10&11).

Fish community is highly endemic with 115 taxa (12) closer to European fish fauna than to that of eastern North America (12). Several non-native species are present as *Gambusia affinis,G. holbrooki, Micropterus salmoides, Alosa sapidissima, Lepomis macrichirus,...*(12&14).

Alloctonous material reach riverbed in autumn, although in undisturbed areas where riparian vegetation unable to survive, sclerofillous leaves arrive to the river along the year (9). Higher decomposition rates in deciduous leaves (15).

Riparian Vegetation mainly with deciduous trees and shrubs (16) as Salix sp., Alnus sp., Corylus cornuta, Populus sp. or Fraxinus sp. Highly altered by non-native species introductions as Arundo donax, Nicotiana sp., or Tamarix sp.

References

1-Orme & Orme, 1998 2-Grove & Rackham, 2001 3-Kahrl et al., 1978 4-Dallman, 1998 5-Thrower & Bradbury, 1973 6-Hornbeck et al., 1983 7-Erman et al., 1988 8-Mouth, 1995 9-del Rosario et al., 2002 10-Cooper et al., 1986 11-Dudley et al., 1986 12-Moyle, 1995 13-Moyle & Herbold, 1987 14-Moyle & Ligth, 1996 15-Hart & Howmiller, 1975 16-Holstein, 1984



Plate 2. Characteristics of the mediterranean region in Mediterranean Basin.



Annual discharge peak. Two annual peaks, in spring and autumn (6 & 22).

Physical and Chemical properties: in siliceous rivers and streams, neutral pH and oligotrophic waters (7), whereas in more calcareous areas higher pH and less oligotrophic waters (8). High natural salinity (by sulphates and chloride) in some semiarid areas (9,10&11).

River and stream typologies: short and steepe streams from coastal ranges longer and wider rivers from inland mountains; karstic rivers frequent (12&13); highly ephemeral rivers called "oueds", "ramblas", "wadis", "torrents" or "barrancos" (14). Some deltas plains are present as in Po, Rhône, Ebre rivers.

Riverbed substrate composed by boulders, cobbles, gravels and bedrock in headwaters and sands and silt downstream.

Biotic river properties

Periphyton community highly diverse, abundant and productive.

Macroinvertebrate community highly diverse (15) and abundant in European area compare with African one (6&16), although a high endemism is present in both regions (6&17). Northern rivers and streams dominated by EPT communities whereas more Ephemeroptera and Coleoptera are present increasing aridity through the south (18). Some non-native macroinvertebrate's species as *Procambarus clarkii* and *Potamopyrgus jenksii*.

Fish community higly diverse, endemic (80% of the exclusive freshwater species) compared with the rest of Europe (19). Several non-native species are present as Oncorhyncus mykiis, Micropterus salmoides, Esox lucius, Carassius auratus, Cyprinus carpio, Gambusia holbrooki, Alburnus alburnus, Hucho hucho, Cichlasoma facetum and others (19)

Alloctonous material reach riverbed in autumn with a second peak between January and April (20), although in undisturbed areas where riparian vegetation unable to survive, sclerofillous leaves arrive to the river along the year. Higher decomposition rates in deciduous leaves compared with sclerofillous ones (21).

Riparian Vegetation mainly with deciduous trees and shrubs as *Salix* sp., *Alnus glutinosa, Ulmus sp., Corylus avellana, Populus* sp., *Fraxinus* sp or *Tamarix* sp. Highly altered by non-native species introductions as *Arundo donax, Nicotiana sp., Robinia pseudoacacia, Platanus hispanica or Ailanhus altissima.*

References

1-Dell et al., 1986	12-Bilandzija et al., 1998
2-Grove & Rackham, 2001	13-Rieradevall et al., in press
3-Le Houérou, 1990	14-Vidal-Abarca, 1990
4-Dallman, 1998	15-Prat et al., 2000
5-di Castri, 1981	16-Lounaci et al., 2000
6-Guidicelli et al., 1985	17-Gallardo et al., 1992
7-Rieradevall et al., 1997	18-Alba-Tercedor et al., 1992
8-Graça et al., 1989	19-Doadrio, 2001
9-Gallardo-Mayenco, 1994	20-Maamri et al., 1994
10-Gallardo-Mayenco et al., 1998	21-Schwarz&Schwoerbel, 1997
11-Moreno et al., 2001	22-Dolédec, 1989







Plate 3. Characteristics of the mediterranean region in Chile.



Periphyton community present in low abundances (10)

Macroinvertebrate community poorly known compare with other med-regions. Several taxonomical works have been done in Plecoptera (11&12), Trichoptera (13) and Chironomidae (14) and ecological studies are increasing (15). The plecopterans Austronemouridae and Diamphipnoidae are endemic.

Fish community highly diversified with 6 endemic families from Central Chile (*Trichomycterus areolatus, Percilia gillissi, Cauque mauleanum, Basilichthys australis, Cheirodon australe* and *Brachygalaxias bullocki*), 1 present in Central Chile and South of Argentina (*Percichthys trucha*) and another shered with Argentina, New Zealand, Australia and Tasmania (*Galaxias maculatus*). Several non-native species are present as Salmo trutta, Oncorhyncus mykiis, Ciprinus carpio and Gambusia affinis holbrooki) (8).

Alloctonous material reach riverbed in summer except for woodland forests.

Riparian Vegetation with evergreen and deciduous trees as *Drimys winteri* and *Salix chilensis* and several shrubs as *Myrceugenella chequen* and *Aristotelia chilensis*.Highly altered by non-native introductions of *Pinus radiata* (16)

References

1-Grove & Rackman, 2001 2-Emberger, 1930, 1955, 1971 3-Thrower & Bradbury, 1973 4-Dallman, 1998 5-Hajek, 1991 6-Waylen et al., 2000 7-Hedin & Campos, 1985 8-Campos, 1985 9-Figueroa et al., 2000 10-Soto & Campos, 1997 11-Illies, 1964, 1965 12-Benedetto, 1974 13-Filnt, 1974 14-Brundi, 1966 15-Figueroa & Araya, 2002 16-Valdovinos, 2001





Plate 4. Characteristics of the mediterranean region in South Africa.





Climatic features

Location and latitude in South-western Cape, 32-35°S. Extension of med-region around 40.000 km² (1). Rank of precipitation from 230mm/y to >2500mm/y (2) with snow unfrequent. Months with highest rain between April and September Range of altitude between 0 to 2300m (3). Population of 3.400.000 inhabitans (4).

Basin characteristics

Geology with sandstones and shales (3) with acidic and low in nutrients soils (5).

Landscape and topography composed by plains and coastal (Table Mountain) and inland ranges (Franschoek and Drakenstein Mountains) separated from the Cretaceous (6). Vegetation dominated by fynbos biome (7&8) with evergreen and sclerofilous shrubs highly endemic (9) and divided in mountain and coastal fynbos, coastal renosterveld and strandveld (7).

Abiotic river properties

Annual discharge peak in winter months.

Physical and Chemical properties: acid and oligotrophic headwaters (10) with brown colour (11), but clearer, higher in nutrients and pH downstream (10).

River and stream typologies: $\ensuremath{\mathbf{c}}$

Riverbed substrate composed by boulders, bedrock, cobbles and gravels in headwaters and stones, pebbles and coarse sand downstream (13).

Biotic river properties

Periphyton community scarce (14) with some macrophytes abundant.

Macroinvertebrate community highly endemic with Gondwanic origin (15&16) but lower in diversity compare with other South African rivers. Presence of taxa adapted to a low fish diversity and abundance (16). Fish community low in diversity and abundance with some non-native species as Salmo trutta, Tilapia sp., Lepomis macrochinus, Ciprinus carpio, Micropterus dolomieu and Micropterus salmoides (17&18).

Alloctonous material reach riverbed in summer. Low decomposition rate (10&14).

Riparian Vegetation shrubby with Prionium serratun, *Metrosideros angustifolia* and *Brabejum stellatifolium*. Non-native species as *Acacia melonoxylon*, *Quercus robur*, *Eucaliptus* sp. and *Hakea* sp. are present (19&20).

References

1-Grove & Rackman, 2001 2-South African Weather Buerau, 1996 3-Thrower & Bradbury, 1973 4-Dallman, 1998 5-Specht & Moll, 1983 6-Walker, 1952 7-Kruger, 1979ab 8-Fuggle & Ashton, 1979 9-Cowling, 1992 10-King et al., 1979 11-Midgley & Schafer, 1992 12-Harrison & Barnard, 1972 13-Brown & Dallas, 1995 14-King, 1982 15-Harrison & Agnew, 1962 16-de Moor, 1992ab 17-de Moor, 1992ab 17-de Moor & Brutton, 1988 18-de Moor, 1992b 19-Hall, 1979 20-Wells et al., 1983; Wells, 1991



Plate 5. Characteristics of the mediterranean region in Australia.





Abiotic river properties

Climatic features

Location and latitude in South-Western Australia (WA) and South Australia (SA and NSW) (1,2&3) disconnected by 1000km, 28-37°S (4).
Extension of med-region around 350.000 km2 (4)
Rank of precipitation from 300mm/y to 1400mm/y (5&6) without snow.
Months with highest rain between April and November.
Range of altitude between 0 to 582m in SW-Australia (7) and to 1300m in S-Australia (8).

Population of 3.000.000 inhabitans (9)

Basin characteristics

Geology composed by sandstones and granites, quarzites and sedimentary deposits (8).

Landscape and topography with a soft relief in SW-Australia with coastal plains an inland plateaus (Darling Plateau) and more abrupt in S-Australia with Flinders Ranges (7). **Vegetation** with by sclerofillous and evergreen shrubs and trees in the coast (called "health" and "mallee"), and Jarrah and Marri forest in the inland plateau in SW-Australia, with *Eucaliptus diversicolor* and *E. marginata* as dominant species (9). Highly endemic (10).

Annual discharge peak in winter months between June and November (11).

Physical and Chemical properties: oligotrophic rivers with slightly low pH (11).

River and stream typologies: slightly steep rivers in S-Australia coming from Flinger Ranges but steepeness streams in SW-Australia flowing from the Darling Plateau to the coast. A big river system "Murray-Darling River" present in S-Australia.

Riverbed substrate composed by gravel and sand with some large rocks (2)

Biotic river properties

Periphyton community scarce and slightly productive (12)

Macroinvertebrate community depauperate in SW-Australia respect S-Australia (2,12&13). Some Gondwanic relicts as some species of the crayfishes *Cherax* sp. and *Engaewa* sp. (14).

Fish community highly diverse in the Murray-Darling river system. In South Western Australia the community is dominated by *Galaxias occidentalis*, *Bostockia porosa* and the endemic *Edelia vittata*. Several non-native species are present as *Oncorhynkus mykiss*, *Gambusia holbrooki* and *Perca fluviatilis*

Alloctonous material reach riverbed in summer (15&16). Few shredders to process CPOM (17).

Riparian Vegetation mainly sclerofillous trees and shrubs as *Dodonea* sp., *Banksia* sp., *Hakea* sp., *Trymalium floribundum, Melaeuca* sp., *Callitris sp. Angianthus* sp., *Eragrostis* sp., and several *Eucaliptus* sp. Altered by nonnative introductions of *Pinus* rsp *Mimosa pigra* are common (18).

References

- 1-Gardner, 1942
- 2-Bunn, 1988 3-Gentilli, 1989

4-Grove & Rackham, 2001

- 5-Beard, 1983
- 6-Chuchward & Dimmock, 1989 7-Conacher & Conacher, 1998
- 7-Conacher & Conacher, 1998 8-Thrower & Badbury, 1973
- 9-Dallman, 1998
- 10-Cowling, 1992
- 11-Bunn et al., 1986 12-Bunn & Davies, 1990
- 13-Smith et al., 1999
- 14-Wardell-Johnson & Horwitz, 1996 15-Attiwill et al., 1978
- 15-Attiwill et al., 1978 16-Lake et al., 1986
- 17-Boulton & Brock, 1999
- 18-Groves, 1991





